

1 **M.R. Yagüe¹ & A.D. Bosch-Serra². 2013. Slurry field management and ammonia**
2 **emissions under Mediterranean conditions. *Short Communication. Soil Use and***
3 ***Management* 29:397-400. DOI: 10.1111/SUM2012.061**

4

5

6 ¹ *Department of Soil & Irrigation (EEAD-CSIC Associated Unit), Agrifood Research*
7 *and Technology Centre Aragon (CITA), Avda. Montañana 930, E-50059, Zaragoza,*
8 *Spain.*

9 ² *Department of Environmental & Soil Sciences, University of Lleida (UdL), Avd.*
10 *Alcalde Rovira Roure 191, E-25198, Lleida, Spain.*

11

12 **Corresponding author:** mryaguue@aragon.es (MR Yagüe).

13

14 **Running title:** Slurry management and ammonia loss

15

Abstract

16
17 In Spain, farmers are interested in applying pig (*Sus scrofa domesticus*) slurry (PS) to
18 their fields throughout the year. During the spring and summer months ammonia (NH₃)
19 volatilization may be high. We studied the potential range of NH₃ losses under a warm
20 and a hot period of the year, using available field practices, and two strategies: PS
21 directly incorporated into the soil, in spring (I-spring); and PS applied by splash-plate,
22 in summer time (SP-summer), both to bare soil. Measurements were conducted, after PS
23 application, using the micrometeorological mass-balance integrated horizontal flux
24 method. The cumulative NH₃-N volatilization was 35% (I-spring) and 60% (SP-
25 summer) of total ammonium nitrogen applied, and half of the total NH₃-N losses
26 happened by 17h and 8h, respectively, after application. Incorporation strategy was less
27 effective in avoiding NH₃ losses than is described in the literature. This fact has
28 important consequences for the implementation of NH₃ mitigation measures in
29 Mediterranean agricultural systems.

30

31 **Keywords:** ammonia-losses, slurry incorporation-method, ammonia volatilization,
32 splash-plate, fertilizer strategies.

33

34 Introduction

35 Spain is the second largest pig producer in Europe and nearly 50% of Spanish pig
36 production is concentrated in the Ebro river valley (MARM, 2013).

37 Pig slurry (PS) is mainly applied to maize (long cycle) at sowing (April-May) and to
38 winter cereals at sowing (October-November) or as a sidedressing (February-March),
39 leaving a gap of 5-6 months when applications are normally avoided. In irrigated areas,
40 a second crop after the cereal harvest or the extension of the maize (short cycle) sowing
41 period can cover this gap.

42 Despite the importance of quantifying European NH₃ emissions, little work done under
43 Mediterranean conditions has been published on the topic (Génermont & Cellier, 1997;
44 Sanz *et al.*, 2010) and existing articles do not cover the whole annual period, nor the
45 range of typical conditions.

46 Our objective was to quantify NH₃ volatilization losses from PS applied according to
47 two different strategies: i) direct incorporation in spring (I-spring), assumed to
48 approximate to minimum likely losses, and ii) splash-plate application in summer time
49 (SP-summer), taken to approximate to maximum likely losses.

50

51 **Materials and methods**

52 Two experiments were conducted in a representative area of the Ebro valley (41° 44'N,
53 0° 49'W, altitude 225 m) on bare ploughed soil (Table 1) before cereal establishment.

54 The first was established on 16-17 May 2007. An incorporation machine was employed.
55 Pig slurry incorporation was by a tube divided into three hoses (12 outlets), with a total
56 application width of 4.80 m. Each outlet was located between two shares, the first one
57 opened a slot in the soil and the one located at the back buried the applied PS at a depth
58 of about 0.15-0.20 m. The second experiment was established on 2-3 August 2007
59 where surface PS spreading was by a tank fitted with a splash-plate; it was spread over
60 the soil without incorporation. For each strategy (Table 2), three replicates were set up
61 plus a control.

62 Weather in the experimental period has two limitations, in May with minimum average
63 temperature (T) and maximum average relative humidity (RH), and in August with
64 maximum T and minimum RH (Fig. 1).

65 The micrometeorological mass-balance integrated horizontal flux method was used,
66 following the description and procedure given by Wood *et al.* (2000). Each rotating

67 mast supported three passive NH₃ flux samplers mounted at three heights (0.375, 0.75,
68 and 1.50 m) with the greatest height being 10% of fetch length, in agreement with
69 estimates by Itier & Perrier (1976), which were confirmed by a previous field test. The
70 ammonium solution obtained in passive samples was analyzed with a continuous flow
71 analyser (AA3–Bran+Luebbe). Sampling started immediately after application and
72 periodicity for the first day was approximately from 1 to 2h, at 3h, from 5 to 8h, at 12h
73 and before 24h after application. Later on, the intensity of sampling decreased with time
74 according to the declining of the intensity of NH₃ flux losses (Fig. 2).
75 Differences between strategies in NH₃-N cumulative emissions and also as a percentage
76 of applied NH₄⁺-N, were analysed using analysis of variance and LSMEANS test (p=
77 0.05). The statistical package SAS V8.2 was used for all statistical analysis.

78

79 **Results and discussion**

80 The average for cumulative NH₃ emissions, measured during the experiments, was 35%
81 of total ammonium nitrogen (TAN) applied (99 kg NH₃-N/ha, 25% of total N) in the I-
82 spring strategy and 60% of applied TAN (122 kg NH₃-N/ha, 42% of total N) in the SP-
83 summer strategy (Table 2). Half of the maximum NH₃-N loss was estimated to occur by
84 17h and 8h after application in I-spring and SP-summer strategies, respectively (Fig. 2).
85 For the SP-summer strategy the emission was in the upper ranges found in other studies
86 (e.g. Sommer *et al.*, 2003; Rochette *et al.*, 2009) but according to Misselbrook *et al.*
87 (2005), the slurry dry matter content in the SP-summer strategy can explain losses
88 equivalent to 60% of applied TAN. Incorporation did not reduce volatilization as much
89 as it was expected to, according to the results from other experiments (Huijsmans *et al.*,
90 2003), probably because dry soil conditions (Table 1) made it difficult to fully bury the
91 PS, favouring NH₃ gas diffusion through it.

92 In the context of Mediterranean agricultural systems, the official advice to bury PS
93 spread over land not later than 24h after application would not be fully effective in
94 reducing NH₃ emissions, as more than 50% of applied TAN could already be lost during
95 the first 24 hours after application. In addition, the low moisture content in the soil, in
96 the hottest months of the year, limits the effectiveness of PS incorporation. New
97 methods must be investigated and they should be orientated to the reduction of slurry-
98 air contact (e.g. trail hoses) or/and to slurry infiltration enhancement (e.g. light
99 irrigation). Nevertheless, their effectiveness will be influenced by soil carbonate content
100 and this aspect needs further practical evaluation. This research is also necessary in the
101 framework of models such as ALFAM (Søgaard *et al.*, 2002) that do not fully cover the
102 special aspects that need to be considered in Mediterranean environments.

103

104 **Acknowledgements**

105 The authors thank Miguel Izquierdo, Jesús Gaudó, Juan M. Acín, and Benjamín Subías
106 for their invaluable field management; Marisa Díez for laboratory assistance; and M.
107 Rosa Teira for assessment of IHF method use. This study was supported by the National
108 Institute for Agricultural and Food Scientific Research and Technology of Spain-INIA
109 (projects RTA2004-114 and RTA2010-126) and by FEDER (European Fund for
110 Regional Development) and FEADER (European Agricultural Fund for Rural
111 Development) funds.

112

113 **References**

- 114 Générumont, S. & Cellier, P. 1997. A mechanistic model for estimating ammonia
115 volatilization from slurry applied to bare soil. *Agricultural and Forest Meteorology*,
116 **88**, 145-167.
- 117 Huijsmans, J.F.M., Hol, J.M.G. & Vermeulen, G.D. 2003. Effect of application method,
118 manure characteristics, weather and field conditions on ammonia volatilization from
119 manure applied to arable land. *Atmospheric Environment*, **37**, 3669-3680.
- 120 Itier, B. & Perrier, A. 1976. Présentation d'une étude analytique de l'advection. I-
121 Advection liée aux variations horizontales de concentration et de température.
122 *Annales Agronomiques*, **27**, 111-140.
- 123 MARM. 2013. Anuario de estadística 2011. Available at:
124 http://www.magrama.gob.es/estadistica/pags/anuario/2011/AE_2011_Completo.pdf
125 ; accessed 13/5/2013.
- 126 Misselbrook, T.H., Nicholson, F.A. & Chambers, F.B. 2005. Predicting ammonia losses
127 following the application of livestock manure to land. *Bioresource Technology*, **96**,
128 159-168.
- 129 Sanz, A., Misselbrook, T.H., Sanz, M.J. & Vallejo, A. 2010. Use of an inverse
130 dispersion technique for estimating ammonia emissions from surface-applied slurry.
131 *Atmospheric Environment*, **44**, 999-1002.
- 132 Sogaard, H.T., Sommer, S.G., Hutchings, N.J., Huijsmans, J.F.M., Bussink, D.W. &
133 Nicholson, F. 2002. Ammonia volatilization from field-applied animal slurry - the
134 ALFAM model. *Atmospheric Environment*, **36**, 3309-3319.
- 135 Sommer, S.G., Générumont, S., Cellier, P., Hutchings, N.J., Olesen, J.E. & Morvan, T.
136 2003. Processes controlling ammonia emission from livestock slurry in the field.
137 *European Journal Agronomy*, **19**, 465-486.

- 138 Rochette, P., Angers, D.A., Chantigny, M.H., MacDonald, J.D., Gasser, M.O. &
139 Bertrand, N. 2009. Reducing ammonia volatilization in a no till soil by
140 incorporating urea and pig slurry in shallow bands. *Nutrient Cycling in*
141 *Agroecosystem*, **84**, 71-80.
- 142 Wood, C.W., Marshall, S.B. & Cabrera, M.L. 2000. Improved method for field-scale
143 measurement of ammonia volatilization. *Communications in Soil Science and Plant*
144 *Analysis* **31**, 581–590.
- 145

146 **Figure Tables**

147

148 **Table 1.** Physico-chemical soil average characteristics (0-0.30 m)

149

150 **Table 2.** Main characteristics^a and rates of the pig slurries applied to each plot
151 according to the two strategies: Incorporation in spring (I-spring) and Splash-plate
152 in summer (SP-summer). Length of measurement periods and absolute emissions
153 are included.

154

155

156 **Figure Legends**

157

158 **Figure 1** (a) Wind speed average (m/s); (b) Relative humidity averages of air (%): maximum
159 (RHmax), medium (RHmed) and minimum (RHmin); (c) Air temperature averages (°C):
160 maximum (Tmax), medium (Tmed) and minimum (Tmin), on a daily basis for a period
161 from 2004 to 2010. The black round symbols (●) are associated to average meteorological
162 data for the period of measurements in each experiment (May and August of 2007). There is
163 a marked contrast between the two measurement periods and no rainfall occurred in
164 either of the periods.

165

166 **Figure 2.** Cumulative ammonia emission (NH₃-N) as a percentage of total ammonium
167 nitrogen applied (NH₄⁺-N, TAN) and measured in two different strategies: slurry
168 incorporated in May and slurry spread (splash-plate method) in August. Both trends
169 were adjusted (***, p<0.001) to the Michaelis-Menten equation
170 [$N_{\text{NH}_3}(t) = N_{\text{max}} * (t / (t + K_m))$], where: N_{NH₃} (% of TAN) is the accumulative

171 ammonia loss at time (t); N_{\max} (% of TAN) is the maximum amount of $\text{NH}_3\text{-N}$ lost
172 and K_m (hours) is the time to reach half of the total losses].

173

174 **Table 1.** Physico-chemical soil average characteristics (0-0.30 m)

Characteristics	Value
pH (potentiometry 1:2.5) ^a	8.3
Humidity (105° C, % w/w) ^b	1.8
Organic matter (Walkley-Black,% w/w)	2.1
Carbonates (Calcimeter Bernard method, % w/w)	39.0
Sand (0.05-2 mm, % w/w)	25.1
Silt (0.002-0.05 mm, % w/w)	53.6
Clay (<0.002 mm, % w/w)	21.3

175 ^a 1:2.5; 1 soil: 2.5 distilled water (v/v).

176 ^b Soil water content was similar in both periods of slurry application.

177

178 **Table 2.** Main characteristics^a and rates of the pig slurries applied to each plot

179 according to the two strategies: Incorporation in spring (I-spring) and Splash-plate in

180 summer (SP-summer). Length of measurement periods and absolute emissions are

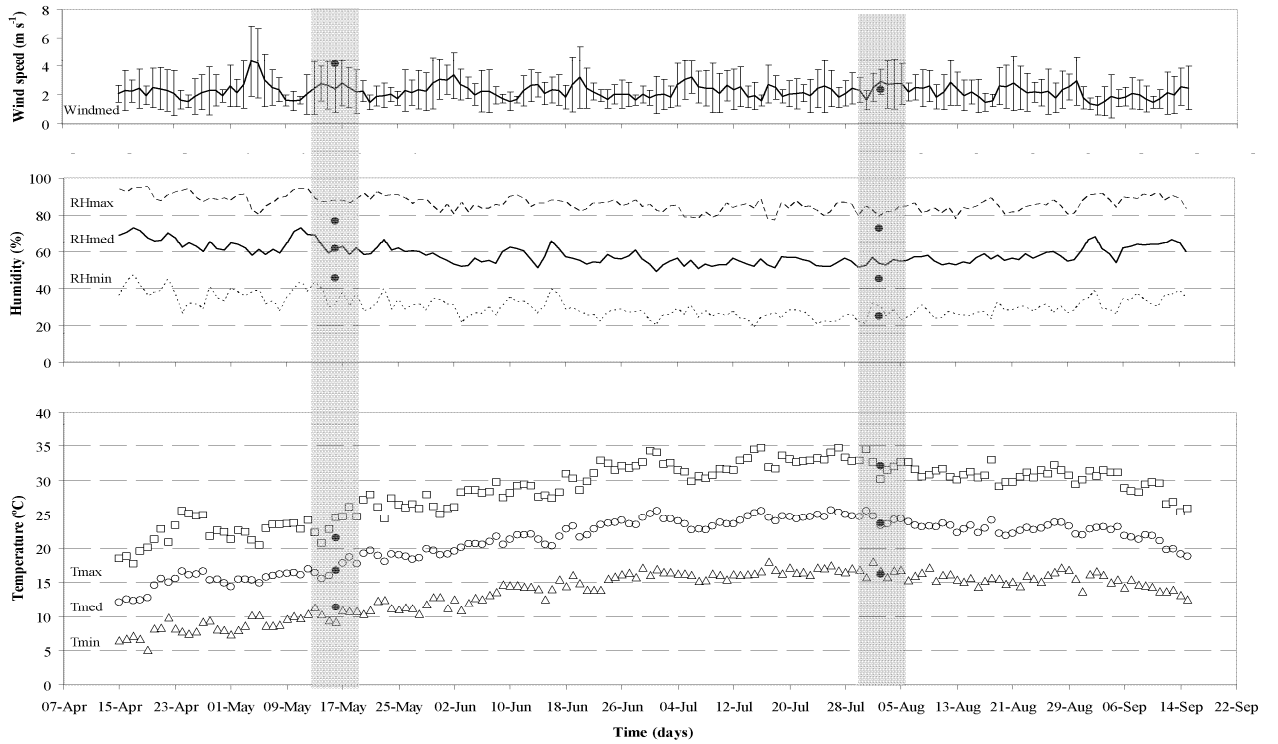
181 included.

Strategy	pH (1:5)	Dry matter (kg/t)	Organic matter (kg/t)	Ammoniacal-N (kg NH ₄ ⁺ -N/ha)	Total N (kg N/ha)	Rate (t/ha)	Sampling period (days)	NH ₃ -N loss (kg N/ha)
I-Spring	8.2	84.8	59.5	267	377	53	6 days	100.7
I-Spring	8.2	84.8	59.5	302	427	60	7 days	97.6
I-Spring	8.2	84.8	59.5	272	385	54	6 days	99.0
SP-Summer	8.1	86.4	60.2	168	239	36	10 days	102.0
SP-Summer	8.1	86.4	60.2	254	359	54	9 days	157.9
SP-Summer	8.1	86.4	60.2	187	266	40	10 days	106.3

182 ^a pH by potentiometry (1:5; 1 pig slurry: 5 distilled water); dry matter by
 183 gravimetric analysis at 105°C; organic matter by calcination at 550°C,
 184 Ammoniacal-N by modified Kjeldahl method and total N by Kjeldahl method.

185

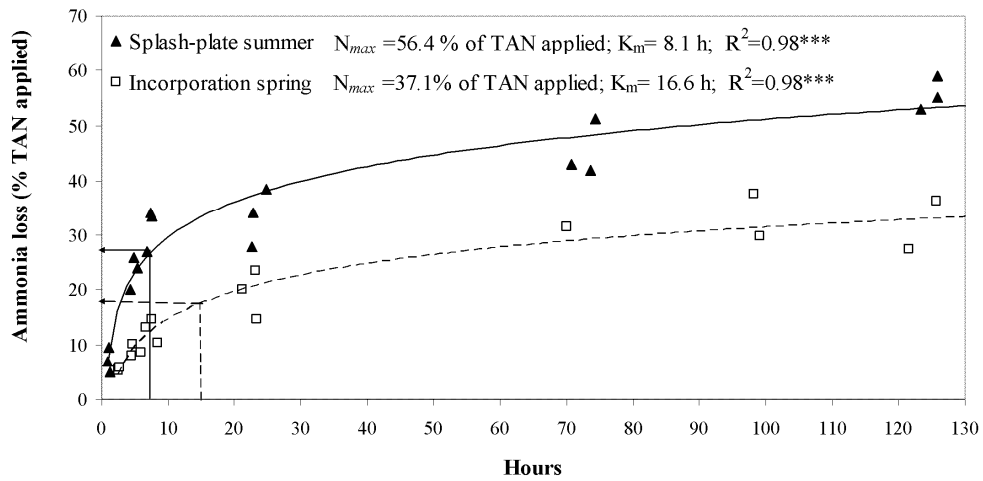
186



187

188 **Fig. 1**

189



190

191 **Fig. 2**