Running title: MEASURING THE CONDITION OF SALINE WETLANDS

Measuring the Condition of Saline Wetlands Threatened by Agricultural

Intensification^{*1}

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ABSTRACT

The saline wetlands, or “saladas”, of Monegros Desert, NE Spain, contain biodiversity that depends on the establishment of a prognostic monitoring system that can warn of approaching irreversible damage. In the context of a lack of existing biodiversity inventories for the saladas, we have determined their state of conservation and vulnerability using seven physical indicators: escarpment continuity, cropping, stone dumping, size, water occurrence, distance to roads, and proximity to irrigated areas. These features are combined into three meaningful indexes, i.e., conservation, current vulnerability, and future vulnerability, thus creating an assessment of the preservation or degradation of saladas in the context of encroaching agricultural irrigation projects. The proposed indexes produced consistent results and showed that a great number of the Monegros wetlands are threatened, regardless of their size or frequency of water occurrence. Only 20% of the saladas studied are classified as being in a *good* or *very good* conservation, whereas 50% are in *bad* or *very bad* shape. A *high*

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current vulnerability was found for 60% of the saladas. For saladas located in land to be irrigated, we predict 73% will have a *high* or *very high* future vulnerability. Currently, 58% of the saladas are in *bad* or *very bad* condition and fully 65% of the saladas, variable in size, present a *bad* or *very bad* prognosis. Our approach provides a monitoring strategy for the conservation of saline wetlands threatened by agricultural intensification, especially irrigation.

Key Words: agricultural landscape, desert, irrigation, Monegros, playa.

INTRODUCTION

The degradation or loss of native wetlands is a matter of major environmental concern worldwide because of the significant biodiversity and the hydrological utility of these lands. Beyond the many physical threats to wetlands already identified, the scientific literature also describes how lack of information, historical legacies of land use and tenure, and political obstacles constrain conservation efforts (Naranjo, 1995; Davis and Froed, 1999; Finlayson and Rea, 1999; Smith, 2003; Uluocha and Okeke, 2004; Junk and Piedade, 2004).

Ecologists have developed indicators which enable them to describe environments with a complexity of components, as in Dale and Beyeler (2001), Tiner (2004), or Müller *et al.* (2006). Using natural and anthropogenic indicators, Klijn and Udo de Haes (1994) introduced abiotic ecosystem classifiers in the Netherlands. In non-ecological situations, abiotic indicators based on soil and geological processes and on geomorphologic features have been used to study the sustainability and the productivity of land (Cendrero *et al.*, 2003) and to assess environmental impact (Rivas *et al.*, 1997). Several sets of environmental indicators have been proposed by international agencies (OECD, 2003; EEA, 2004; UNEP, 2004). In Europe, recent studies have sought to establish indicators

that can be processed by a geographic information system with the aim of monitoring Natura 2000 Habitats (Boteva *et al.*, 2004; Langanke *et al.*, 2005; Lang and Langanke, 2005).

Wetlands, as sensitive environments, require specific diagnostic indicators. Numerous visual features have been considered and ecological indicators have been developed (i) to identify morphological features and to quantify geomorphic attributes (Jurmu and Andrlé, 1997; Adamowicz and Roman, 2005); (ii) to detect alterations of hydrological functions (Hruby, 2001; Lichvar *et al.*, 2004; Cole, 2006); (iii) to assess the ecosystem integrity based on biological metrics (Cohen *et al.*, 2005) or on remote sensing surveys (Phillips *et al.*, 2005); or, (iv) to define the components of conceptual models for restoration (Mayer *et al.*, 2004). An operational synthesis of the above approaches, the Wetland Rapid Assessment Procedure (WRAP) was adapted by Breaux *et al.* (2005). Of the wetland assessment methods reviewed by Fennessy *et al.* (2004), only the WRAP allowed rapid measurement that is conducted on site and is verifiable. Frequently, wetlands are rated using outside evaluators (Hruby, 2001; Breaux *et al.*, 2005). These methods were conceived for humid countries, i.e., where there is permanent, fresh water; these metrics are not suitable for arid lands. Moreover, these assessment tools depend strongly on databases with site information, which is not the case for our study area or for many other arid lands.

Due to their scientific and ecological importance, half of the saladas territory has been classified by the local government as habitats to be included in the European Natura2000 network. The remainder of the saladas territory is being transformed for irrigation purposes (a project is set to irrigate over 60 000 ha in southern Monegros). The irrigation works, which will supply water from Pyrenean rivers, will be managed by the local government. For 10 years the project has been on hold, as European Union

funds have been blocked by the lobbying efforts of ecology activists. The project involves serious landscape transformations due to farm consolidation and the works associated with agricultural intensification.

The aim of this paper is to set up useful indicators and indexes to assess the conservation, vulnerability and prognosis of the Monegros saline wetlands. For this purpose we use explicit criteria that are adaptable to changes in the area, and to other arid lands. This integrated procedure can easily be applied to other wetlands in arid environments.

STUDY AREA

Saline wetlands are found in the Monegros desert, one of the most arid regions of Europe (Herrero and Snyder, 1997). Rainfall is very irregular, with a minimum annual average of 175 mm, a maximum of 535 mm, and an average of 354 mm recorded by the Petris Weather Station for the period 1973–2006. The mean annual ET_0 is 1255 mm, based on the estimate of Faci and Martínez-Cob (1991) for the nearby Bujaraloz Weather Station.

These saline depressions are scattered across a structural geological platform which is almost 200 m above the Ebro River water level, forming an endorheic complex closely connected to the hypersaline aquifers. Their name “salada” refers to the presence of salts in soils, surface water and groundwater, and also reminds one of their past use as salterns. They could be classified as depressional, isolated wetlands according to the Leibowitz and Nadeau (2003) definition. The saladas are fed primarily by groundwater. Their seasonal dry-wet cycles are relevant for their biodiversity and the endemic organisms that have adapted to fluctuations of brine level and chemistry. Halophilous vegetation, scarcely represented in the European inlands, appears in concentric fringes according to salinity tolerance, water availability and soil texture; the center of the

wetland remains bare due to salinity and flooding. Elaboration of detailed maps of priority habitats and endemic vegetation in each salada are in progress (Domínguez et al., 2006).

More than a hundred saladas were inventoried in the region two decades ago. Of these, only 53 (Fig. 1) were recognized in our updated *Inventory 2003* (Castañeda and Herrero, *in revision*). Ranging from less than 1 ha to more than 200 ha, the saladas lie hidden in the agricultural landscape, occupying the lowest positions of the gentle undulated topography. The saladas bottom elevation ranges between 310 and 370 m a.s.l. and they are shaped by escarpments of up to twenty meters tall which are aligned in the prevalent tectonic and wind direction, NW-SE.

The total surface of the saladas, almost 1000 ha, represents less than 3% of the whole agricultural platform. The platform is under dry farming, often surpassing the natural limits imposed by intermittent water occurrence in the saladas or by soil salinity, with winter cereal crops in soils having an electrical conductivity of 4 dS m^{-1} in the top soil layer (0-60 cm). The subsidies for rainfed cereal crops encouraged the agricultural invasion of marginal lands. As a result, recent decades have seen the disappearance and deterioration of some wetlands. Land consolidation, new roads, new power lines, drainage pipes and trenches, irrigation ponds, machinery traffic, and other construction are changing the appearance of the area, reducing the native vegetation and smoothing the salada escarpments, as observed during the field campaigns. Moreover, the close proximity of currently irrigated lands in the northern part of the platform (Fig. 1) and their projected extension towards the south will disturb their hydrological cycle. Several saladas are already artificially flooded with fresh and polluted water from the return flows of the newly irrigated lands.

[Fig. 1]

MATERIAL AND METHODS

We use the saladas *Inventory 2003*, (Castañeda, 2002, pages 33-37; Castañeda and Herrero, *in revision*) containing their conservation status, as established in the field using external expert judgment, and systematic descriptions of these saline wetlands obtained by ground surveys, remote sensing and orthophotographs (Table I). For the purposes of this article, we have added to the above database the water occurrence in each salada, and its proximity both to any new roads and to the border of the forthcoming irrigated area, which is delineated following arbitrary straight-line limits.

Following the ecosystem approach that is frequently used in wetland study, we integrated the selected perceptible components of the Monegros saladas at different scales, from the field survey to the satellite images analysis; then a spatial analysis was performed using a geographical information system (GIS) to provide a perspective difficult to obtain from field surveys alone.

Hydrology, geomorphology, and biology, were included in our examination of the Monegros saladas by means of the following interrelated perceptible components: water occurrence, the appearance of the depressions, and the preservation of native vegetation. Though it is important to protect and conserve extremophile microbes and endemisms, these and other intrinsic values can not be used for ranking because their study has just started (Domínguez et al., 2006). We merged crop and natural vegetation borders with abiotic elements to maximize measurement efficiency, validity and reliability. Usually the abiotic elements control the biotic ones; however, the geomorphic and hydrologic control of Monegros saladas has not been recognized in conservation policy. Moreover, the paucity of hydrology data and the lack of soil and vegetation maps have imposed the approach used here for studying the status of the saladas.

In order to measure the status of the saladas in terms of conservation and vulnerability we have developed several indexes. We treat them jointly in a GIS database in order to assess the current condition of these wetlands and for monitoring programs.

TABLE I

Main features registered in the *Inventory 2003*. (+): saladas detected in Landsat images.

Data obtained from: field observations⁽¹⁾, remote sensing⁽²⁾, and GIS analysis⁽³⁾.

Salada name	Size ^{(2) (3)} (ha)	Conservation Status ⁽¹⁾	Escarpment ⁽¹⁾	Dumping ⁽¹⁾	Cropping ⁽¹⁾	Water occurrence ⁽²⁾ (%)	Roads proximity ⁽³⁾ (m)
Agustín (+)	68.1	heavily disturbed	flat bottomed valley	center	partially	27	384
Albacar (+)	10.2	pristine	unrecognizable	without	totally	4	284
Alforjeta I (+)	1.4	slightly disturbed	unrecognizable	center	partially	4	199
Alforjeta II	2.3	slightly disturbed	smooth	center	partially		199
Amarga Alta (+)	22.3	advanced destruction	semi-confining	center and escarpment	partially	27	395
Amarga Baja (+)	7.1	unrecognizable	semi-confining	center	without	31	292
Balsa (+)	2.9	heavily disturbed	smooth	escarpment	totally	4	189
Benamud (+)	19.9	heavily disturbed	smooth	without	partially	35	459
Berchel (+)	2.6	pristine	unrecognizable	without	totally	0	244
Bernab (+)	4.3	heavily disturbed	smooth	center and escarpment	partially	4	310
Berzas	8.8	disturbed	smooth	escarpment	totally		233
Camarón (+)	44.7	unrecognizable	confining	center and escarpment	without	54	362
Catio I	2.3	slightly disturbed	smooth	escarpment	totally		164
Catio II (+)	3.7	slightly disturbed	unrecognizable	escarpment	partially		320
Cerrajero (+)	10.1	pristine	unrecognizable	without	totally	0	703
Chamarqueta (*)	0.5	disturbed	smooth	escarpment	totally		215
clota-19 (+)	1.7	pristine	unrecognizable	without	totally	0	388
clota-37 (+)	4.9	disturbed	smooth	without	totally	4	302
clota-50 (+)	2.5	slightly disturbed	unrecognizable	center and escarpment	partially	0	599
clota-51 (+)	4.1	heavily disturbed	unrecognizable	center	partially	0	123
clota-52	2.8	slightly disturbed	smooth	center	partially		351
clota-94 (+)	9.3	slightly disturbed	smooth	center	partially	0	431
Corral Viejo (+)	4.9	heavily disturbed	smooth	center and escarpment	without	4	491
Correo	4.6	heavily disturbed	smooth	center and escarpment	partially		294
Don Roque (+)	10.0	slightly disturbed	flat bottomed valley	center	partially	12	533
Escobedo	6.3	disturbed	smooth	without	totally		194
Farnaca (+)	12.2	heavily disturbed	flat bottomed valley	center	partially	19	210
Gramenosa (+)	10.7	unrecognizable	confining	center and escarpment	without	4	462
Gros	3.9	heavily disturbed	flat bottomed valley	center and escarpment	partially		241
Guallar (+)	15.0	unrecognizable	semi-confining	center and escarpment	partially	58	278
Herrero	3.2	heavily disturbed	smooth	escarpment	partially		296
Hoya del Pez	2.5	pristine	unrecognizable	center	without		331
La Playa (+)	234.5	unrecognizable	confining	center and escarpment	without	50	601
Larga	10.4	heavily disturbed	flat bottomed valley	center	partially		275
Lisonfer	5.5	heavily disturbed	smooth	center and escarpment	partially		318
Muerte (+)	17.9	unrecognizable	semi-confining	center and escarpment	without	58	303
Nieves (+)	7.0	pristine	smooth	center and escarpment	totally	0	251
Pez (+)	8.3	unrecognizable	semi-confining	without	without	23	502
Piñol (+)	15.7	unrecognizable	confining	center	without	35	459
Pitamar (+)	4.2	pristine	unrecognizable	center and escarpment	partially		296
Pito	67.2	unrecognizable	confining	escarpment	without	35	565
Pueyo (+)	24.3	unrecognizable	confining	escarpment	without	38	421
Rebollón (+)	16.1	advanced destruction	semi-confining	center and escarpment	partially	27	651
Rollico (+)	40.3	advanced destruction	semi-confining	center and escarpment	totally	38	810
Salineta (+)	22.7	unrecognizable	confining	escarpment	without	100	430
San Miguel (+)	10.4	pristine	unrecognizable	center	partially	0	506
Valdecarreta (+)	6.9	heavily disturbed	flat bottomed valley	center and escarpment	partially		244
Valdefrancín	25.7	disturbed	flat bottomed valley	center and escarpment	totally		354
Valdespartosa (+)	14.8	heavily disturbed	flat bottomed valley	center	partially	8	442
Vinagrero I (+)	5.0	slightly disturbed	flat bottomed valley	center and escarpment	partially	4	434
Vinagrero II (+)	5.0	slightly disturbed	flat bottomed valley	center and escarpment	partially	4	249
Yesera I (+)	3.1	slightly disturbed	smooth	center	partially	0	252
Zaborros (+)	9.0	slightly disturbed	smooth	without	totally	4	147

Indicators, criteria and rating

Observations of escarpment, stone dumping, cropping, salada size and water occurrence were used because they are the most outstanding features of these wetlands and because these features are easily observed and interpreted. The distance to roads and future irrigated developments were included as the main conveyors of human pressure. **Escarpment** indicates the confinement of the salada: the higher and more continuous is the escarpment, the better the salada preservation. Usually the escarpment slope is not homogeneous; the northern and southern borders are higher and/or more continuous while salada borders are typically open to the east and the west (Gutiérrez et al., 2002). About 44% of the saladas have a smooth or unrecognizable escarpment. **Cropping** indicates degradation and implies the destruction of native vegetation fringes, the alteration of soil, modified run-off, increased erosion and sedimentation in playas (Haukos and Smith, 1994). Moreover, native vegetation lost to cropping is also lost as a monitoring tool (Finlayson and Rea, 1999). **Stone dumping** is also an indicator of degradation, affecting 77% of the saladas; stones removed from surrounding fields are dumped over the escarpment, or spread in piles over the vegetated bottom or the mud flat. More than 80% of the saladas are affected by stone dumping.

Salada size was obtained from satellite images and orthophotomaps. Size ranges from 0.5 ha to 235 ha, with a median of 7.1 ha. From a developmental point of view, the enlargement of the saladas is related to their maturity; therefore, size is noticeably related to the escarpment and the presence of water, and indirectly related to the risk of disappearance. In general, the largest saladas are the best preserved because they are flooded for some weeks every year; while other non-annually flooded saladas are covered with halophytes and are at risk from plowing to earn subsidies from the European Union. **Water occurrence**, an outstanding feature in this arid zone,

represents aquifer discharge. This feature was the most frequently used metric in early studies of Monegros (Table II), but the lack of standard procedures hampers use of this indicator. The most suitable data source was remote sensing as it provides repetitive synchronic and synoptic observations, and spans the largest period of time. Remote sensing gives surface area, but as water extent cannot be measured in the field because the muddy bottoms are impassable, so our remotely sensed data were converted to presence/absence to allow continuity with field observations. This indicator is expressed as the percentage of satellite images with water in the 39 saladas (Table I) detected with Landsat images (Castañeda *et al.*, 2005).

The **distance to roads** has been computed in order to establish different levels of human influence. Distances range from 810 m to 123 m, with a median of 318 m. Agricultural fields, previously formed as an irregular mosaic adapted to the round-shaped depressions and sinkholes, have become square-shaped due to land consolidation. The associated new road network has fragmented many saladas and has increased their exposure to machinery traffic; previously untrodden soils have been altered by compaction and are slow to recover, typical of arid regions (Belnap and Warren, 2002; Kade and Warren, 2002; Webb, 2002).

The **proximity to the irrigated area** was considered since it increases the risk of degradation, due to the earth movements and the flows from the irrigation. Also, some saladas are planned to host pumping stations and other hydraulic infrastructures.

TABLE II

Number of wetlands listed by different authors, with the type of water observation, period and frequency of recording.

Source	Number of surveyed wetlands	Water observation		
		Type	Period	Frequency
Comín and Sanz (1989)	85	presence / absence	1988	single record
Berga (1993)	15	presence / absence	1989-1990	monthly
Castañeda (2002)	8	water depth	1993-1997	weekly
Castañeda and Herrero (2005)	39	water surface	1985-2000	semiannual

The type and the degree of changes to be expected in these wetlands are unknown; however, previous studies (García-Vera, 1996; Samper and García-Vera, 1998) predict the flooding and salinization of the soils unless drainage is installed. The technology of irrigation water distributed by pressurized pipes has resulted in the artificial segregation of the platform in non irrigated and irrigated land (Fig. 1) affecting many saladas according to the Irrigation Plan. We considered a buffer of 500 m as the area of influence around the future irrigated area. Irrigation effects are not ignored but cannot be included as an indicator, due to the lack of hydrogeological models for each salada.

The above observations are considered as the key indicators of preservation or degradation (Table III) together with the specific process defined by them. The indicators were categorized, and then rated using specific scales according to their significance. Since the indicators came from three different sources, i.e., field, remote sensing and cartography, a custom scheme was required to establish the categories. For these to be significant and helpful in determining the relative status of the saladas, the scales were set taking into account their correspondence with field evidence, and the nature of each indicator. To produce comparable index values, and to avoid disparate

figures in some saladas, most indicators are integers scaled from one to five. Some scales included zero, negative or non-integer values.

TABLE III

Set of indicators used in this study with their criteria and specific processes. The last two columns show the categories established for each indicator to describe the cases found in the saladas and the rating generated according to their significance.

Indicator	Criterion / process	Category	Rating
visual indicators	Escarpment continuity and height	unrecognizable	0
		smooth	1
		flat bottomed valley	2
		semi-confining	3
		confining	4
	Cropping	totally cultivated	-2
		partially cultivated	1
		not cultivated	5
	Stone dumping	completely covered	1
		in center	2
		in escarpment	3
		absence	5
remotely sensed data	Size	≤ 10 ha	1
		10 - 20 ha	2
		20 - 40 ha	3
		40 - 70 ha	4
		>200 ha	5
	Water occurrence	>75%	1
		50 - 75%	2
		25 - 50%	3
		1 - 25%	4
		0%	5
cartographic information	Proximity to roads in intervals of 200 m	0 - 200	1
		200 - 400	2
		400 - 600	3
		600 - 800	4
		>800	5
	Proximity to the future irrigated area	inside	1
		< 500 m far	1.2
		> 500 m far	1.5

Proposed Indexes

A conservation index (CI) was established as a linear combination of the three visual indicators (Table III) weighted as following:

$$CI = 3 \times \{escarpment\} + 2 \times \{cropping\} + 1.5 \times \{stone \cdot dumping\}$$

where {escarpment}, {crop} and {stone dumping} are the scaled indicators. The weight assigned to each indicator was tested and tuned in successive trials by examining the agreement with the current status of the saladas determined in the field by external expert as part of the *Inventory 2003* (Castañeda and Herrero, *in revision*). The agreement was measured with Spearman coefficient ($P < 0.05$) on 25 saladas, and evaluated with the other 28 saladas.

Two indexes were established for measuring the exposure to degrading forces. First, in order to provide the degree of exposure to deterioration at present, a current vulnerability index (VIc) was created as

$$VIc = \{proximity \cdot to \cdot roads\} + \{size\} + \{water \cdot occurrence\}$$

where {proximity to roads}, {size} and {water occurrence} are the rated indicators. Taking into account the application of water during the life-time of the irrigation district, we predict this negative effect of the forthcoming irrigation in the area. Then, the proximity to the planned irrigated area is introduced (Table III), establishing the predictable future vulnerability index (VI_f) as

$$VI_f = \varphi \times VIc$$

where φ takes one of three values, depending of the location of each salada (Table III).

The resulting values of the above indexes were grouped in intervals using the natural-breaks classification method (Jenks, 1977). Each interval was represented by a code and its corresponding level of the thematic meaning (Table IV).

TABLE IV

Intervals established for the Conservation Index (CI), the Current Vulnerability Index (VIc), and the Future Vulnerability Index (VIf); code assigned, and their thematic value.

Index	Intervals	Code	Thematic value
CI	0.5-5.5	1	very bad
	5.6-8.0	2	bad
	8.1-14.5	3	Conservation status fair
	14.6-22.0	4	good
	22.1-26.5	5	very good
VIc	6.0-8.9	1	low
	9.0-11.9	2	Current vulnerability fair
	12.0-15.9	3	high
VIf	6.0-8.9	1	low
	9.0-11.9	2	Future vulnerability fair
	12.0-15.9	3	high
	16.0-22.5	4	very high

Finally, conservation and vulnerability indexes were linked to obtain an integrated index of the salada condition (SCI). By crossing CI with VIc and VIf, we obtained the current condition (SCIc) and the future condition (SCIf), respectively. Each combination that was encountered of Conservation Index with Current Vulnerability Index and with Future Vulnerability Index was assigned a value (Table V).

TABLE V

In grey shading, values assigned to the saladas condition (SDS) according to a combination of Conservation and Vulnerability values.

Conservation status	Current vulnerability			Future vulnerability			
	low	medium	high	low	medium	high	very high
very bad	*	bad	very bad	*	*	very bad	very bad
bad	*	bad	very bad	*	*	very bad	very bad
fair	*	fair	very bad	*	Bad	very bad	very bad
good	very good	good	fair	very good	good	fair	*
very good	very good	good	fair	very good	good	fair	fair

RESULTS AND DISCUSSION

The state of conservation of the saladas

The {escarpment} had the highest significance as observed in the field: it has favored the preservation of saladas by making it difficult for machinery to invade the salada. {Stone dumping} had the lowest power in differentiating degradation, as it affects all the saladas. {Cropping} had an intermediate level of significance and depends in part on the escarpment.

The CI values had a slight positive relationship with the water occurrence with a Pearson correlation coefficient of $r = 0.78$ ($p = 0.000$) and had no observed relation ($r = 0.39$; $p = 0.15$) to the size of the salada (Fig. 2). Even when the more frequently flooded saladas yield high values of CI, the weak relationships denote an anthropic control of the conservation state rather than a natural one.

[Fig. 2]

From the thematic values (Table IV), 20% of the saladas are in *good* or *very good* conservation (Fig. 3); they are of playa-lake type, holding water at least some time each year. Almost all of them are surrounded by crops and are affected by stone dumping to different degrees. Half of the saladas have a *bad* or *very bad* conservation; these are primarily the smallest saladas and those strongly influenced by cropping; of these, nine have been completely plowed and nine more are unrecognizable because of stone dumping, debris, and construction. Some 30% of the saladas present a *fair* conservation, with a smooth escarpment and the bottom covered by halophytes. Their soils, with an electrical conductivity $> 100 \text{ dS m}^{-1}$ (Herrero, 1982), are considered unproductive by the rural population; this has facilitated their conservation, except for stone dumping.

[Fig. 3]

Vulnerability of the saladas

The current vulnerability index (VIc) represents the common vulnerability for all the saladas since it computes the effects of land consolidation and machinery traffic. When roads are adjacent to the saladas, they increase stone dumping, debris, and earth movements. This is the case of the saladas Amarga Baja and Muerte, where the native vegetation and the dune formed leeward have been partially destroyed. Moreover, other

saladas have been crisscrossed by new roads, disturbing the hydrological function of the wetland.

This VIc index was applied to the 39 saladas detected by remote sensing. Following the thematic values of Table IV, 60% of these saladas have a *high* current vulnerability (Fig. 4); they are the smallest saladas, ranging from 1.4 to 12.2 ha, and have the least water occurrence. They usually have smooth escarpment but halophytes remain in spite of stone dumping and cropping. Some of them are very close to or are within the irrigated area.

[Fig. 4]

We found that 28% of the saladas studied present a *fair* current vulnerability; some are outside the planned irrigation area. Within the irrigated area, saladas Agustín, Benamud and Gramenosa have less water occurrence; they are the most representative of those saladas hosting some of the priority habitats listed by the Habitats Directive (97/62CEE) of the European Union. Only 13% present a *low* vulnerability; they are the largest playas and all but one will remain outside the irrigated area.

The VIf yields a number of saladas exposed to a risk of degradation higher than the VIc (Fig. 5). We found that 73% have a *high* or *very high* prognosis of future vulnerability. Only the largest playas (10%) have a diagnosis of *low* future vulnerability. The predicted vulnerability index value is *fair* for the most frequently flooded wetlands; however, alteration of dry-wet cycles and chemical characteristics can be foreseen for those closely surrounded by irrigated fields. Fig. 5 shows the difference between the current (VIc) and the future (VIf) vulnerability indexes. We predict that almost all the small saladas in the southern part will suffer an increase in their vulnerability, as well as the Agustín, Farnaca and Benamud saladas.

[Fig. 5]

The ongoing irrigation works are producing effects in agreement with the VIf categories (Fig. 6). Much more intense alteration is expected when irrigation starts, due to foreseeable irrigation returns and other water flows.

[Fig. 6]

As shown in Fig. 7, the relationships between the vulnerability indexes (VIf and VIf) and the water occurrence or the size of the saladas yield negative Pearson correlation coefficients. The only significant, but slight, correlations are water occurrence with VIf ($r = -0.73$; $p = 0.000$), and size with VIf ($r = -0.24$; $p = 0.009$). These slight correlations parallel that of the CI (Fig. 2) and highlight the rise of anthropic influence. The highest vulnerability values occur in saladas < 10 ha, which are also those with less water occurrence.

[Fig. 7]

Condition of the saladas

A separate assessment of the conservation status (CI) and the current and future vulnerabilities (VIf and VIf, respectively) can be used for designing conservative or preventive strategies. The integration of these indexes in the current and future condition of the saladas, SCIf and SCIf, respectively, provides (Figs. 8 and 9) a comprehensive qualification of the saladas that can illustrate the effects produced by advancing irrigation. Obviously, the best situation corresponds to those saladas having the highest state of conservation and the lowest vulnerability.

After applying values (Table V), more than a half of saladas, some 58%, have a *bad* or *very bad* current condition. They are the smallest, less than 13 ha, and were without water presence on the dates the satellite images were obtained. Some of them are unrecognizable due to cropping.

[Figs. 8, 9]

A *good* or *very good* current condition was determined for 28% of the saladas. These are the largest playa-lakes. Also, Benamud and Gramenosa are classified as in *good* condition in spite of being dry on the dates studied. From among the 15% of saladas with a *fair* current condition, Agustín, Las Amargas and Valdespartosa are of medium size and they stand out because of their valuable halophilous vegetation, scarcely represented in inland wetlands of Europe.

The prognosis is worse for the saladas located near or within the irrigated area. Some of these saladas are medium sized and have a *very good* conservation; however, since they occupy the lowest topographic positions, they are destined to become drainage ponds, permanently flooded with less saline, polluted waters.

Summarizing, 65% of the saladas, variable in size, have a *bad* or *very bad* predicted future condition; the majority of these are close to or within the irrigated area. The best future condition is predicted for 28% of the lot, among them the following playa-lakes: La Playa, Pueyo, Pito, Guallar, Muerte, Piñol, Camarón, Rebollón, and Rollico. Others such as Salineta, Agustín and Gramenosa will be the most affected by irrigation plans in spite of their current *very good* conservation.

The resulting scores, *good*, *very good*, *bad* and so on do not measure absolute values or have intrinsic meaning, but they do allow for comparisons among saladas and provide an understand of the status and threat to these wetlands. These descriptors can be used to choose conservation measures or to evaluate the resilience of the lakes. The better the condition of the saladas, the less effort required to conserve them. In the case of *very bad* condition, intervention would be useless and ineffective because the saladas are probably unrecoverable. The GIS database includes historical and current homogenized data for these wetlands and is prepared to integrate data from future

surveys. The standardized monitor procedures presented here can help insure that public policies targeting the saladas are in fact achieving their goals.

CONCLUSIONS

The methodology here developed yields a coherent ranking of the saladas according to their conservation and their vulnerability, validated with expert judgment. The use of easily observable features allows for the rapid and, consistent assessment of the condition of the saladas and also provides warning signals of current and future threats to these habitats. Only 20% of the saladas studied are in a *good* or *very good* state of conservation; 30% of them are in *fair* state; half of the saladas are in a *bad* or *very bad* state, primarily the smallest and those largely invaded by cropping. From the saladas studied, 60% of them have *high* current vulnerability; they are the smallest saladas and those with a lower water occurrence; 28% present a *fair* current vulnerability and 13%, the largest playa-lakes, are of *low* vulnerability. The prognosis on the effect of the irrigation plan on these wetlands is that 73% would have a *high* or *very high* future vulnerability. Only 10% have a *low* future vulnerability, the largest playa-lakes of the center and southern sections of the region.

A majority, 58% of saladas, are in *bad* or *very bad* condition. These are those of less than 13 ha, and without water presence on the dates studied; 28% are in *good* or *very good* current condition and 15% are in a *fair* current condition. Finally, 65% of the saladas, variable in size, are predicted to have a *bad* or *very bad* future condition; the majority of this group is close to or within the irrigated area. The best future condition is predicted for 28% of the saladas, found in the southwestern quadrant. Salineta, Agustín, and Gramenosa are in a *very good* conservation state and are *very highly* vulnerable: these lakes can easily be saved. The conservation or the vulnerability are

not reflected in any particular salada feature, such as size or flood frequency, because of the strong agricultural intensification in recent years.

This salada assessment allows us to establish priorities for surveillance, conservation, and restoration purposes. In general, the loss of saladas could be mitigated through the adoption of enhancement and restoration measures. Other criteria could be used depending on the point of view, the social demand, and the resources available to make observations. This methodology can be applied to other saline wetlands under agricultural or other human pressures.

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