

# Adapted conservation measures are required to save the Iberian lynx in a changing climate

D. A. Fordham<sup>1</sup>, H. R. Akçakaya<sup>2</sup>, B. W. Brook<sup>1</sup>, A. Rodríguez<sup>3</sup>, P. C. Alves<sup>4,5</sup>, E. Civantos<sup>4,6</sup>, M. Triviño<sup>6,7</sup>, M. J. Watts<sup>1,8</sup> and M. B. Araújo<sup>6,9,10\*</sup>

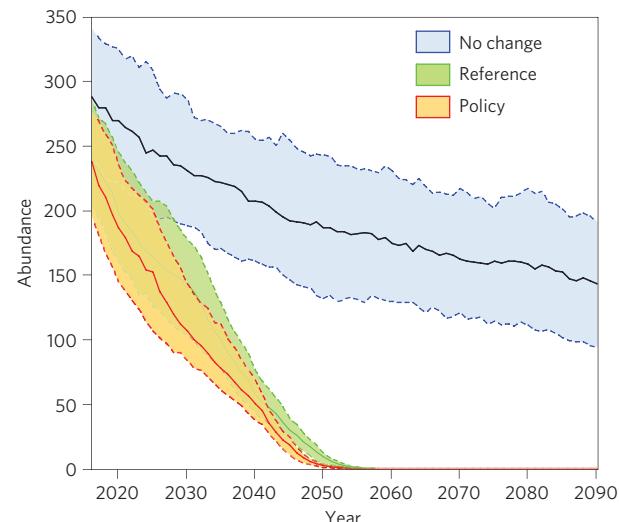
The Iberian lynx (*Lynx pardinus*) has suffered severe population declines in the twentieth century and is now on the brink of extinction<sup>1</sup>. Climate change could further threaten the survival of the species<sup>2</sup>, but its forecast effects are being neglected in recovery plans<sup>3,4</sup>. Quantitative estimates of extinction risk under climate change have so far mostly relied on inferences from correlative projections of species' habitat shifts<sup>5</sup>. Here we use ecological niche models coupled to metapopulation simulations with source-sink dynamics<sup>6,7</sup> to directly investigate the combined effects of climate change, prey availability and management intervention on the persistence of the Iberian lynx. Our approach is unique in that it explicitly models dynamic bi-trophic species interactions in a climate change setting. We show that anticipated climate change will rapidly and severely decrease lynx abundance and probably lead to its extinction in the wild within 50 years, even with strong global efforts to mitigate greenhouse gas emissions. In stark contrast, we also show that a carefully planned reintroduction programme, accounting for the effects of climate change, prey abundance and habitat connectivity, could avert extinction of the lynx this century. Our results demonstrate, for the first time, why considering prey availability, climate change and their interaction in models is important when designing policies to prevent future biodiversity loss.

The Iberian lynx is the world's most endangered cat. Its range contracted from 40,600 km<sup>2</sup> in the 1950s to 1,200 km<sup>2</sup> in 2005<sup>1,8</sup>. Recent count estimates only 250 Iberian lynx in the wild in 2005<sup>1,8</sup>. This major decline is associated closely with sharp reductions in European rabbit (*Oryctolagus cuniculus*) abundance<sup>8</sup>, caused by myxomatosis virus in the 1950s and, more recently, rabbit haemorrhagic disease<sup>9</sup> (RHD). Over-hunting of rabbits and the modification and fragmentation of mixed grassland and forest habitat has exacerbated the problems of prey scarcity, because rabbits account for more than 80% of Iberian lynx consumption<sup>10</sup>. Non-natural mortality of Iberian lynx such as trapping, poaching and road kills have further reduced lynx population numbers<sup>11</sup>. Only two Iberian lynx populations persist in the wild at present compared with nine in the 1990s<sup>1</sup>.

*Ex situ* captive breeding programmes have been launched to facilitate the reintroduction of a genetically diverse pool of Iberian lynx into suitable areas within their historical range<sup>12</sup>. Simultaneously, management effort is being directed

towards increasing the carrying capacity of reintroduction sites through habitat restoration, relocating rabbits and limiting direct anthropogenic-related fatalities<sup>4,13</sup>. Although a well-financed effort to avert the extinction of this charismatic species is underway (> €94 million funding since 1994)<sup>4</sup>, non-accounted threats such as climate change and its influence on prey abundance, are not being considered in recovery plans.

Here we provide the most comprehensive analysis of the likely effects of climate change yet for a threatened vertebrate. So far, models used to investigate how climate change will



**Figure 1 | Forecast lynx abundance in the Iberian Peninsula from 2015 to 2090 under three climate change scenarios.** Scenarios are: no climate change—where temperature and precipitation remains unchanged from the year 2000; high CO<sub>2</sub> concentration Reference scenario (WRE750); and an alternative low CO<sub>2</sub> concentration Policy scenario that assumes strong mitigation of greenhouse gas emissions (LEV1). The solid lines show mean estimates for each scenario. Band widths represent 5th and 95th percentiles of population size derived from a stochastic metapopulation model with spatial habitat dynamics. Model variation characterizes demographic stochasticity, which is only one component of model uncertainty (see Supplementary Methods).

<sup>1</sup>Environment Institute and School of Earth and Environmental Sciences, University of Adelaide, North Terrace, South Australia 5005, Australia,

<sup>2</sup>Department of Ecology and Evolution, Stony Brook University, Stony Brook, New York 11794, USA, <sup>3</sup>Department of Conservation Biology, Estación

Biológica de Doñana, CSIC, Américo Vespucio s/n, 41092 Seville, Spain, <sup>4</sup>CIBIO, Centro de Investigação em Biodiversidade e Recursos Genéticos &

Faculdade de Ciências, Universidade do Porto, 4485-661 Vairão, Portugal, <sup>5</sup>Wildlife Biology Program, College of Forestry and Conservation, University of

Montana, Montana 59812, USA, <sup>6</sup>Department of Biogeography and Global Change, National Museum of Natural Sciences, CSIC, 28006 Madrid, Spain,

<sup>7</sup>Department of Biological and Environmental Sciences, University of Jyväskylä, PO Box 35 40014 Jyväskylä, Finland, <sup>8</sup>Information Technology Programme,

AIS St Helens, PO Box 2995, Auckland 1140, New Zealand, <sup>9</sup>Rui Nabeiro Biodiversity Chair, CIBIO, University of Évora, Évora, 7000, Portugal, <sup>10</sup>Center for

Macroecology, Evolution and Climate, University of Copenhagen, 2100 Copenhagen, Denmark. \*e-mail: maraujo@mncn.csic.es

**Table 1 | Extinction risk for the endangered Iberian lynx in the twenty-first century according to different climate and conservation scenarios.**

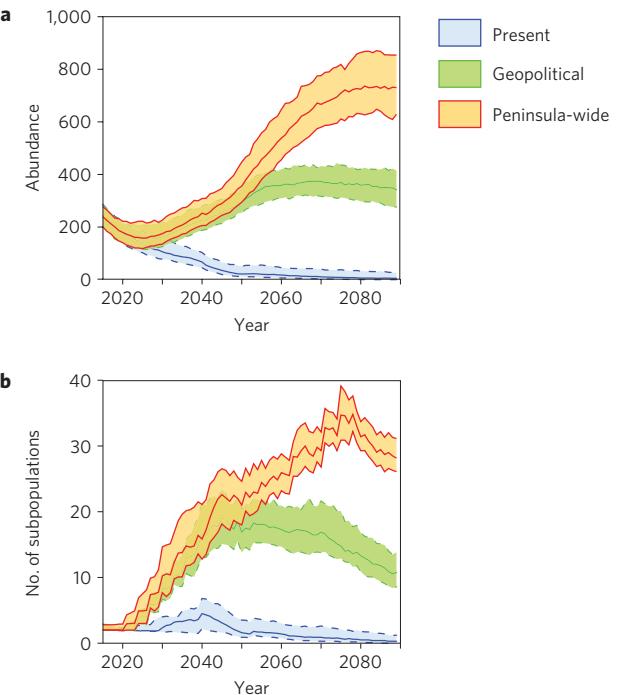
Conservation scenario	Climate scenario	EMA	Probability of extinction	Median time to extinction
No extra manag.	No change	58	12	—
	Policy	0	100	2051
	Reference	0	100	2054
Extra manag.	Policy	3	89	2066
	Reference	3	86	2065
Geopolitical	Policy	52	2	—
	Reference	52	4	—
Peninsula-wide	Policy	68	0	—
	Reference	78	1	—

Climate scenarios are a 'No change scenario' where temperature and precipitation remains unchanged from the year 2000, a stabilizing Policy scenario that assumes substantive mitigation of greenhouse gas emissions and a high CO<sub>2</sub> concentration Reference scenario. Conservation scenarios are: no additional management intervention beyond what is occurring in Doñana and Sierra Morena (No extra manag.); an increase in active population management in potentially suitable habitat patches across the Iberian peninsula (Extra manag.); moving animals with an underlying aim to establish viable lynx breeding populations in every autonomous region within its recent historical range (Geopolitical); moving animals to areas of most favourable habitat across the entire Iberian peninsula (Peninsula-wide). Extinction risk during this century is measured using expected minimum abundance (EMA), probability (%) of total population size declining to zero (Probability of extinction) and median time to extinction (Median time to extinction).

affect biodiversity at broader geographical scales have failed to account explicitly for dynamic feedbacks of species interactions<sup>14</sup>. We address this key limitation using coupled ecological niche–population models with source–sink dynamics<sup>6,7</sup>, simulating the stochastic demographic responses of Iberian lynx to: spatial patterns of rabbit abundance conditioned by disease, climate and land-use variation; and changes in climate suitability and landscape modification. We consider the interaction between rabbits and Iberian lynx as unidirectional, because lynx are extremely rare, whereas rabbits are abundant and have greater than 30 vertebrate predators, many of them widely distributed and locally abundant<sup>15</sup>. Rare efforts to account for species interactions in climate–ecological forecasts have used overly simple approaches; by adding an interacting species as an additional predictor in a correlative model, or by restricting the distribution of one species to the modelled distribution of the other<sup>16</sup>. Here we directly account for important per capita effects of rabbit abundance on the population growth rate of Iberian lynx in a climate change setting.

We show that climate change is predicted to have a rapid and severe negative influence on Iberian lynx abundance (Fig 1), exceeding its ability to adapt and disperse to more climatically favourable regions where prey densities are sufficient to support viable populations. We estimate time to extinction to be less than 50 years (Table 1), even with rapid and deep global cuts to anthropogenic greenhouse gas emissions (for instance, stabilization at equivalent CO<sub>2</sub> concentrations of 50 ppm (MiNiCAM LEV1; ref. 17)) if forecasted population decline will occur slightly faster than under a strong mitigation Policy scenario (Fig 1) causing slightly shorter median time to extinction (2051 compared to 2054 under the Reference scenario; Table 1). This apparently counter-intuitive result is because coal-fired power stations also produce atmospheric aerosols that suppress global temperature through dimming<sup>18</sup>. Replacing oil combustion with cleaner energy alternatives results in conditions that are initially more adverse (warmer and drier before 2050) for Iberian lynx and rabbits under the Policy scenario compared with high-CO<sub>2</sub> concentration Reference scenario (WRE750; ref. 19).

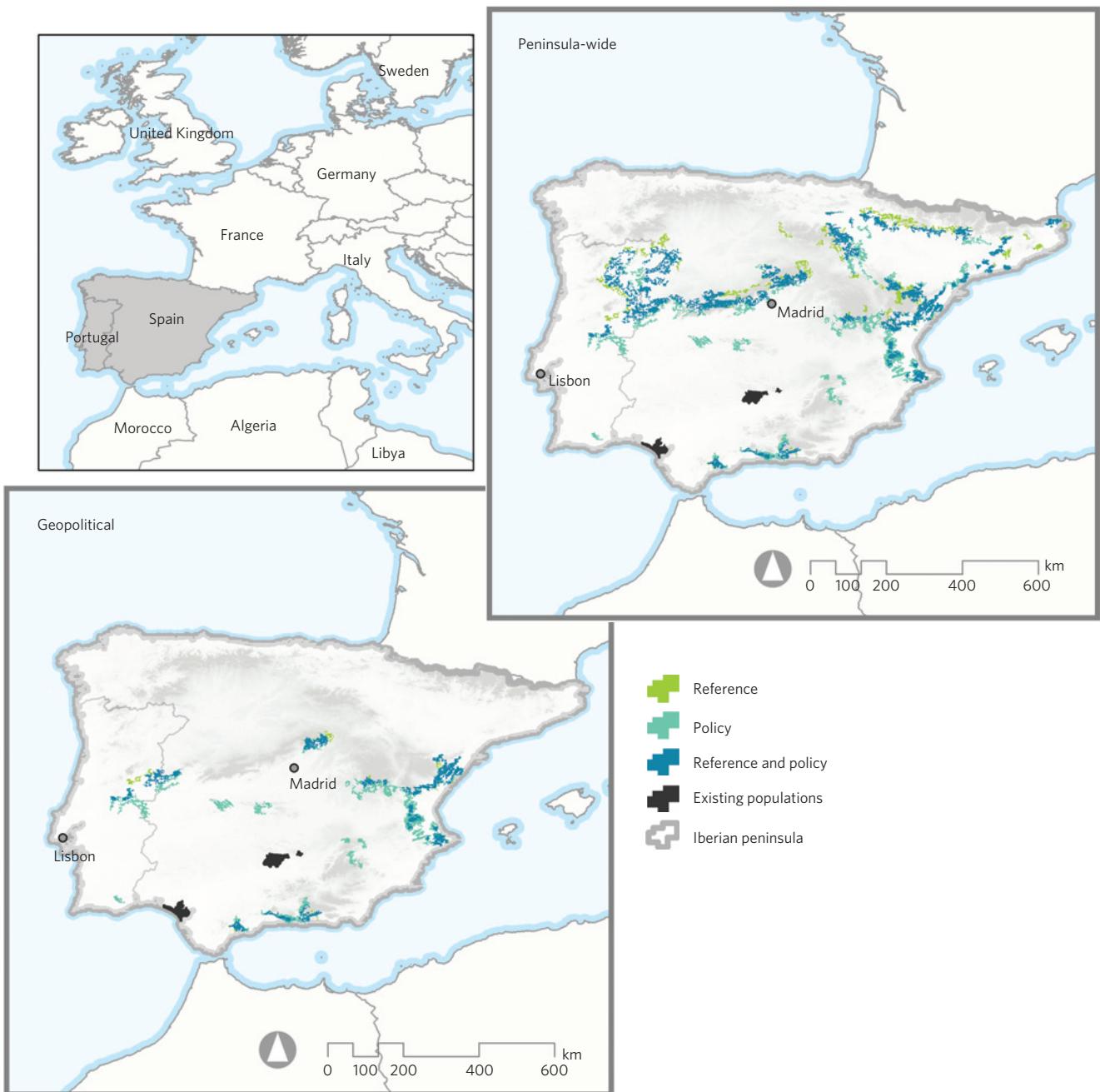
In trying to avert extinctions in the wild, managers have done considered the use of human-assisted colonization of individuals



**Figure 2 | The effect of management intervention on the persistence of Iberian lynx.** **a,b**, Forecast lynx abundance (**a**) and number of populations (**b**) in the Iberian Peninsula from 2015 to 2090 according to three possible management options. Forecasts are for a Policy climate scenario that assumes strong mitigation of greenhouse gas emissions (LEV1). The interventions are: present-day conservation practices, including increasing prey (lagomorph) densities, habitat alteration, preventing disease and non-natural mortality (Present); reintroducing captive-bred lynx to unoccupied habitat according to a Geopolitical scheme that favours establishing lynx populations in every autonomous region in Spain (Andalucía, Castilla-La Mancha, Comunidad Valenciana, Extremadura and Murcia), plus Portugal as an additional region, within its recent historical range (Geopolitical); and a strategy focused on releasing animals into the best-quality habitat regardless of region (Peninsula-wide). The solid lines show mean estimates for each scenario. Band widths represent 5th and 95th percentiles, and characterize variation due to demographic stochasticity. They do not include error propagating from uncertainties in demographic parameters or forecasts of climate suitability (see Supplementary Methods).

to more suitable habitat<sup>20</sup>. However, as a climate adaptation strategy for Iberian lynx, such action has been criticized as inadequate if important factors relating to prey and habitat availability are ignored<sup>21</sup>. To avoid this simplification, we simulated reintroductions that accounted simultaneously for the velocity, extent and nonlinearities in rates of climate-induced change, habitat quality and connectivity, prey availability and physiologic conditions important to Iberian lynx. We show that a yearly release of six animals (aged between 1 and 4 years) of each sex into habitat patches ranked according to a combination of carrying capacity, initial population size, connectivity and rate of survival and fecundity would avert likely extinction of Iberian lynx this century (Fig 2 and Table 1 and Supplementary Fig S1). This reintroduction number is lower than that considered under the Iberian lynx reintroduction plan (release of 20–40 animals per year) and would not jeopardize the persistence of the breeding population (see Supplementary Methods).

In contrast, we show that implementing present-day conservation practices (increasing prey densities, habitat management, preventing disease and non-natural mortality)<sup>1,4</sup> at a regional level



**Figure 3 | Forecast location of lynx populations in the Iberian Peninsula in 2090.** Occupied habitats are shown for the Peninsula-wide and Geopolitical reintroduction scenarios and two climate change scenarios: a high-CO<sub>2</sub> concentration stabilizing Reference scenario (WRE750) and an alternative Policy scenario that assumes strong mitigation (LEV1). Maps capture lynx demographic responses to spatial patterns of rabbit abundance (conditioned by disease, climate and environmental variation) and changes in climate suitability and landscape modification. Only grid cells where lynx were present in 90% of stochastic iterations of the demographic model were treated as populated. See Supplementary Information and Methods for further details.

wouldn't serve to delay mediartimextinction by <15 years, irrespective of climate policy (Table 1). This is primarily because extensive clearing and destruction of natural habitat by humans has disrupted processes that underpin Iberian lynx dispersal and establishment. We predict that >40 discrete areas of suitable habitat would be available for potential colonization in 2050 but in the absence of managed reintroductions most of these would remain unoccupied because of limited connectivity caused by human modification of the landscape.

Policymakers are considering the feasibility of establishing viable Iberian lynx breeding populations in every autonomous region within its recent historical range.<sup>22</sup> We show however that

constraining reintroduction to only those suitable patches within the recent historical range of Iberian lynx (Supplementary Fig S2), and spreading animals evenly across autonomous regions using Geopolitical reintroduction strategy results in fewer Iberian lynx living in fewer populations in the future compared with a Peninsula-wide reintroduction strategy where animals are released into the best-quality habitat regardless of region (Fig 2 and Supplementary Fig S1). We predict final population sizes in 2090 to be between 190 and 275 animals (5th and 95th percentiles, respectively), approximately the same as today's population size<sup>4</sup>, living in 7–10 subpopulations for the Geopolitical reintroduction strategy, compared with 654–896 animals living in 25–31 subpopulations.

for the Peninsula-wide strategy. Uncertainties in these estimates are due to demographic variation, driven by multiple stochastic runs of the meta-population model and the coupled dynamics of the rabbit–lynx bi-trophic interaction. They do not include error propagating from uncertainties in demographic parameters or forecasts of climate suitability (see Supplementary Methods for a detailed description of model uncertainty). The reason for this large difference is that the Peninsula-wide strategy considers habitats in the northern half of Iberia as potential refugia from climate-induced shifts in physiological conditions, rabbit abundance and habitat availability (Fig 3 and Supplementary Fig. S3). However, even then, a negative imbalance between the rate of establishment of new potential locations sites and the extirpation of old expected to occur in the Peninsula-wide strategy after 2065 (Fig 2). Thus in the future, managers might need to consider moving Iberian lynx beyond their Pleistocene refugia where they evolved<sup>23</sup>, to other parts of Europe where the biophysical characteristics of the new location better match the needs of the species and its prey. For example, fossil data show that the Iberian lynx roamed near relatives were distributed in southeastern France in the Pleistocene<sup>8</sup>.

Another potential management option is to improve the quality, connectivity and permeability of the non-habitat land-scapematrix of Iberian lynx's present and future habitats by developing habitat corridors along climatic pathways<sup>25</sup>. However, the immediacy of climate threat faced by Iberian lynx (high probability of extinction in the wild within the next 60 years Fig 1) paired with both the tremendous costs associated with establishing large-scale movement corridors<sup>26</sup> and the technical feasibility of translocating Iberian lynx<sup>4</sup>, implies strongly that the dispersal processes that have been disrupted by loss of habitat connectivity can be most cost-effectively and quickly restored using a fully managed reintroductions program<sup>24</sup>.

Our results suggest that Iberian lynx abundance is declining (Fig 1, see also ref. 1). However, lynx abundance at Doñana and Sierra Morena (the two extant populations) may have increased in recent years in response to intensive management operations<sup>4</sup>. Recent conservation measures have included efforts to decrease anthropogenic mortality among dispersing Iberian lynx and broker conservation agreements with private landholders<sup>4</sup>, two management actions not fully captured in our models, both of which could potentially influence trajectories of population growth in the absence of global warming. Our systematic sensitivity analysis (see Supplementary Information), which identified primary sources of demographic parameter uncertainty in model predictions, supports this premise. Although our forecast final mean and expected minimum abundance estimates for Iberian lynx proved to be most sensitive to the frequency of outbreaks of feline leukaemia virus (see Supplementary Methods), lynx survival (aged  $\geq 2$  years) was the second most important parameter in the model (Supplementary Table S1). Although it remains unclear whether management intervention in the two extant populations of Iberian lynx in Doñana and Sierra Morena is definitely reversing population declines<sup>1,4,22</sup>, present-day efforts will be insufficient for achieving the long-term conservation of Iberian lynx. Our dynamic meta-population model results clearly show that habitat connectivity today is inadequate to support sufficient range movements for an effective natural-adaptation response to the velocity of environmental change forecast for Iberian lynx this century. Refining field estimates of Iberian lynx survival will strengthen model precision and should be a priority for future Iberian lynx research.

The risk of extinction faced by Iberian lynx within the next 50 years is high. Although existing breeding programmes and reintroduction plans promote the persistence of Iberian lynx (at least in the near term), we have demonstrated that the implementation of a carefully planned relocation strategy,

accounting simultaneously for the effects of climate change, prey availability conditioned by disease and habitat connectivity, will be crucial, irrespective of any global decision on mitigating greenhouse gas emissions.

## Methods

We used coupled Iberian lynx and European rabbit<sup>7</sup> demographic models to simultaneously model Iberian lynx and European rabbit range dynamics and climate change. We generated annual time series of climate change daysof the year, rainfall and near temperature<sup>7</sup> that were used to estimate monthly birth rates for the period 2000–2100 using a geographically skilled ensemble-averaged climate forecast<sup>27</sup>. Occurrence records (since 1950) were used to fit ecologic niche models (ENMs) separately for Iberian lynx and European rabbit models, which were parameterized to represent present climatic–habitat preferences and forecast potential future distributions. The rational is that ENM-modelled suitability provides a surrogate for species carrying capacity<sup>28</sup>, capturing northerly physiological constraints that define presence/absence at a given location, which are used in demographic models<sup>7</sup>. Projections from alternative ENMs can vary substantially under the same climate change scenario and were computed to evaluate different ENM approaches (BIOCLIM, Mahalanobis and Euclidean distances, generalized linear models, random forest, maximum entropy and GARP) and obtained consensus forecasts by assigning equal weights to a subset of skilled methods. We separately identified CORINE land-use categories (centred by year 2000) appropriate for breeding habitat for Iberian lynx and rabbit occupancy (Supplementary Table S2 and S3). Land-use modification has a strong historical legacy on the distribution of Iberian lynx and rabbits<sup>8</sup>; this information was used to constrain the suitability of ENM predictions both for the present-day and future<sup>29</sup>. Our projection assumes that lynx and rabbits will conserve their climatic–habitat preferences in the future and that land-use remains spatially constant.

We built a demographic projection for rabbits with grid-based spatial structure, whereby each grid cell was modelled with its own scalar–stochastic model. The carrying capacity of rabbits in each cell was based on ENM predictions, scaled by the maximum and minimum recorded annual rabbit abundance per hectare. The impact of myxomatosis on rabbit abundance was modelled explicitly, whereas the impact of RHD was modelled explicitly. Diseases were modelled as temporally uncorrelated random interactive because empirical data directly describing disease interactions and their effect on rabbit vital rates across a wide spectrum of rabbit densities are not available. The account of uncertainty in RHD estimates (0.0 rabbit model) was built with different grid cell frequencies of RHD occurrence and severity (sampled from within the likely upper and lower parameter bounds) and using glidation under both Reference and Policy CO<sub>2</sub> concentration scenarios and no-climate change scenario where temperature and precipitation remained unchanged from the year 2000. Using these outputs, we mapped rabbit density each year (2000–2100) for each model, capturing stochasticity in interaction between rabbit demography, disease and source–sink dynamics.

Following the rabbit modelling step, we built a spatially structured metapopulation model of Iberian lynx where each subpopulation was represented with a hex-structured stage-structured stochastic model. This is a grid of locations, subpopulations, which were generated using ENM predictions for each spatial pattern of rabbit abundance<sup>7</sup>. Previous step had estimated the ground management effort. The density-dependent sub-model accounted for individuals within each grid cell and modified survival, viability and fecundity functions of rabbit density. Population located outside protected areas were modelled with an additional 0% fixed increase in mortality. Dispersal accounted for the interaction between movement rates and density over the modelled area. The probability of an outbreak of feline leukaemia virus as a per-population catastrophe.

Using single-population lynx model, we calculated sustainable numbers of young animals aged between 1 and 4 years that could be removed annually from a captive breeding population of lynx. We evaluated and ranked which patches to release Iberian lynx into the basis of composite metrics that accounted for important ecological processes. Reintroductions were spread wherever possible, even across autonomous regions (Supplementary Fig S2) for the Geopolitical strategy – targeting the most suitable patch of habitat in each of the autonomous regions for three consecutive years, where a forth Peninsula-wide strategy, the two best ranked habitats were used regardless of provincial location. Vulnerability of lynx during the 80-year interval 2010 and 2090 was measured using expected minimum abundance probability of a population size declining to zero and median time to extinction for persistent models<sup>30</sup>. Annual average abundance and metapopulation patch occupancy were also used to provide a measure of population stability.

The development of the complexity in the rabbit and Iberian lynx models reflected a balance between the need for management actions to be based on a realistic model that does not exclude major biological processes and making the model as robust as possible to underlying uncertainties. Bayesian approaches can be used to explore mode error in niche–population models<sup>30</sup>, however using them to propagate uncertainty in complex models such as the Iberian lynx model, would be computationally demanding. Furthermore, independent evaluation

of such models is extremely difficult<sup>7</sup> and beyond the scope of this study. We report uncertainties in Iberian lynx abundance derived from the internal dynamics of the stochastic population model and the coupled dynamic of the rabbit–lynx relationship. We used sensitivity analysis to determine how the rabbit model was sensitive to assumptions surrounding disease carrying capacity and population growth rates and whether assumptions surrounding spatial and/or non-spatial demographic parameters largely influenced the lynx model. A more detailed explanation and justification of what and method is given in the Supplementary Methods.

## References

1. Palomares, F., Rodríguez, A., Revilla, E., López-Bao, J. & Calzada, J. Assessment of the conservation efforts to prevent extinction of the Iberian lynx. *Conserv. Biol.* **25**, 4–8 (2011).
2. Thomas, C. D. Translocation of species climate change and hindcasting to recreate ecological communities. *Trend Ecol Evol.* **26**, 216–221 (2011).
3. Gil-Sánchez, J. M., Arenas-Rojas, R., García-Tardío, M., Rodríguez-Siles, J. & Simón-Mata, M. A. Habitat assessment to release or reintroduce the endangered Iberian lynx. *Wildl Biol Pract.* **7**, 1–19 (2011).
4. Simón-Mata, M. A. Reversing the decline of the endangered Iberian lynx. *Conserv Biol.* **26**, 731–736 (2012).
5. Thomas, C. D. Extinction risk from climate change. *Nature* **427**, 145–148 (2004).
6. Keith, D. A. & Delibes, M. Predicting extinction risk under climate change: Coupling stochastic population models with dynamic bioclimatic habitat models. *Biol Lett.* **4**, 560–563 (2008).
7. Fordham, D. A., Akçakaya, H. R., Araújo, M. B., Keith, D. A. & Brook, B. W. Tools for integrating range change and extinction risk under climate change information into conservation management. *Ecohydrology* <http://dx.doi.org/10.1111/j.1600-0587.2013.00147.x> (2013).
8. Rodríguez, A. & Delibes, M. Internal structure and patterns of contraction in the geographic range of the Iberian lynx. *Ecohydrology* **25**, 314–328 (2002).
9. Delibes-Mateos, M., Ferreras, P. & Villafuerte, R. European rabbit population trends and associated factors: A review of the situation in the Iberian Peninsula. *Mammal Rev.* **39**, 124–140 (2009).
10. Delibes, M., Rodríguez, A. & Ferreras, P. Action Plan for the Conservation of the Iberian Lynx (*Lynx pardinus*) in Europe. *Nature and Environment Series* (Vol. 111) Council of Europe Publishing (2000).
11. Rodríguez, A. & Delibes, M. Patterns and causes of non-natural mortality in the Iberian lynx during a 10-year period of range contraction. *Biol Conserv.* **118**, 151–161 (2004).
12. Vargas, A., et al. The Iberian lynx *Lynx pardinus* conservation breeding program. *Int Zool Yearb.* **42**, 190–198 (2008).
13. López-Bao, J. V., Rodríguez, A. & Palomares, F. Behavioural responses of trophic specialists to the Iberian lynx: supplementary food patterns, food use and implications for conservation. *Biol Conserv.* **141**, 1857–1867 (2008).
14. Bellard, C., Bertelsmeier, C., Leadley, P., Thuiller, W. & Courchamp, F. Impacts of climate change on the future of biodiversity. *Ecol Lett.* **15**, 365–377 (2012).
15. Delibes-Mateos, M., Delibes, M., Ferreras, P. & Villafuerte, R. Key role of European rabbits in the conservation of the western Mediterranean basin hotspot. *Conserv Biol.* **22**, 1106–1117 (2008).
16. Kissling, W. D. & Jetz, W. Towards novel approaches to modelling biotic interactions in multi-species assemblages at large spatial extents. *J Biogeogr.* **2163–2178** (2012).
17. Wigley, T. M. L. Uncertainties in climate stabilization. *Climatic Change* **97**, 85–121 (2009).
18. Wigley, T. M. L. Could reducing fossil-fuel emissions cause global warming? *Nature* **349**, 503–506 (1991).
19. Wigley, T. M. L., Richels, R. & Edmonds, J. A. Economic and environmental choices in stabilizing atmospheric CO<sub>2</sub> concentrations. *Nature* **379**, 240–243 (1996).
20. Griffith, B., Scott, J. M., Carpenter, J. W. & Reed, C. Translocation as a species conservation tool: Status and strategy. *Science* **245**, 477–480 (1989).
21. Vilà, M. & Hulme, P. Jurassic Park? Thanks. *Trend Ecol Evol.* **26**, 497–498 (2011).
22. Rodríguez, A., Calzada, J., Revilla, E., López-Bao, J. V. & Palomares, F. Bringing science back to the conservation of the Iberian lynx. *Conserv Biol.* **26**, 737–739 (2012).
23. Ferrer, M. & Negro, J. J. Threats to extinction of wild European predators: Super-specialist vs. generalist. *Conserv Biol.* **18**, 344–349 (2004).
24. Hoegh-Guldberg, O. et al. Assisted colonization and rapid climate change. *Science* **321**, 345–346 (2008).
25. Alagador, D. et al. Linking likelihood with like: Optimising connectivity between environmentally similar habitats. *Landscape Ecol.* **27**, 291–301 (2012).
26. Fischer, C. *Aetiology, Impacts, Adaptation and Vulnerability* (ed Parry, M. L.) 211–272 (Cambridge Univ Press, 2007).
27. Fordham, D. A., Wigley, T. M. L., Watts, M. J. & Brook, B. W. Strengthening forecasts of climate change impacts with multi-model ensemble averaged projections using MAGICC/SCENGEN 3. *Ecohydrology* **35**, 4–8 (2012).
28. Van Der Wal, J., Shoo, L. P., Johnson, C. N. & Williams, S. E. Abundance and environmental niche: Environment suitability estimated from niche models predicts the upper limit to local abundance. *Am Natur.* **174**, 282–291 (2009).
29. Stanton, J. C., Pearson, R. G., Horning, N., Ersts, P. & Akçakaya, H. R. Combining static and dynamic variables in species distribution models under climate change. *Methods Ecol Evol.* **3**, 349–357 (2012).
30. Pagel, J. & Schurr, F. M. Forecasting species range by statistical estimation of ecological niche and spatial population dynamics. *Glob Ecol Biogeogr.* **21**, 293–304 (2012).

## Acknowledgements

This research was initiated at a workshop at the National Museum of Natural Sciences, Madrid, led by M. B. A. and funded by FP6-MACISSP/STREP/0778 Australian Research Council (ARC) supported D. A. F. B. W. Band, M. J. W. (LP0989420/DP1096427, FT100100200 and FS110200051), NSANNX09AK19G and NSDEB-1146198 supported H. R. A. F. C. I. PTDC/AAC-AMB/98163/2008 Integrated Program of C&DT ND/SAESCTN/ALENTE-07-0224-FEDER-001755 and Danish NSI supported M. B. A.; Spanish Research Council (CSIC) grant E008301195 and the Andalusian Department of Innovation and Science (P06-RNM-01903) supported R. REVIRGÓ and C. Ferreira provided further advice on rabbit distribution and abundance. C. Mellin and N. C. Lisby generated the figures.

## Author contributions

The design of this project was the result of discussions involving all the authors. M. B. A. generated the ENMs. D. A. F. B. H. R. A. A. R. P. C. A. E. C. and M. T. coupled the ENMs to metapopulation simulations. D. A. F. B. H. R. A. B. W. Band, M. J. W. performed the analysis. D. A. F. and M. B. A. wrote the initial draft of the manuscript and all authors contributed to the writing of the final version of the paper. All authors discussed the results and commented on the manuscript.

## Additional information

Supplementary information is available online at the online version of the paper. Reprints and permissions information is available online at [www.nature.com/reprints](http://www.nature.com/reprints). Correspondence and requests for materials should be addressed to M. B. A.