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Potential soil carbon sequestration in a semiarid Mediterranean  
agroecosystem under climate change: Quantifying management and climate  
effects

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24 **Keywords**

25 Soil Organic Carbon; Climate Change; Modeling; Tillage; Mediterranean systems

26

27 **Abstract**

28 Climate change is projected to significantly impact vegetation and soils of managed  
29 ecosystems. In this study we used the ecosystem Century model together with climatic  
30 outputs from different atmosphere-ocean general circulation models (AOGCM) to study  
31 the effects of climate change and management on soil organic carbon (SOC) dynamics in  
32 semiarid Mediterranean conditions and to identify which management practices have the  
33 greatest potential to increase SOC in these areas. Five climate scenarios and seven  
34 management scenarios were modeled from 2010 to 2100. Differences in SOC  
35 sequestration were greater among management systems than among climate change  
36 scenarios. Management scenarios under continuous cropping yielded greater C inputs and  
37 SOC gain than scenarios under cereal-fallow rotation. The shift from rainfed conditions  
38 to irrigation also resulted in an increase of C inputs but a decrease in the SOC sequestered  
39 during the 2010-2100 period. The effects of precipitation and temperature change on  
40 SOC dynamics were different depending on the management system applied.  
41 Consequently, the relative response to climate and management depended on the net  
42 result of the influences on C inputs and decomposition. Under climate change, the  
43 adoption of certain management practices in semiarid Mediterranean agroecosystems  
44 could be critical in maximizing SOC sequestration and thus reducing CO<sub>2</sub> concentration  
45 in the atmosphere.

46

47 **Abbreviations**

48 AOGCM, atmosphere-ocean general circulation models

49 BF, barley-fallow

50 CB, continuous barley

51 IRRI, irrigated

52 CT, conventional tillage

53 NT, no-tillage

54 SOC, soil organic carbon

55 SOM, soil organic matter

56 SR, straw removal

57

58 **Introduction**

59 Soil organic carbon (SOC) stocks in Mediterranean semiarid agroecosystems are  
60 constrained by different factors. Limited C input because of low precipitation and high  
61 evapotranspiration rates; centuries of agriculture under intensive tillage systems  
62 combined with the use of long bare fallows (16-18 months between crops) and the  
63 removal of crop residues for animal feed are some of these factors (Austin et al. 1998;  
64 Hernanz et al. 2009). In the Mediterranean region, several long-term experiments were  
65 initiated in the 1980's and 1990's to determine the effects of different management  
66 practices on crop development and soil properties (i.e. Hernanz et al. 1995; López and  
67 Arrúe 1997; Moreno et al. 1997). These experiments were originally established to study  
68 crop growth and soil fertility characteristics under different tillage and cropping systems.  
69 In the past decade, data on management effects on SOC sequestration and dynamics in

70 these agroecosystems have been collected. For example, López-Fando et al. (2007), in  
71 central Spain, reported 13% more SOC in no-tillage (NT) compared to conventional  
72 tillage (CT) in the 0-30 cm depth. Also, Ordóñez-Fernández et al. (2007) in southern  
73 Spain measured 20% more SOC stock under NT than under CT in the top 26 cm soil  
74 depth. Álvaro-Fuentes et al. (2008) in northeast Spain observed up to 15% more SOC in a  
75 continuous barley (CB) system compared to a barley-fallow (BF) system in the 0-30 cm  
76 layer. These studies support the potential for SOC sequestration as a result of the  
77 adoption of alternative management practices in Mediterranean regions. To our  
78 knowledge, there are no region-specific assessments of the potential effects that climate  
79 change could have in SOC dynamics in semiarid Mediterranean agroecosystems.

80 Applications of atmosphere-ocean general circulation models (AOGCM) in the  
81 Mediterranean region suggest that climate change could result in significant warmer  
82 conditions and lower precipitations (Gibelin and Déqué 2003). Results from the  
83 PRUDENCE project using regional climate models in Europe showed that the largest  
84 warming is projected to occur in the Mediterranean region, with an increase of greater  
85 than 6 °C in the Iberian Peninsula during summer (Christensen et al. 2007). Recently, the  
86 effects of warming on SOC dynamics have received particular attention (e.g., Davidson  
87 and Janssen 2006; Conant et al. 2008). Despite some remaining uncertainty about the  
88 effects of warming on different compartments of SOC (Kirschbaum 2006), there is  
89 scientific consensus that soil organic matter (SOM) overall is sensitive to an increase in  
90 temperature (Knorr et al. 2005; Conant et al. 2008). However, temperature is not the only  
91 factor affecting soil microbial activity and SOM turnover but the soil water content also  
92 plays a major role (e.g. Linn and Doran 1984; Skopp et al. 1990). Furthermore, in

93 semiarid regions climate change could have a significant impact not only on soil  
94 microbial processes but also on crop growth and the return of C inputs to the soil  
95 (Mínguez et al. 2007). Previous model analyses suggest that climate impacts on both  
96 SOM turnover and crop growth may be modified by management practices (e.g. Paustian  
97 et al. 1996). For example, Lugato and Berti (2008), in northeast Italy, observed  
98 significant differences in SOC sequestration under climate change depending on the  
99 management applied.

100 Since the soil C sink could take even 100 years to reach a new equilibrium, simulation  
101 models are a valuable tool to study the interactions of climate change, management  
102 practices and ecological processes (Rosenzweig 1990; Ojima et al. 1993; Paustian et al.  
103 1997). In this study we used the Century ecosystem model together with climatic outputs  
104 from different AOGCM's to study the effects of climate change and management on  
105 SOC dynamics in semiarid Mediterranean conditions and to identify what management  
106 practices have the greatest SOC sequestration potential in these areas.

107

## 108 **Materials and methods**

### 109 **Experimental site and model description**

110 An experimental site located in the Zaragoza province NE Spain (41°44'30''N,  
111 0°46'18''W, 270 m) was chosen as broadly representation of conditions in the semiarid  
112 cropland of Spain. The climate is semiarid, with an average annual precipitation of 340  
113 mm and an average annual air temperature of 14.7 °C. The soil is a fine-loamy, mixed,  
114 thermic Xerollic Calciorthid (Soil Survey Staff, 1975) with the following main properties  
115 for the 0-20 cm soil layer: pH (H<sub>2</sub>O, 1:2.5): 8.3; electrical conductivity(1:5): 0.25 dS m<sup>-1</sup>;

116  $\text{CaCO}_3$ :  $432 \text{ g kg}^{-1}$ ; sand (2000-50  $\mu\text{m}$ ), silt (50-2  $\mu\text{m}$ ), and clay (<2  $\mu\text{m}$ ) content: 293,  
117 484 and  $223 \text{ g kg}^{-1}$ , respectively. The long-term experiment was established in 1989 and  
118 consisted of a long-term tillage (three tillage systems: CT, reduced tillage and NT) and  
119 two cropping systems (BF vs. CB) comparison experiment. In the present study, data  
120 was only presented from the CT and BF system, as the baseline historical management  
121 performed in the area during decades. Primary tillage consisted of mouldboard ploughing  
122 to a depth of 30 cm implemented in early spring every two seasons during the fallow  
123 phase of the rotation. Secondary tillage was implemented in late spring with a cultivator  
124 pass to a depth of 15-20 cm. Inorganic nitrogen was applied in all the treatments since  
125 1998. The N fertilization rates have been changed every season and ranged from 26 to 60  
126  $\text{kg N ha}^{-1}$ . From the records found, it is known that prior to the establishment of the long-  
127 term experiments fields had been under CT and BF rotation for several decades.

128 We chose this experiment mainly for two reasons. Firstly, the experiment was located in  
129 a representative Mediterranean semiarid area with typical soil, climate and landscape.  
130 The second is that we previously used the Century model in this same long-term  
131 experiment to validate the Century model in Mediterranean semiarid areas (Álvaro-  
132 Fuentes et al. 2009).

133 The Century model is a general ecosystem model designed to simulate C, N, S and P  
134 dynamics in a monthly time step. The model was described in detail by Parton et al.  
135 (1987, 1994). The parameterization and initialization of the model for the same  
136 experimental plots was done in a previous study (Álvaro-Fuentes et al. 2009). Briefly,  
137 passive and slow SOM pools were initialized simulating a 5000-yr period with a tree-  
138 grass system and a 100-years period with a barley-fallow rotation with intensive tillage,

139 respectively. Furthermore, parameter constants controlling crop growth (e.g. harvest  
140 index (HIMAX), the effect of water deficit on harvest index (HIWSF and HIMONN), the  
141 fraction of N which goes to the grain (EFGRN) or potential aboveground production  
142 (PRDX)) were calibrated to better represent crop growth according with the values  
143 measured during the experimental period. Also, we implemented the procedure proposed  
144 by Metherell et al. (1993) to simulate SOC dynamics in the 0-30 cm soil depth. For this  
145 experiment, both simulated and measured SOC values from the BF-CT treatment during  
146 the 1989-2005 period were taken from Álvaro-Fuentes et al. (2009).

147 The Century model is able to simulate the impacts of increased atmospheric CO<sub>2</sub> on plant  
148 processes. Century considers the following effects as a result of an increase in  
149 atmospheric CO<sub>2</sub>: (1) higher photosynthesis rates, (2) increased water use efficiency due  
150 to reduced stomatal conductance, (3) decrease in plant N concentration, (4) increase in C  
151 allocation to roots (Metherell et al. 1993; Paustian et al. 1996). The effects of CO<sub>2</sub> change  
152 on plant growth can be parameterized for each crop. In our study where we only modeled  
153 barley we used the same parameterization used by Paustian (1996) for a wheat crop. This  
154 parameterization considered a 30% increase in the potential enhanced photosynthesis and  
155 a decrease in transpiration per unit canopy biomass of 23%. These values are in the range  
156 of responses found for C3 crops in Free Air Carbon-dioxide Enrichment (FACE) studies  
157 (Ainsworth and Long 2005).

158

159 Climate and management scenarios

160 In order to evaluate the effects of climate change on SOC dynamics seven management  
161 scenarios and five climate scenarios were built and simulated over a 90 yr period (from

162 2010 to 2100). Management scenarios are summarized in Table 1. The planting and  
163 harvest dates and the N fertilization rates were similar in all the management scenarios.  
164 Barley crop were planted in November and harvested in June and fertilized with 45 kg N  
165 ha<sup>-1</sup> at planting. In the irrigated scenarios (IRRI), 400 mm of water was applied from  
166 March to June in every cropping season. Tillage and cropping system practices in the  
167 management scenarios were the same used during the experimental period (1989-2005).  
168 Climate scenarios included a baseline scenario with neither CO<sub>2</sub> increase nor climate  
169 change. The climate data used for this baseline scenario was obtained from the average  
170 monthly temperature and monthly accumulated precipitation measured during the 1989-  
171 2005 period. The other four scenarios were obtained from two AOGCM simulations  
172 (ECHAM4 and CGCM2) forced by two IPCC emissions scenarios (SRES: A2 and B2)  
173 (Nakicenovic et al. 2000). The A2 and B2 scenarios were equivalent to a CO<sub>2</sub>  
174 concentration at the end of the simulation period of 856 and 621 ppmv. We assumed a  
175 linear CO<sub>2</sub> concentration increase over time. The climate data was produced by the  
176 Meteorology State Agency (Ministry of the Environment and Rural and Marine Environs  
177 of Spain) using a regionalization technique explained in Brunet et al (2008) to better  
178 adjust the climate change scenarios to the conditions of the area studied. In all the four  
179 climate change scenarios, precipitation decreased in the next order compared to the  
180 baseline scenario in the order: CGCM2-A2 > CGCM2-B2 > ECHAM4-A2 > ECHAM4-  
181 B2 (Table 2). Also the mean maximum and minimum air temperature increased in all of  
182 the climate change scenarios in the order: ECHAM4-A2 > ECHAM4-B2 > CGCM2-A2  
183 > CGCM2-B2. The only exception was the maximum air temperature in the CGCM2-B2  
184 scenario which was lower than in the baseline scenario. In addition to the variation in



185 total annual precipitation, the annual distribution pattern was significantly modified in the  
186 climate change scenarios (Fig. 1). Basically, in this semiarid area, the two typical rain  
187 peaks (in fall and spring) decline, in particular the fall peak. In contrast there was a  
188 precipitation increase during the summer period. However, the annual temperature  
189 distribution was not modified by climate change (Fig. 1).

190

## 191 **Results**

192 The average annualized grain yield, C inputs and SOC variation predicted during the  
193 2010-2100 period for the different management scenarios are shown in Table 3. The  
194 lowest grain yield and C inputs were predicted in the BF-CT-SR scenario. Also, this was  
195 the only management scenario with SOC loss (Table 3). Management scenarios under CB  
196 had greater grain yields, C inputs and SOC gain than scenarios under BF rotation. The  
197 shift from rainfed conditions (CB-CT and CB-NT) to irrigation (CB-CT-IRRI and CB-  
198 NT-IRRI) also resulted in an increase of grain yields and C inputs but a decrease in the  
199 SOC sequestered during the 2010-2100 period (Table 3). Also, within the same cropping  
200 system, NT had lower C input compared to CT but greater SOC gain (Table 3).

201 In the BF management scenarios, C inputs were similar among climate scenarios.  
202 However, greater differences were obtained in the SOC variation during the 2010-2100  
203 period (Fig. 2). In general, the model predicted the lowest SOC gain (in the BF-CT-SR  
204 rotation the greatest SOC loss) in the baseline scenario followed by the CGCM2-A2. The  
205 exception was in the BF-CT in which the lowest SOC gain was in the CGCM2-A2  
206 followed by the baseline. The greatest SOC gain was predicted in the ECHAM4 scenarios  
207 in all the three BF management scenarios.

208 In the CB scenarios under rainfed conditions (i.e. CB-CT and CB-NT), differences in C  
209 inputs and SOC gain among climate scenarios were similar. The greatest C inputs and  
210 SOC gain were predicted in the baseline and both CGCM2 scenarios and the lowest in  
211 the ECHAM4-A2 and ECHAM4-B2 (Fig. 3).

212 In the CB scenarios under irrigated conditions (i.e. CB-CT-IRRI and CB-NT-IRRI),  
213 similar C inputs were observed among scenarios (Fig. 4). However, the greatest SOC  
214 gain was obtained in the CGCM2-A2 and CGCM2-B2 scenarios in both management  
215 scenarios (Fig. 4).

216 The temporal SOC dynamics in the CGCM2-B2 and ECHAM4-A2 climate scenarios  
217 during the 2010-2100 period are shown in Fig. 5. Both climate scenarios had similar SOC  
218 dynamics among management scenarios. The BF-CT-SR showed an initial SOC decrease  
219 and a stabilization of the SOC by the end of the study period. The BF-CT management  
220 scenario kept almost steady over the entire study period (Fig. 5). In both climate  
221 scenarios, the CB management showed an almost linear SOC increase during the entire  
222 studied period. However, in the latest 10 years of the simulation, the SOC gain in CB  
223 under ECHAM4-A2 scenario kept almost steady (Fig. 5). However, this trend was not  
224 shown in the CGCM2-B2 scenario.

225

## 226 **Discussion**

227 In semiarid Mediterranean areas, the effects of cropping intensification and tillage  
228 reduction on SOC sequestration under current climate conditions have been widely  
229 studied (i.e. Virto et al. 2007; Álvaro-Fuentes et al. 2008; Hernanz et al. 2009). In the  
230 same study area, we used the Century model to simulate SOC dynamics in a tillage-

231 cropping system long-term experiment under current climate conditions (Álvaro-Fuentes  
232 et al. 2009). In a previous study (Álvaro-Fuentes et al. 2009), we observed threefold  
233 higher SOC sequestration rates under a NT-CB system than under a CT-CB system and  
234 under a NT-BF system. Similarly, in our present study under climate change conditions  
235 both CB and NT also showed greater SOC gain than the scenarios under BF and CT.  
236 However, greater C input was predicted under CT than under NT in both current climate  
237 and climate change conditions. Moret et al. (2007) in the same experimental plots  
238 concluded that the greater biomass production observed under CT than under NT was  
239 explained by higher soil evaporation in NT compared to CT due to lower ground cover  
240 provided by the crop during growth.

241 In the study area selected, climate change scenarios predicted an increase in air  
242 temperature and a reduction in total annual precipitation. As commented previously,  
243 annual precipitation distribution was modified under climate change. In the four climate  
244 change scenarios, precipitation increased during summer compared with the baseline.  
245 Furthermore, reduction in precipitation was predicted during the two water recharge  
246 periods (autumn and spring), which are critical for crop growth. However, in the BF and  
247 IRRRI management scenarios, the average C inputs were reasonably steady among climate  
248 change scenarios. Though in semiarid Mediterranean agroecosystems crop production is  
249 strongly dependant on rainfall (Austin et al. 1998), the drop in precipitation due to  
250 climate change did not have a strong effect on C inputs. Increase in atmospheric CO<sub>2</sub> has  
251 been associated with both the stimulation of crop photosynthesis (i.e. CO<sub>2</sub> fertilization  
252 effect) (Friedlingstein et al. 1995; Lobell and Field 2008) and the increase in water use  
253 efficiency as a result of lower stomatal conductance (Morgan et al. 2004). Consequently,

254 the effect of precipitation reduction on crop growth could be ameliorated by the increase  
255 in water use efficiency and crop photosynthesis due to atmospheric CO<sub>2</sub> increase.  
256 However, in the CB scenarios, there was a slight decrease in C inputs predicted for the  
257 two ECHAM4 scenarios compared to the CGCM2 scenarios. This fact could suggest that  
258 exceptional reductions in precipitation (e.g. more than 60 mm per year in the ECHAM4  
259 scenarios compared to the baseline) could lead to situations in which water use efficiency  
260 improvement by increased atmospheric CO<sub>2</sub> could not completely ameliorate water stress  
261 effect on plant growth.

262 At the same time, climate modification due to increases in atmospheric CO<sub>2</sub> has a  
263 significant impact over SOC turnover (McGuire et al. 1995). A positive relationship  
264 between warming and SOC decomposition has been experimentally demonstrated  
265 (Kirschbaum, 1995; Trumbore et al., 1996; Conant et al., 2008). In our study, the effects  
266 of temperature increase on SOC decomposition was observed in the irrigated (IRRI)  
267 scenarios where despite similar C inputs slightly lower SOC gain was predicted in the  
268 ECHAM4 scenarios compared with the CGCM2 scenarios. Similar soil moisture among  
269 climate scenarios due to irrigation supply together with an increase in more than 3°C in  
270 both minimum and maximum temperatures predicted in the ECHAM4 scenarios  
271 compared with the CGCM2 scenarios resulted in greater SOC decomposition. However  
272 opposite behavior was observed in the barley-fallow (BF) scenarios where lower SOC  
273 decomposition was predicted for the two ECHAM4 scenarios despite the similar C inputs  
274 among climate scenarios. Soil moisture limits SOC decomposition (Stott et al. 1986),  
275 particularly in semiarid conditions (Wildung et al. 1975; Paustian et al. 1996).  
276 Furthermore, in semiarid southeastern Spain but under current climate conditions,

277 Almagro et al. (2009) established a threshold value of soil water content above which soil  
278 respiration was controlled basically by soil temperature and below which was controlled  
279 by precipitation only. In our study, it is likely that the low soil moisture in the ECHAM4  
280 scenarios in the BF system led to limited conditions for soil microorganism to decompose  
281 SOC.

282 The increase in precipitation predicted during summer months in the four climate change  
283 scenarios led to increases on SOC decomposition compared to the baseline scenario.  
284 Despite this increased precipitation during the summer months (July and August), the  
285 model predicted a decline in SOC in all the climate and management scenarios; under  
286 climate change SOC losses were greater compared to the baseline scenario. For instance,  
287 in the CB-NT and CB-CT management scenarios under the ECHAM4-A2 climate change  
288 scenario, the average SOC loss during July and August was 9.9 and 7.4 gC m<sup>-2</sup>,  
289 respectively. However, for the same period and management scenarios, under the  
290 baseline scenario SOC losses were 6.9 and 4.5 gC m<sup>-2</sup>, respectively. The BF management  
291 scenarios showed a similar SOC loss between the climate change scenarios and the  
292 baseline scenarios. In the irrigated (IRRI) scenarios, during summer differences in SOC  
293 loss between climate change scenarios and the baseline scenario were negligible.  
294 Consequently, in the climate change scenarios higher precipitation and elevated soil  
295 temperatures during summer increased SOC decomposition compared to the baseline  
296 scenario. The increase on SOC decomposition can be assessed with the Century output  
297 factor *defac* (i.e. decomposition factor based on the temperature and the soil moisture).  
298 During July and August, the Century model predicted between 40% and 50% increases in  
299 the *defac* parameter of the CB and BF management scenarios under climate change

300 scenarios compared to the same management scenarios under baseline conditions (data  
301 not shown).

302 As commented before, the CB management scenarios showed different C inputs among  
303 climate scenarios. The lower C inputs predicted in the ECHAM4 scenarios led to lower  
304 SOC gain under these climate scenarios compared to the CGCM2 scenarios.  
305 Consequently, under CB, climate effects on SOC dynamics were primarily due to an  
306 increase in C inputs. Basically, the relative response to climate and management in this  
307 study depended on the net result of the influences on C inputs and decomposition.  
308 Consequently, different management systems showed different responses to climate  
309 change scenarios.

310 Following a change of management practice, SOC content tends to reach a new steady  
311 state (West and Six 2007). The time it takes to achieve this new steady state varies  
312 between ecosystems, climate regimes and land management. In our study, different  
313 climate scenarios had different effects on the duration of SOC sequestration. The  
314 ECHAM4-A2 climate scenario with higher increase in temperature and lower annual  
315 precipitation achieved a steady state earlier than the CGCM2-B2 scenario. West and Six  
316 (2007) observed that sequestration activities with greater impacts on decomposition rates  
317 result in lower sequestration durations.

318 Results from simulation models are associated with imprecision and bias known as model  
319 uncertainty (Ogle et al. 2007). Lugato and Berti (2008), in a similar study in northeast  
320 Italy, identified three main sources of uncertainty: associated with the model, associated  
321 with the climate scenarios and associated with the management scenarios. These same  
322 authors pointed out that the uncertainty associated with the model is basically related

323 with the fact that Century is a model based on a first-order decomposition kinetics,  
324 resulting in a SOC increase without limits as C input increases. However, as suggested by  
325 Stewart et al. (2007), some long-term experiments showed no change in SOC content in  
326 response to different C input levels. Therefore, in our experiment, the almost linear SOC  
327 gain in the four CB scenarios with significant increase in SOC stock in the 90 years of  
328 simulation (Fig. 5) could be overestimated. Another possible uncertainty source could be  
329 the climate scenarios. The use of AOGCM could result in significant biases in the  
330 precipitation and temperature predicted when used for regional studies, particularly in  
331 areas of complex topography and land use distribution like the Mediterranean basin  
332 (Christensen et al. 2007). However, in our study, as commented in the Methods section,  
333 the climate data used has been transformed with a regionalization technique (Brunet et al.  
334 2008) by the Spanish Meteorology State Agency in order to better adjust the climate  
335 change scenarios to the physiographic conditions of the area studied. The third source of  
336 uncertainty was associated with the management scenarios. As suggested Lugato and  
337 Berti (2008), farmers would react to climate change in a dynamic way by adopting new  
338 management practices or by using new genetic material. In our simulation, management  
339 scenarios remained the same during the whole simulation period. This means that  
340 possible changes in fertilization, crop varieties or species, irrigation or event planting and  
341 harvest dates, were not contemplated in the study. However, we considered that  
342 maintaining the same management during the simulation period could help to better  
343 understand the interactions between climate and management, which was the purpose of  
344 this experiment. Another source of uncertainty could be related to the adjustment  
345 between measured and simulated baseline conditions. As discussed in a previous study

346 (Ávaro-Fuentes et al. 2009), during the experimental period (1989-2005) the Century  
347 model somewhat overestimated SOC gain in the BF-CT system. This fact could result in  
348 a slight bias of the final SOC value for all the management scenarios. However, this bias  
349 did not necessarily have an impact on the differences found in SOC among management  
350 and climate scenarios.

351 Because this study was only focused on one site, it is not necessarily representative of the  
352 entire region. Especially recently when new simulation tools and approaches are being  
353 created in order to simulate SOC stocks in regional/national scales (e.g. Milne et al. 2007;  
354 Tornquist et al. 2009). However, the main purpose of our study was to evaluate the  
355 effects of management on SOC dynamics under climate change in semiarid  
356 Mediterranean conditions. Therefore, an approach based on one specific site that was  
357 representative of the climate, historic management and soil characteristics could give us a  
358 better interpretation of the interaction between management and climate on SOC  
359 dynamics.

360

## 361 **Conclusions**

362 Our study investigated the role of management practices on SOC dynamics under climate  
363 change in a semiarid Mediterranean agroecosystem. The adoption of certain management  
364 practices could be essential in order to maximize SOC sequestration under climate  
365 change. In our study, differences on SOC sequestration were greater among management  
366 systems than among climate change scenarios. In general, cropping intensification and  
367 NT had greater SOC sequestration than cereal-fallow and CT, respectively. At the same  
368 time, rainfed systems compared to irrigated systems resulted in greater SOC gain and the



369 removal of the straw resulted in SOC loss over time. The effect of precipitation and  
370 temperature change on SOC dynamics was different depending on the management  
371 system applied. Consequently, the relative response to climate and management  
372 depended on the net result of the influences on C inputs and decomposition. Under  
373 climate change, the adoption of certain management practices in semiarid Mediterranean  
374 agroecosystems could be critical in maximizing SOC sequestration and thus reducing  
375 CO<sub>2</sub> concentration in the atmosphere.

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## REFERENCES

392

393 Ainsworth EA, Long SP (2005) What have we learned from 15 years of free-air CO<sub>2</sub>  
394 enrichment (FACE)? A meta-analytic review of the responses of photosynthesis,  
395 canopy properties and plant production to rising CO<sub>2</sub>. *New Phytol* 165:351-371.

396 Almagro M, Lopez J, Querejeta JI, Martinez-Mena M (2009) Temperature dependence of  
397 soil CO<sub>2</sub> efflux is strongly modulated by seasonal patterns of moisture availability  
398 in a Mediterranean ecosystem. *Soil Biol Biochem* 41:594-605

399 Álvaro-Fuentes J, López MV, Arrúe JL, Moret D, Paustian K (2009) Tillage and  
400 cropping effects on soil organic carbon in Mediterranean semiarid agroecosystems:  
401 testing the Century model. *Agr Ecosyst Environ* 134:211-217

402 Álvaro-Fuentes J, López MV, Cantero-Martínez C, Arrúe JL (2008) Tillage effects on  
403 soil organic carbon fractions in Mediterranean dryland agroecosystems. *Soil Sci Soc  
404 Am J* 72:541-547

405 Austin RB, Cantero-Martínez C, Arrúe JL, Playán E, Cano-Marcellán P (1998) Yield-  
406 rainfall relationships in cereal cropping systems in the Ebro river valley of Spain. *Eur  
407 J Agron* 8:239-248

408 Brunet M, Casado MJ, de Castro M, Galán P, López JA, Martín JM, Pastor A, Petisco E,  
409 Ramos P, Ribalaygua J, Rodríguez E, Sanz I, Torres L (2008) Generación de  
410 escenarios regionalizados de cambio climático para España. Agencia Estatal de  
411 Meteorología. Ministerio de Medio Ambiente, Medio Rural y Marino. Gobierno de  
412 España

413 Christensen JH, Carter TR, Rummukainen M, Amanatidis G (2007) Evaluating the  
414 performance and utility of regional climate models: the PRUDENCE project.  
415 Climatic Change 81:1-6

416 Conant R, Drijber RA, Haddix ML, Parton W, Paul EA, Plante A, Six J, Steinweg JM  
417 (2008) Sensitivity of organic matter decomposition to warming varies with its quality.  
418 Global Change Biol 14:868-877

419 Davidson EA, Janssens IA (2006) Temperature sensitivity of soil carbon decomposition  
420 and feedbacks to climate change. Nature 440:165-173

421 Friedlingstein P, Fung I, Holland E, John J, Brasseur G, Erickson D, Schimel D (1995)  
422 On the contribution of CO<sub>2</sub> fertilization to the missing biospheric sink. Global  
423 Biogeochem Cycles 9:541-556

424 Gibelin AL, Déqué M (2003) Anthropogenic climate change over the Mediterranean  
425 region simulated by a global variable resolution model. Clim Dynam 20:327-339

426 Hernanz JL, Girón V, Cerisola C (1995) Long-term energy use and economic evaluation  
427 of three tillage systems for cereal and legume production in central Spain. Soil Till  
428 Res 35:183-198

429 Hernanz JL, Sanchez-Giron V, Navarrete L (2009) Soil carbon sequestration and  
430 stratification in a cereal/leguminous crop rotation with three tillage systems in  
431 semiarid conditions. Agr Ecosyst Environ 133:114-122

432 Kirschbaum MUF (1995) The temperature dependence of soil organic matter  
433 decomposition and the effect of global warming on soil organic carbon storage. Soil  
434 Biol Biochem 27:753-760

435 Kirschbaum MUF (2006) The temperature dependence of organic-matter decomposition -  
436 still a topic of debate. *Soil Biol Biochem* 38:2510-2518

437 Knorr W, Prentice IC, House JI, Holland EA (2005) Long-term sensitivity of soil carbon  
438 turnover to warming. *Nature*, 433:298-301

439 Linn DM, Doran JW (1984) Effect of water-filled pore-space on carbon-dioxide and  
440 nitrous-oxide production in tilled and nontilled soils. *Soil Sci Soc Am J* 48:1267-1272

441 Lobell D, Field C (2008) Estimation of the carbon dioxide (CO<sub>2</sub>) fertilization effect using  
442 growth rate anomalies of CO<sub>2</sub> and crop yields since 1961. *Global Change Biol* 14:39-  
443 45

444 López, MV, Arrúe, JL (1997) Growth, yield and water use efficiency of winter barley in  
445 response to conservation tillage in a semi-arid region of Spain. *Soil Till Res* 44: 35-54

446 López-Fando C, Dorado J, Pardo MT (2007) Effects of zone-tillage in rotation with no-  
447 tillage on soil properties and crop yields in a semi-arid soil from central. Spain *Soil*  
448 *Till Res* 95:266-276

449 Lugato E, Berti A (2008) Potential carbon sequestration in a cultivated soil under  
450 different climate change scenarios: A modelling approach for evaluating promising  
451 management practices in north-east Italy. *Agr Ecosyst Environ* 128:97-103

452 McGuire AD, Melillo JM, Kicklighter DW, Joyce LA (1995) Equilibrium responses of  
453 soil carbon to climate change: Empirical and process-based estimates. *J Biogeogr*  
454 22:785-796

455 Metherell AK, Harding LA, Cole CV, Parton WJ (1993) CENTURY Soil Organic Matter  
456 Model Environment Technical Documentation. Agroecosystem Version 4.0 Great

457 Plains System Research unit Technical Report No. 4. USDA-ARS, Fort Collins,  
458 Colorado. 245 p.

459 Milne E, Al Adamat R, Batjes NH, Bernoux M, Bhattacharyya T, Cerri C, Cerri C,  
460 Coleman K, Easter M, Falloon P, Feller C, Gicheru P, Kamoni P, Killian K, Pal DK,  
461 Paustian K, Powlson DS, Rawajfih Z, Sessay M, Williams S, Wokabi S (2007)  
462 National and sub-national assessments of soil organic carbon stocks and changes: The  
463 GEFSOC modelling system. *Agr Ecosyst Environ* 122:3-12

464 Mínguez M, Ruiz-Ramos M, Díaz-Ambrona C, Quemada M, Sau F (2007) First-order  
465 impacts on winter and summer crops assessed with various high-resolution climate  
466 models in the Iberian Peninsula. *Clim Change* 81:343-355

467 Moreno F, Pelegrín F, Fernández JE, Murillo JM (1997) Soil physical properties, water  
468 depletion and crop development under traditional and conservation tillage in  
469 southern Spain. *Soil Till Res* 41:25-42

470 Morgan JA, Pataki DE, Korner C, Clark H, Del Grosso SJ, Grunzweig JM, Knapp AK,  
471 Mosier AR, Newton PCD, Niklaus PA, Nippert JB, Nowak RS, Parton WJ, Polley  
472 HW, Shaw MR (2004) Water relations in grassland and desert ecosystems exposed to  
473 elevated atmospheric CO<sub>2</sub>. *Oecologia* 140:11-25

474 Nakicenovic N, Alcamo J, Davis G, de Vries B, Fenhann J, Gaffin S, Gregory K, Grubler  
475 A, Jung TY, Kram T, La Rovere EL, Michaelis L, Mori S, Morita T, Pepper W,  
476 Pitcher H, Price L, Riahi K, Roehrl A, Rogner H-H, Sankovski A, Schlesinger M,  
477 Shukla P, Smith S, Swart R, van Rooijen S, Victor N, Dadi Z (2000) Special Report  
478 on Emissions Scenarios: A Special Report of Working Group III of the

479 Intergovernmental Panel on Climate Change. Cambridge University Press,  
480 Cambridge, UK.

481 Ogle SM, Breidt FJ, Easter M, Williams S, Paustian K (2007) An empirically based  
482 approach for estimating uncertainty associated with modelling carbon sequestration in  
483 soils. *Ecol Model* 205:453-463

484 Ojima DS, Parton WJ, Schimel DS, Scurlock JMO, Kittel TGF (1993) Modeling the  
485 effects of climatic and CO<sub>2</sub> changes on grassland storage of soil C. *Water Air Soil*  
486 *Poll* 70:643-657

487 Ordóñez Fernández R, González Fernández P, Giráldez Cervera JV, Perea Torres F  
488 (2007) Soil properties and crop yields after 21 years of direct drilling trials in  
489 Southern Spain. *Soil Till Res* 94:47-54

490 Parton WJ, Schimel DS, Cole CV, Ojima DS (1987) Analysis of factors controlling soil  
491 organic matter levels in Great Plains grasslands. *Soil Sci Soc Am J* 51:1173-1179

492 Parton WJ, Schimel DS, Ojima DS, Cole CV (1994) A general model for soil organic  
493 matter dynamics: sensitivity to litter chemistry, texture and management. In: Bryant  
494 RB, Arnold RW (Eds.), *Quantitative Modeling of Soil Forming Processes*, SSSA  
495 Spec. Pub. 39. ASA, CSSA and SSSA, Madison, WI, p. 147-167

496 Paustian K, Elliott ET, Peterson GA, Killian K (1996) Modelling climate, CO<sub>2</sub> and  
497 management impacts on soil carbon in semi-arid agroecosystems. *Plant Soil* 187:351-  
498 365

499 Paustian K, Levine E, Post WM, Ryzhova IM (1997) The use of models to integrate  
500 information and understanding of soil C at the regional scale. *Geoderma* 79:227-260

501 Rosenzweig C (1990) Crop response to climate change in the southern Great Plains: A  
502 simulation study. *Prof Geogr* 42:20-37

503 Skopp J, Jawson MD, Doran JW (1990) Steady-state aerobic microbial activity as a  
504 function of soil-water content. *Soil Sci Soc Am J* 54:1619-1625

505 Soil Survey Staff. 1975. Soil taxonomy, a basic system of soil classification for making  
506 and interpreting soil surveys. USDA-SCS Agric. Handbook 436. US Gov. Print.  
507 Office, Washington, DC, USA.

508 Stewart C, Paustian K, Conant R, Plante A, Six J (2007) Soil carbon saturation: concept,  
509 evidence and evaluation. *Biogeochemistry* 86:19-31

510 Stott DE, Elliott LF, Papendick RI, Campbell GS (1986) Slow-temperature or low water  
511 potential effects on the microbial decomposition of wheat residue. *Soil Biol Biochem*  
512 18: 577-582

513 Tornquist CG, Giasson E, Mielniczuk J, Pellegrino Cerri CE, Bernoux M (2009) Soil  
514 organic carbon stocks of Rio Grande do Sul, Brazil. *Soil Sci Soc Am J* 73:975-982

515 Trumbore SE, Chadwick OA, Amundson R (1996) Rapid exchange between soil carbon  
516 and atmospheric carbon dioxide driven by temperature change. *Science* 272:393-  
517 396

518 Virto I, Imaz MJ, Enrique A, Hoogmoed W, Bescansa P (2007) Burning crop residues  
519 under no-till in semi-arid land, Northern Spain-effects on soil organic matter,  
520 aggregation, and earthworm populations. *Aust J Soil Res* 45:414-421

521 West T, Six J (2007) Considering the influence of sequestration duration and carbon  
522 saturation on estimates of soil carbon capacity. *Clim Change* 80:25-41

523 Wildung RE, Garland TR, Buschbom RL (1975) Interdependent effects of soil  
524 temperature and water-content on soil respiration rate and plant root decomposition  
525 in arid grassland soils. Soil Biol Biochem 7:373-378  
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546 **Tables**

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548 Table 1. Summary of the management scenarios used for simulation during the 2010-  
549 2100 period.

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Management scenarios	Description
BF-CT-SR	Barley-fallow system under conventional tillage and straw removal
BF-CT	Barley-fallow system under conventional tillage and straw incorporated into the soil
BF-NT	Barley-fallow system under no-tillage and straw left on soil surface
CB-CT	Continuous-barley system under conventional tillage and straw incorporated into the soil
CB-NT	Continuous-barley system under no-tillage and straw left on soil surface
CB-CT-IRRI	Continuous-barley system under conventional tillage and straw incorporated into the soil with irrigation
CB-NT-IRRI	Continuous-barley system under no-tillage and straw left on soil surface with irrigation

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563 Table 2. Average total annual precipitation, maximum and minimum air temperature  
564 predicted by the different climate scenarios during the 2010-2100 period.

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	Precipitation (mm yr <sup>-1</sup> )	Tmax (°C)	Tmin (°C)
Baseline	340	21.5	8.4
CGCM2-B2	300	20.9	8.9
CGCM2-A2	302	21.7	9.5
ECHAM4-B2	280	25.7	12.6
ECHAM4-A2	277	26.3	13.1

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582 Table 3. Annualized grain yield, C input and SOC variation averaged across all climate  
583 scenarios during the 2010 to 2100 period for the management scenarios.

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Management scenario	Grain yield (kg ha <sup>-1</sup> )	C inputs (g C m <sup>-2</sup> )	ΔSOC (g C m <sup>-2</sup> )	ΔSOC (%)
BF-CT-SR <sup>b</sup>	1458±257	54±4 <sup>a</sup>	-459±64	-13
BF-CT	1778±307	215±4	437±148	13
BF-NT	1467±265	177±2	1485±166	42
CB-CT	1920±437	283±27	2749±209	79
CB-NT	1751±424	252±28	3827±337	109
CB-CT-IRRI	4000±193	356±13	2300±189	66
CB-NT-IRRI	3702±173	329±11	3179±243	91

585 <sup>a</sup> Average value ± standard deviation

586 <sup>b</sup> BF, barley-fallow rotation; CB, continuous barley system; CT, conventional tillage; IRRI, irrigation; NT,  
587 no-tillage; SR, straw removal

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## Figure captions

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603 Fig.1. Precipitation, mean air maximum and minimum temperature distribution for the  
604 different climate scenarios (Baseline, CGCM2-A2, CGCM2-B2, ECHAM4-A2,  
605 ECHAM4-B2).

606 Fig. 2. Average C inputs and SOC change during the 2010-2100 period for the different  
607 climate scenarios (Baseline, CGCM2-A2, CGCM2-B2, ECHAM4-A2, ECHAM4-B2)  
608 and for the barley-fallow management scenarios (BF-CT-SR, Barley-fallow system under  
609 conventional tillage and straw removal; BF-CT, Barley-fallow system under conventional  
610 tillage and straw incorporated into the soil; BF-NT, Barley-fallow system under no-  
611 tillage and straw left on soil surface).

612 Fig. 3. Average C inputs and SOC change during the 2010-2100 period for the different  
613 climate scenarios (Baseline, CGCM2-A2, CGCM2-B2, ECHAM4-A2, ECHAM4-B2)  
614 and for the continuous barley under rainfed conditions scenarios (CB-CT, Continuous-  
615 barley system under conventional tillage and straw incorporated into the soil; CB-NT,  
616 Continuous-barley system under no-tillage and straw left on soil surface).

617 Fig. 4. Average C inputs and SOC change during the 2010-2100 period for the different  
618 climate scenarios (Baseline, CGCM2-A2, CGCM2-B2, ECHAM4-A2, ECHAM4-B2)  
619 and for the continuous barley under irrigated conditions scenarios (CB-CT-IRRI,  
620 Continuous-barley system under conventional tillage and straw incorporated into the soil  
621 with irrigation; CB-NT-IRRI, Continuous-barley system under no-tillage and straw left  
622 on soil surface with irrigation).

623 Fig. 5. Temporal SOC dynamics from 2010 to 2100 for the different management  
624 scenarios (BF-CT-SR, Barley-fallow system under conventional tillage and straw  
625 removal; BF-CT, Barley-fallow system under conventional tillage and straw incorporated  
626 into the soil; BF-NT, Barley-fallow system under no- tillage and straw left on soil  
627 surface; CB-CT, Continuous-barley system under conventional tillage and straw  
628 incorporated into the soil; CB-NT, Continuous-barley system under no-tillage and straw  
629 left on soil surface; CB-CT-IRRI, Continuous-barley system under conventional tillage  
630 and straw incorporated into the soil with irrigation; CB-NT-IRRI, Continuous-barley  
631 system under no-tillage and straw left on soil surface with irrigation) and for the  
632 CGCM2-B2 and ECHAM4-A2 climate scenarios.

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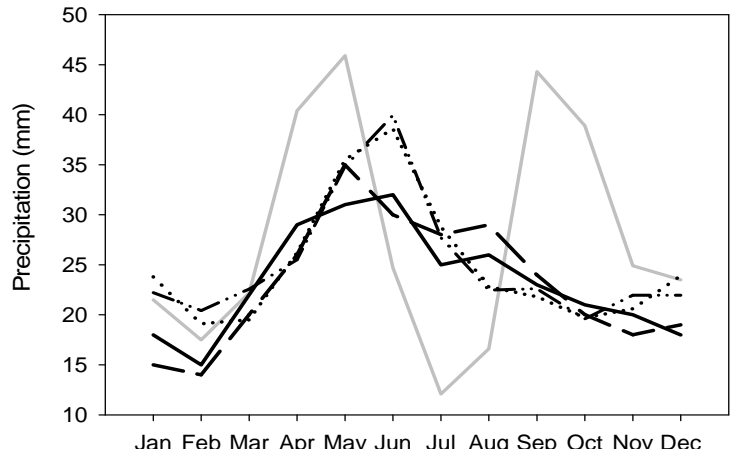
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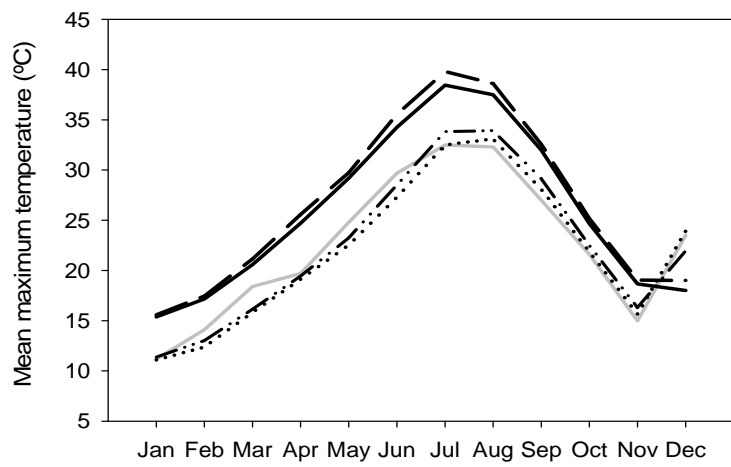
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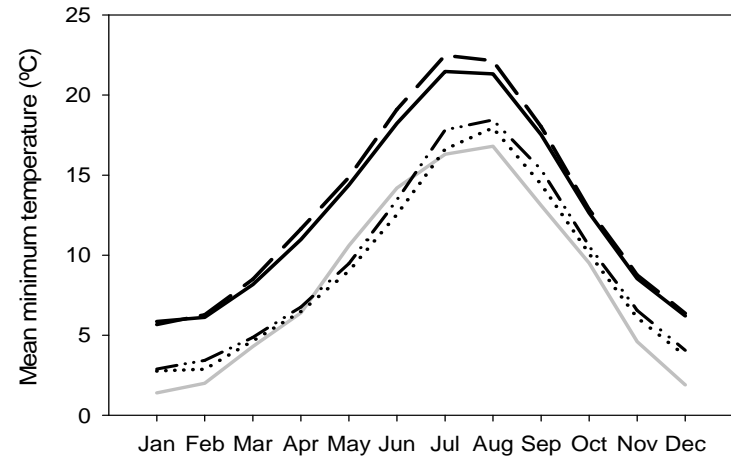
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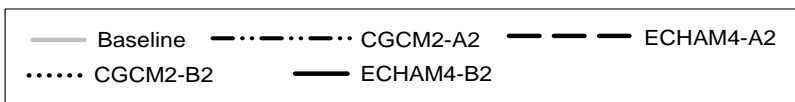
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668 Fig. 1

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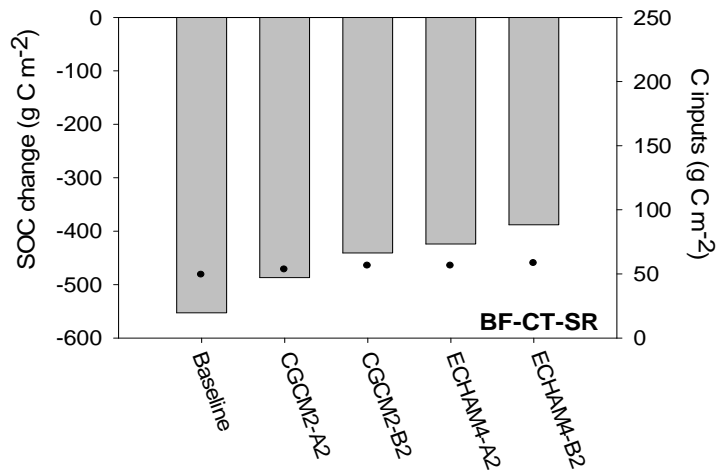
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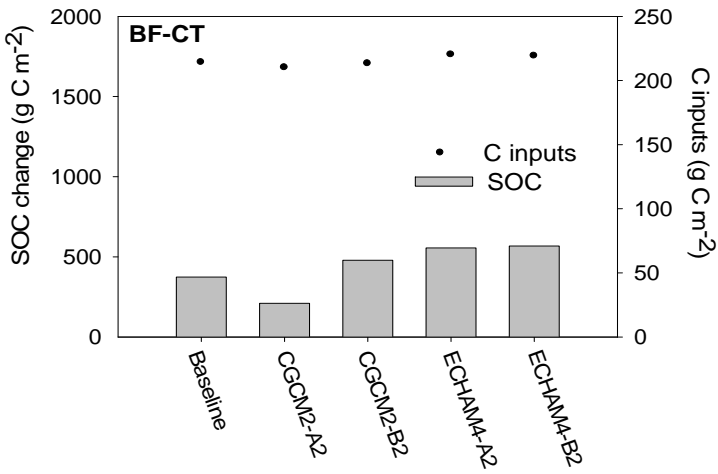
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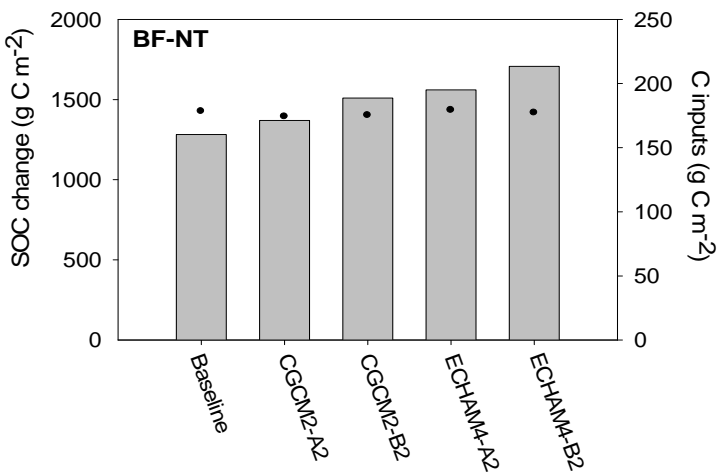
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691 Fig. 2

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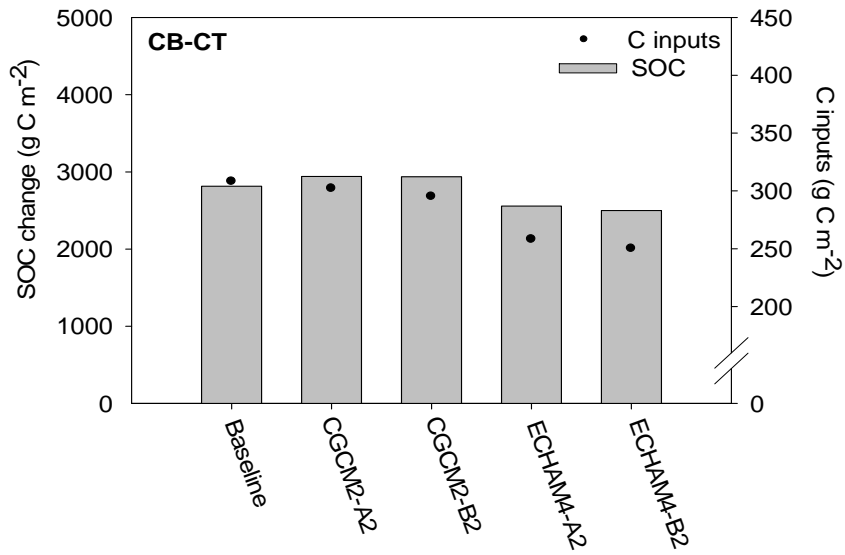
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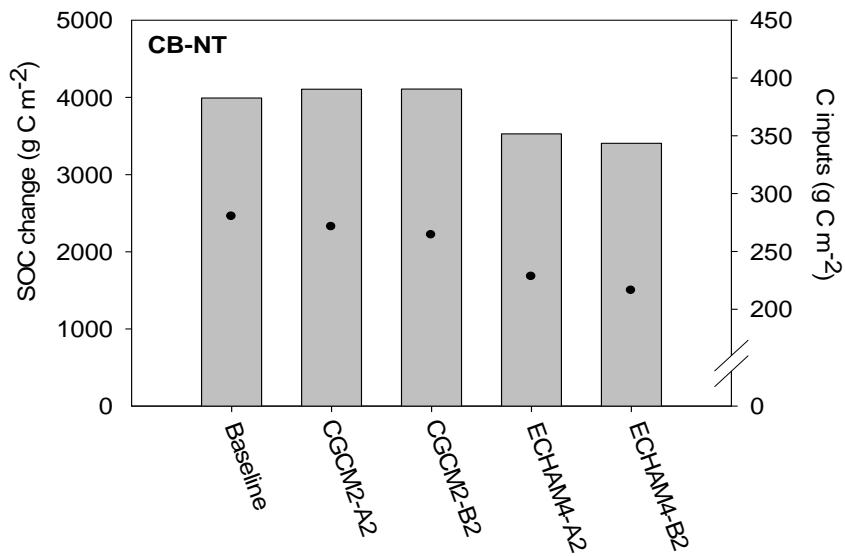
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711 Fig. 3

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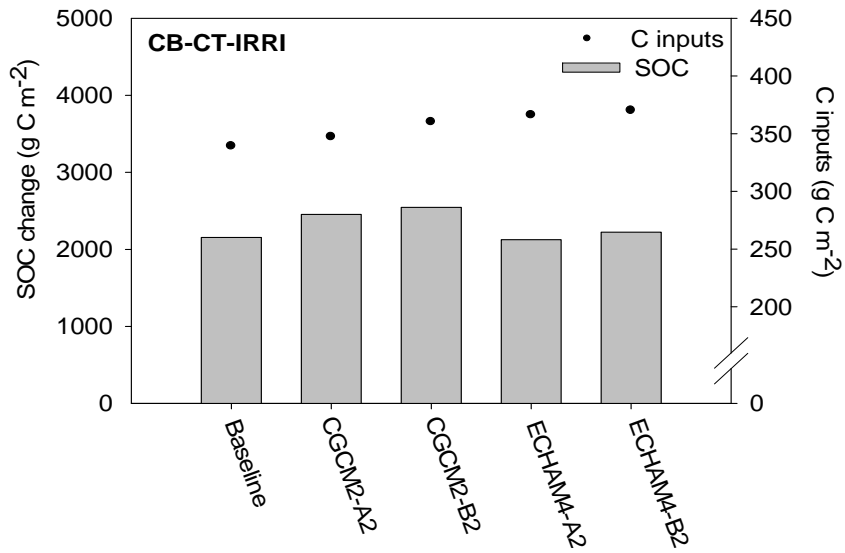
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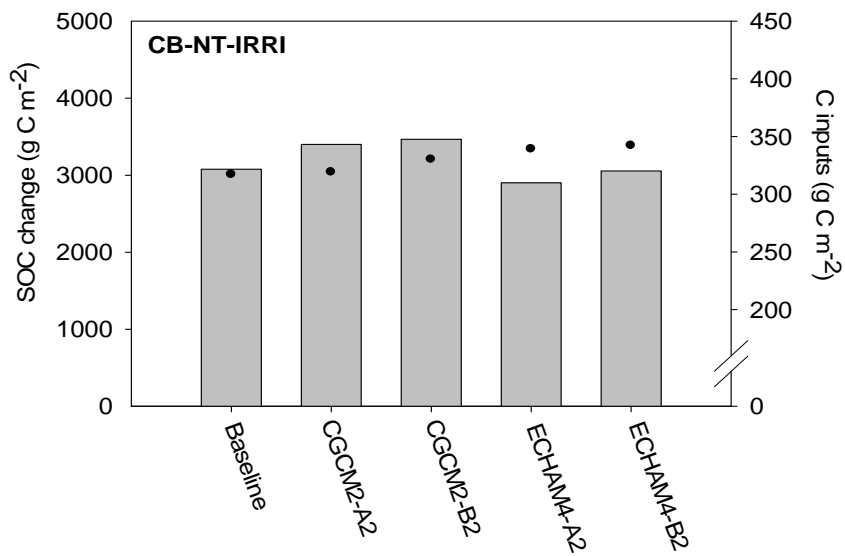
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735 Fig. 4

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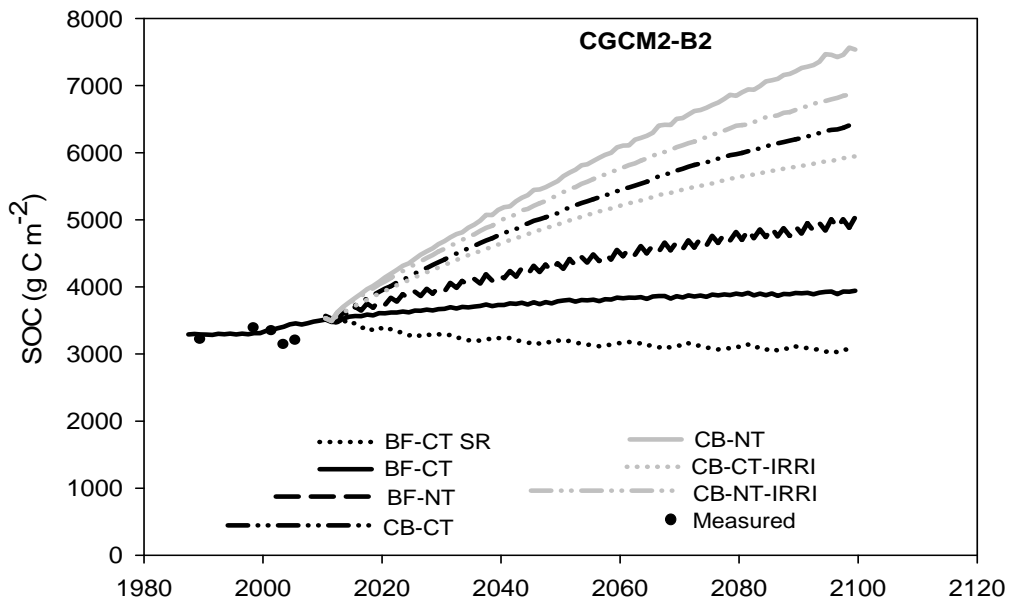
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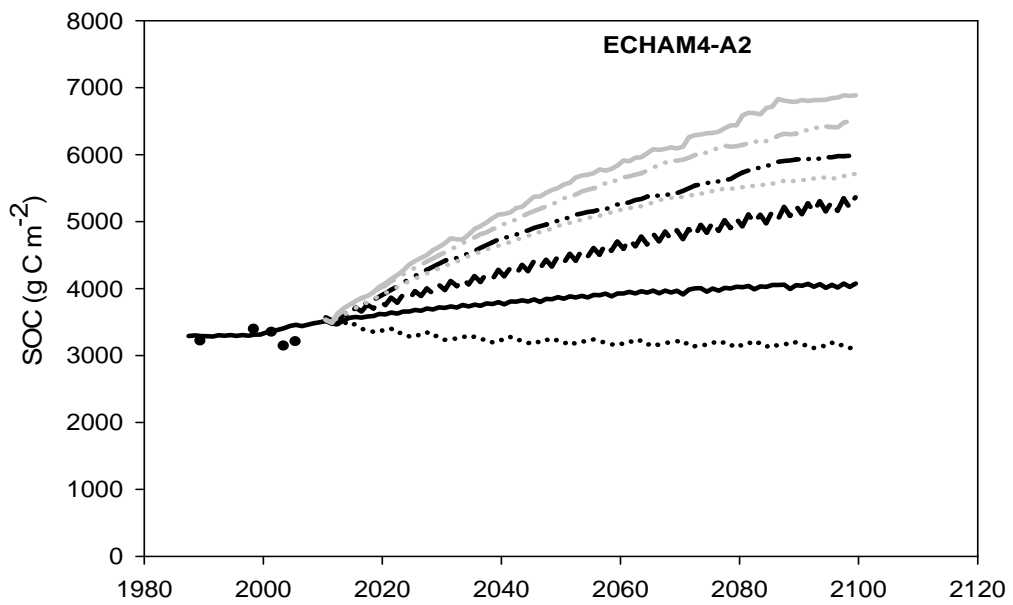
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758 Fig. 5

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