

1 **Spatiotemporal analyses of tracking data reveal fine-scale, daily cycles in seabird-**
2 **fisheries interactions**

3 Jazel Ouled-Cheikh^{1*}, Carola Sanpera^{1,2}, Juan Bécares^{3,5}, José Manuel Arcos^{1,3}, J.L.
4 Carrasco⁴, Francisco Ramírez^{1,2}

5 **1** Departament de Biologia Evolutiva, Ecologia i Ciències Ambientals, Universitat de
6 Barcelona, Barcelona, Spain

7 **2** Institut de Recerca de la Biodiversitat (IRBio)

8 **3** Programa marino, Delegació de Catalunya, SEO/BirdLife, Barcelona, Spain

9 **4** Departament de Fonaments Clínics, Bioestadística, Facultat de Medicina, Universitat
10 de Barcelona, Barcelona, Spain

11

12 **5** Cory's – Investigación y Conservación de la Biodiversidad (current address)

13

14

15 * Corresponding author: jazelouled@gmail.com

16 **ABSTRACT**

17 Human fisheries provide scavengers with abundant and predictable feeding opportunities
18 that may schedule their behavioural patterns. Yet, quantitative assessments on scavenger-
19 fisheries interactions are largely oriented towards assessing their spatial overlap. Using
20 GPS tracking technology, we evaluated how the Audouin's gull, a Mediterranean
21 endemic seabird that makes extensive use of feeding opportunities provided by fisheries,
22 co-occurs (i.e. presumably interacts) with the main fishing gear (i.e., diurnal trawlers and
23 nocturnal purse seiners), both in space and time. Results showed that some individuals
24 were able to adapt their distribution and activity patterns to the scheduled routines of
25 human fisheries. Waveform analyses based on co-occurring positions revealed that most
26 interactions with trawlers occurred during the afternoon (around 16:00 h) when
27 discarding occurs as vessels approach the ports. In contrast, gull-purse seiner interactions
28 largely occurred at night (between 2:00 h and 4:00 h) coinciding with the hauling of nets.

29 Moreover, we found an individual component in seabird-fishery interactions, showing
30 that there may be differential use of fisheries by individuals within the population. In
31 addition to implications for our understanding of the behavioural ecology of these species,
32 these results may have important management implications as this food subsidy becomes
33 increasingly restricted (e.g., EU Common Fisheries Policy).

34 **Keywords:** Discards, gulls, fisheries, interaction, Mediterranean

35

36 INTRODUCTION

37 Among the Earth's biomes, the marine environment is likely one of the most greatly
38 impacted by humans (Halpern *et al.* 2015; Ramirez *et al.* 2017). In addition to human-
39 driven climate impacts, pollution and habitat degradation, marine resource
40 overexploitation can also result in severe changes to marine ecosystems and biodiversity
41 (Cury *et al.*, 2011). For instance, fishing activities have resulted in the complete
42 exploitation of 60% of fish stocks worldwide whereas 33% are overexploited and 7% are
43 depleted (FAO, 2018). Changes in fish abundances caused by fisheries may have further
44 implications within the marine food webs (Pauly, 1998; Essington *et al.*, 2006) with
45 ultimate, often exacerbated, impacts on top predators through bottom-up trophic cascades
46 (Frederiksen *et al.*, 2006; Lynam *et al.*, 2017). Concurrently, human fisheries may also
47 impact these marine predators through direct mortality (i.e., bycatch, Lewison *et al.*,
48 2014), food depletion (through marine resource overexploitation), or by providing
49 resources that would not be naturally available otherwise (Hudson and Furness, 1988).
50 These new feeding opportunities are largely driven by fisheries' discards, which refers to
51 the part of the catch returned to the sea, often dead (Damalas, 2015).

52 A large amount of discards are generated daily by industrial and artisanal fisheries and
53 thrown back into the sea. Historically, global estimated discards increased from under 5
54 million t/year (t = 1,000 kg) in the early 1950s to a peak of 18.8 million t in 1989, and
55 gradually declined thereafter to the levels of the late 1950s of less than 10 million t/year.
56 Globally, this represents between 10% and 20% of the global worldwide catch (Zeller *et*
57 *al.*, 2018). Discarding occurs at specific times and locations, thus resulting in one of the
58 most important and predictable anthropogenic food subsidies in the marine ecosystems
59 worldwide (Oro *et al.*, 2013). Many species take advantage of this food subsidy, and have
60 adapted their distribution and activity patterns to the scheduled routines of human
61 fisheries (Oro *et al.*, 2013). This is the case for some seabirds, whose foraging behaviour,
62 habitat use, and movement patterns are highly affected by the presence/absence of fishing
63 activity and thus, of discards (Bodey *et al.*, 2014, Tyson *et al.*, 2015, Bartumeus *et al.*,
64 2010). This can have an influence on species habits with ultimate consequences on life
65 history traits, population dynamics and community structure (Oro *et al.*, 1999; Bearhop
66 *et al.*, 2008). In this scenario, opportunistic species with high adaptability can take
67 advantage of this resource (Oro *et al.*, 2013) since discards can lead to highly competitive
68 feeding interactions (Arcos *et al.* 2001; Calado *et al.*, 2018). The favoured species are
69 those that make up the communities of scavengers that feed on discards. These
70 communities can vary greatly across different geographic locations in terms of species
71 (Weimerskirch *et al.*, 2000, Tyson *et al.* 2015, Louzao *et al.*, 2011). In the Western
72 Mediterranean, these communities are typically dominated by yellow-legged gulls (*Larus*
73 *michahellis*), Balearic shearwaters (*Puffinus mauretanicus*), the Audouin's gull
74 (*Ichthyaetus audouinii*) (Louzao *et al.*, 2011) and the Cory's shearwater (*Calonectris*
75 *diomedea*) (Abelló *et al.*, 2003).

76 Changes in the dynamics of discard availability may have some direct implications on
77 these scavenger communities. For instance, a decline in the availability of discards can
78 cause food shortages and a subsequent search for locations with better food availability
79 (Calado *et al.*, 2018). The absence of discards has also been tied to greater seabird
80 attendance at longliners and a consequent increase in seabird bycatch (Laneri *et al.* 2010;
81 Soriano-Redondo *et al.* 2016). Other costs can arise even at the level of biotic interactions.
82 For example, some species may offset food shortages and increased competition for
83 natural resources by shifting their trophic regime from discards to other small seabirds,
84 causing potential negative impacts to their populations (Bicknell *et al.*, 2013).

85 Such a shortage in discards may occur in the near future in the EU's Exclusive
86 Economic Zones (EEZ). A discards ban policy (the so-called landing obligation) is being
87 implemented under the current European Union (EU) Common Fisheries Policy (Borges,
88 2015). EU marine scavenger communities may thus be subjected to some of the impacts
89 discussed above (Bicknell *et al.*, 2013). Within this scenario, reliable assessments on
90 seabird-fishery interactions are key to taking proper management decisions and to
91 providing answers on how communities will respond when discards are no longer
92 available (Oro *et al.*, 2013).

93 The Audouin's gull is an appropriate model species to study the interactions between
94 fisheries and seabirds. This is particularly true given the forthcoming implementation of
95 the European landing obligation, as the Audouin's gull is known to greatly rely on
96 discards (Oro *et al.*, 1999; Arcos *et al.*, 2001). Many of the previous assessments on the
97 interaction between seabirds and fisheries have considered either the temporal or the
98 spatial dimension, with special attention to the latter (Cama *et al.*, 2011; Cama *et al.*,
99 2013; Yorio *et al.*, 2010, Bécarea *et al.* 2015). Few have addressed this issue by
100 integrating both dimensions simultaneously (but see Votier *et al.*, 2010; Granadeiro *et al.*,

2014). This approach can provide further insights into these interactions and can be key to assessing or predicting possible responses or consequences for seabirds in advance of changes in discard availability. Based on GPS tracking data for gulls and fishing vessels (Vessel Monitoring System, hereafter VMS), we combined temporal and spatially-explicit information to evaluate gull-fishery co-occurrences in both space and time. This allowed us to investigate at a finer scale how the Audouin's gull interacts with the fishing fleet of the NE Iberian Peninsula in a *pre-ban* scenario. Regarding the association between the Audouin's gull and the resources provided by the fisheries, we predicted that the gulls will adjust their feeding strategies to the activity patterns of the co-occurring fishing vessels. In other words, we predicted that the interactions would occur at those times and locations at which the fishing boats generate the best feeding opportunities. On the other hand, and taking into account the existence of individual and distinguishable strategies within the opportunistic/generalist species' populations (e.g., Navarro et al. 2010), we predicted a heterogenic usage of this trophic resource between individuals of this population. The information provided could be useful in making comparisons to *post-ban scenarios*, and to assessing other future changes in the interaction of birds and fisheries, especially when human food subsidies such as discards are involved in the interaction.

MATERIAL AND METHODS

Study area and species

The study area was defined from the movements of the GPS-tracked Audouin's gulls breeding at the Punta de la Banya colony (40°40'N, 0°45'E), a protected sandy peninsula with salt pans in the Ebro Delta Natural Park (NE Spain) (Fig. 1). The area comprised the NE Levantine coast of Spain and extended from the coast over the continental shelf to the upper slope. There are numerous fishing ports scattered along the coast of the study area,

126 which is the most important fishing ground for clupeids and demersal resources in the
127 Mediterranean due to the wide continental shelf and the nutrients contributed by the Ebro
128 River (Maynou *et al.*, 2008). This supports two main fishing activities: trawling (diurnal
129 activity 7:00 h to 17:00 h GMT+1, Fig. 2A and 2B) and purse seining (nocturnal activity,
130 starting at 23:00 h GMT+1 and with no return limit, Fig. 2A and 2C). The fishing activity
131 of both fleets is concentrated on the weekdays (Monday to Friday), with no fishing
132 activity on the weekend. Trawling is a non-selective fishing practice that produces large
133 quantities of discards (Stithou *et al.*, 2019). These discards are thrown back to the sea
134 after every trawl and two to four trawls can be carried out per day. In the Ebro Delta, the
135 trawling fishing vessels begin to produce discards around 11:00 h. However, it is at the
136 end of the fishing day, between 16:00 h and 17:00 h, when all the fishing vessels discard
137 simultaneously as they approach the fishing ports. This results in an abundant and highly
138 predictable food resource for marine scavengers (Martínez-Abraín *et al.*, 2002; Karris *et*
139 *al.*, 2018). This contrasts with the nocturnal purse-seining activity, which produces few
140 discards but can affect the foraging behaviour of scavengers through a process of resource
141 facilitation, as it concentrates epipelagic fish close to the surface (Arcos and Oro, 2002).
142 The study period coincided with a trawling moratorium established north of the Ebro
143 River (Fig. 1).

144 The Audouin's gulls breeding at the Ebro Delta typically share their foraging
145 distribution between terrestrial (mainly rice fields) and marine areas close to the colony
146 (Bécares *et al.* 2015; Christel *et al.* 2012), where they often interact with fishing vessels
147 (Oro *et al.*, 1996). This colony has seen some fluctuation in its numbers. Before the 1980s,
148 the Audouin's gull was a scarce species in the Mediterranean, but during the 1980s and
149 the 1990s, the studied colony in the Ebro Delta grew exponentially coinciding with the
150 development of the fishing activity in the study area (Oro and Martinez-Vilalta, 1992).

151 This exponential growth was likely due to the exploitation of the highly available human
152 subsidies, particularly of discards. In 2011, when the study was carried out there were
153 11.967 breeding pairs, representing ca. 60% of the global population (Ebro Delta Natural
154 Park, *personal communication*).

155 **Fieldwork procedure**

156 Between the 8th and the 26th of May 2011, 60 breeding gulls were captured in randomly
157 chosen nests, with either box or tent-labelled traps (Bub, 1991), and equipped with
158 CatTrack GPS loggers (Perthold, 2011). This coincided with the incubation period of the
159 species. These loggers were programmed to record locations (10 m accuracy, Perthold,
160 2011) every 5 minutes. Devices were sealed using a rubber shrink tube to make them
161 waterproof, and attached to the back of the gulls using a Teflon adjustable harness
162 (Bécares et al., 2010). The total weight of sealed devices (ca. 25 g) roughly represented
163 3-5% of the bird's body mass, thus below the limit for deleterious effects on individual
164 birds (Wilson *et al.*, 2002; Phillips *et al.*, 2003). Thirty-six tagged birds were recaptured
165 between one and two weeks after the deployment of GPS devices. Recorded data included
166 GPS positions for these 36 individuals between May 8th and 26th. No adverse weather
167 conditions (e.g. rain or strong winds) that could potentially affect gulls' foraging
168 behaviour occurred during the study period.

169

170 **Data analyses**

171 *Habitat use*

172 We compared the differential use of the sea by the gulls on weekdays (Monday to
173 Friday; period with fishing activity) and weekends (Saturday and Sunday; period without
174 fishing activity) using the proportion of the time spent at sea or inland (mainly in rice

175 fields). A foraging trip was defined to include the locations from when a bird left the
176 colony until it returned (BirdLife International, 2005). For each study day, all the foraging
177 trips were taken into account. The proportion of time spent at sea on each foraging trip in
178 terms of the trip total duration was calculated, thus obtaining the daily use of the sea by
179 the gulls. A linear mixed model was fitted, with the logit transformation of the proportion
180 of time spent in each habitat as the dependent variable and the type of day (i.e. weekday,
181 Monday to Friday, and weekend, Saturday and Sunday) as the explanatory variable.

182 *Activity rhythms & Gull-fisheries interaction*

183 We performed a waveform analysis on the daily use of the sea of both the fishing vessels
184 and the gulls, to determine their daily temporal patterns of activity (Fig. 2A). GPS-
185 locations for gulls were grouped into 2-hour intervals, following the temporal resolution
186 for VMS data. Total number of bird or boat positions per time interval were subsequently
187 averaged to obtain a representative 24 h profile (the waveform) of the 17 days of
188 sampling. The phase, defined as the significant increase in sea use by gulls and fisheries,
189 was determined for each waveform by calculating the Midline Estimating Statistic Of
190 Rhythm (MESOR; Aguzzi *et al.*, 2015). The MESOR was computed by re-averaging all
191 waveform values and was plotted as a threshold in the waveform plot. Waveform values
192 above the MESOR indicated a significant use of the sea in a cyclic way, i.e., the phase.

193 We combined spatio-temporal information on the distribution of gulls and fishing
194 vessels to assess gull-fishery interaction. To do so, we first retained bird positions within
195 a 500 m and 20 min (± 10 min) buffer around fishing vessel positions (based on VMS).
196 We selected this spatial threshold after a sensitivity analysis revealed that the number of
197 individuals interacting within a given spatial buffer increased between 0 m and 200 m,
198 but that it stabilised between 300 m and 500 m. Thus, we selected the 500 m buffer for a
199 more conservative approach. Filtered positions were subsequently included in a

200 waveform analysis to test when the interactions occurred, and the interaction magnitude
201 (as revealed by the number of bird positions within our spatiotemporal buffer) in a
202 specific time interval. Finally, we carried out a kernel analysis of density for the
203 interacting positions to assess where the interactions were produced.

204

205 *The individual component of seabird-fishery interactions*

206 We assessed the repeatability in the individual feeding strategies to consider whether
207 there were different strategies within the population regarding fishing discard use. For
208 every gull trip, we calculated a minimum, dimensionless distance between gulls and
209 fishing boats that accounted for both the temporal and the spatial dimensions and was
210 standardised to the above-defined spatiotemporal buffer (see details in supporting
211 information). This minimum distance can be interpreted as the number of spatiotemporal
212 buffers between a fishing vessel and a gull for a specific trip. We used these distances as
213 an indicator of the degree of gull-fishery interaction to assess the repeatability in feeding
214 strategies. These strategies can range from a complete dependence on fishing discards
215 (high degree of interaction – consistently small distances) to a complete independence
216 (low degree of interaction – consistently large distances), with some intermediate
217 strategies. The repeatability was evaluated by estimating the intraclass correlation
218 coefficient (ICC) using a linear mixed model. ICC values range from 0 to 1, with higher
219 ICC values indicating high intra-individual repeatability in feeding strategies, while lower
220 ICC values denote individuals behaving randomly.

221 R Statistical Software was used to compute the spatiotemporal buffers, and to carry out
222 the kernel analysis and all statistical analyses (R Development Core Team, 2008).

223

224 **RESULTS**

225 Based on 36.251 recorded positions outside the colony, Audouin's gulls preferentially
226 used the terrestrial environment (21.084 positions) with less contribution by the marine
227 environment (15.167 positions) to overall habitat use by gulls. However, the relative
228 contributions to habitat use differed between weekdays and weekends, with a three-fold
229 (CI 95%: 1.85 – 4.84; $p < 0.001$) greater proportion of time spent at sea during weekdays.

230 Regarding the temporal dimension, gulls' daily activity patterns also differed between
231 weekdays and weekends. On weekdays, the temporal patterns of sea use by gulls matched
232 those of trawling boats; i.e., from 7:00 h to 17:00 h, thus both being diurnal (Fig. 2A). In
233 the case of the purse seiners, the phase of the activity pattern was from 22:00 h to 8:00 h
234 (Fig. 2A), and thus coincided less with that of the gulls. However, there was an overlap
235 of 2 hours, from 6:00 h to 8:00 h, when both fleets and the gulls were at sea. On weekends,
236 when there was no fishing activity, the temporal pattern of sea usage by the gulls changed
237 noticeably, with significant sea use from 23:00 h to 6:00 h, and thus being mainly
238 nocturnal (Fig. 3B).

239 We detected a stepwise increase in the magnitude of gull interaction with trawlers from
240 9:00 h to 16:00 h, with a maximum at 16:00 h (38.8%). The magnitude of birds'
241 interactions with trawlers sharply decreased after 16:00 h, reaching 8.3% by 18:00 h.
242 Regarding interactions with purse seiners, there was a marked increase in the magnitude
243 of the interactions from 2:00 h to 4:00 h, during which interactions peaked (33.1%). From
244 4:00 h on, there was a clear decrease in the magnitude of the interaction, reaching 8.3%
245 just before 10:00 h (Fig. 3A). Standard deviations for every analysis are provided in Table
246 1.

247 Regarding the geographic location of seabird-fishery interactions, our results showed
248 that the interactions were concentrated within the 30 southernmost kilometres of the
249 colony, near the main fishing port of the Ebro Delta (Fig. 3C).

250 Using the minimum distances as an indicator of gull-fishery interaction, we estimated
251 that 47.6% of the trips at sea entered the spatiotemporal buffer (500 m, \pm 10 min.) of
252 distance around fishing vessels. However, we also detected a large heterogeneity among
253 individuals showing minimum distances that ranged from 0.02 to 19.3 buffers of distance
254 thus indicating a degree of structuring in the foraging strategies within the population
255 (Fig.4). We observed a repeatability of 34% (ICC: 0.34, 95% CI: 0.1963 – 0.4765, $p <$
256 0.001), a value rated as “fair” on the scale provided in Landis and Koch, 1977.

257

258 **DISCUSSION**

259 It is well known that fishing activity provides substantial food for opportunistic seabirds
260 (Tasker *et al.*, 2000). However, most studies on this topic have addressed the seabird-
261 fishery interaction either from a temporal or a spatial perspective (Yorio *et al.*, 2010),
262 with few studies integrating the two dimensions simultaneously (Votier *et al.*, 2010;
263 Granadeiro *et al.*, 2011; Bécarea *et al.*, 2015). Based on our new spatiotemporal approach,
264 we show that Audouin’s gulls scheduled their behaviour to that of fishing vessels, either
265 to benefit from fish concentrations near the surface or to exploit fishing discards.
266 However, this feeding strategy was not homogeneous within the population. As predicted,
267 there was a differential usage of the discards by the different tracked individuals, thus
268 showing the characteristic behaviour of opportunistic species, with a highly plastic
269 trophic strategy (Navarro *et al.* 2010; Ceia *et al.*, 2014). These results provide further
270 insights into the dependence of scavenger communities on human food subsidies, but may

271 also have implications for the management and conservation of these species, particularly
272 within the current context of changes in fishing policies.

273 Opportunistic scavenger species are capable of shaping their schedules and their use of
274 habitat depending on human activities (Tyson *et al.*, 2015). Accordingly, the Audouin's
275 gulls in our study showed a differential use of the sea depending on whether it was a
276 weekday or a weekend. This result suggests that there is a driver favouring this day-
277 dependent habitat usage. The fishing activity could be a suitable driver to explain this
278 difference, as this activity is only carried out from Monday to Friday in the study area
279 (Bécares *et al.*, 2015).

280 The daily pattern of sea-use by gulls matched that of the trawlers and purse seiners on
281 weekdays, providing some evidence regarding gull-fishery interactions. A study by
282 Bécares *et al.* (2015) showed a spatial overlap between Audouin's gulls and fishing fleets.
283 However, an interaction cannot be assumed when assessing temporal and spatial overlaps
284 independently. A spatial overlap does not necessarily imply a temporal one nor does a
285 temporal overlap imply that birds and boats are exploiting the same areas. Quantitative
286 assessments on seabird-fishery interactions require a detailed spatiotemporal approach as
287 the one provided in the current study.

288 The interaction with trawlers can be explained by the great amount of discards provided
289 (Stithou *et al.*, 2019). The magnitude of the interaction increased gradually beginning at
290 the start of the fishing day until 16:00 h, when it reached its maximum. This peak of
291 interaction coincides with the time at which trawlers are returning to port and are thus
292 more concentrated and closer to the coast and the colony (i.e., when discards are most
293 readily available).

294 Purse seiners operate nocturnally, and produce few discards (Arcos and Oro, 2002).
295 There was a sharp increase in the magnitude of interaction from 4:00 h to 6:00 h (i.e., half
296 of the working day for purse seiners). At that time, the nets are usually pulled out of the
297 water and there is a large concentration of available fish at the surface (Arcos and Oro,
298 2002). Purse seiners also use a large lamp to attract fish improving visibility for birds
299 making the fish easier to catch (Arcos and Oro, 2002). This can be considered a process
300 of resource facilitation (Daleo *et al.*, 2005), as it allows the gulls to easily pick fish from
301 the surface by dipping (Gaston, 2004). This is somewhat similar to the natural feeding
302 strategy of the Audouin's gull. In natural circumstances (those not associated with
303 fisheries), the Audouin's gull takes advantage of the diel vertical migrations that some
304 small epipelagic fish carry out at night, from the epipelagic zone, to near-surface waters,
305 catching fish at the surface (Blaxter and Hunter, 1982; Arcos and Oro, 2002). This species
306 has also been seen interacting with tuna schools (Oro, 1995), similar to purse seine
307 facilitation, as they pull fish to the surface. However, interacting with purse seiners may
308 be energetically costlier than interacting with trawlers, as it has been shown that discards
309 from trawlers can supply enough energy to get through the breeding season (Oro, 1999),
310 whereas with purse seiners, birds must catch live fish themselves. In regards to the
311 nocturnal activity, this could be explained by individual specialisation or some sort of
312 competitive exclusion (Hardin, 1960), as discarding (produced by the diurnal activity of
313 trawlers) generates highly competitive interactions (Calado *et al.*, 2018).

314 Our spatio-temporal approach revealed that gulls interact with fishing vessels (i.e.,
315 entered the spatiotemporal buffer) during ca. 50% of their trips to sea. However, these
316 values are very likely underestimated, as the VMS data are collected every two hours,
317 thus limiting the evaluation of interactions to two-hour intervals. Despite this constraint,
318 we were able to detect an individual component in seabird-fishery interactions, thus

319 indicating individual strategies within this population. Individual specialisation refers to
320 the use of available resources resulting in inter-individual niche variation. This inter-
321 individual variation could be consistent over short- and/or long-term periods, and is
322 known to be widespread across a diverse set of taxa (Bolnick *et al.*, 2003). It is especially
323 common amongst generalist predators (Woo *et al.*, 2008), like the Audouin's gull
324 (Christel *et al.*, 2012). In fact, many generalist populations are composed of ecologically
325 heterogeneous individuals that repeatedly differ in foraging behaviour and use different
326 subsets of the available resources (Bolnick *et al.*, 2003). The specialisation of seabirds
327 on fisheries discards was assessed in Tyson *et al.* (2014), which identified some discards
328 specialists amongst a population of lesser black-backed gulls in the Wadden Sea.

329 Our results on the individual component of the interactions with the two fishing fleets
330 were obtained using the ICC. The ICC result (repeatability of 34%) suggests that there
331 was some consistency in the feeding strategies, with some individuals showing a discard-
332 dependent strategy, and others tending to feed naturally. However, rather than a bimodal
333 distribution of foraging strategies, most individuals showed intermediate foraging
334 strategies within this gradient. The reasons for the individual specialisation in our study
335 case could be related to the stage of the annual cycle of the gulls (i.e., the breeding
336 season). All seabirds are central-place foragers during the breeding season (Rayner *et al.*
337 *et al.*, 2010), which could lead to partitioning in the use of available resources between
338 individuals, according to the principle of competitive exclusion (Ceia and Ramos, 2015).

339 The fact that some individuals tend to interact more often with fisheries than others is
340 important in terms of conservation of the Audouin's gull colony in a *post-ban* scenario.
341 These birds could be more affected by a depletion in discards suggesting that the effects
342 of the discard ban will not be homogeneous across the Audouin's gull population. In
343 contrast, the presence of individuals that do not rely on discards would be key to the

344 population overcoming the discard ban, as these individuals would be able to feed more
345 easily in the absence of discards.

346 The location where the interactions took place was also assessed using a spatiotemporal
347 buffers. The results identified an area located ca. 30 Km south of the colony for both types
348 of fishing gear. However, the fishing activity is carried out all along the Levantine Iberian
349 coast (Fig. 2B, 2C), with some hotspots in particular locations. In addition, a trawling
350 moratorium was implemented north of colony, such that interactions could not occur in
351 that area. On the other hand, the proximity of the interaction hotspot to the colony can
352 again be explained by the breeding stage and central-place foraging, as during the
353 breeding season, the birds are energetically constrained and do not search far from the
354 colony to find their prey (Orians and Pearson, 1979).

355

356 *Conclusions and perspectives*

357 In this study, we showed the fine-scale spatiotemporal overlap between the Audouin's
358 gull and the fisheries operating near the Ebro Delta. We interpreted this spatiotemporal
359 overlap as an interaction between gulls and fishing vessels. The interactions were not
360 constant throughout the day and showed some variability depending on the fishing fleet
361 and the time of the day, with a larger magnitude of interaction during either discarding or
362 resource facilitation. Furthermore, differences between individuals regarding the feeding
363 strategy were found, indicating a lack of homogeneity within the population. The studied
364 population has seen some fluctuations in the last 40 years (García-Tarrasón, 2014), with
365 a demographic explosion coinciding with an increase in fishing activity in the area. This
366 indicates the importance of the fishing activity, and particularly trawling, for the
367 subsistence of this population. However, since 2009, the population has declined to about

368 2000 pairs. Fishing activity has not changed significantly in the study area, and discards
369 are still produced. Thus, the decline in the gull population could be attributed to other
370 factors, such as predation episodes. Therefore, human subsidies provided by fishing
371 activity (i.e., discards) could be key to conserving the remaining pairs. However, in the
372 near future, changes to the discarding policies will be implemented in the EU (i.e.,
373 Discard ban/Landing Obligation). This could have impacts on this population as well as
374 on other species in the EU and Mediterranean scavenger community. A discard ban could
375 imply a food shortage for this species (Bicknell *et al.* 2013), as an important portion of
376 their energy is obtained from discarded fish, especially in the breeding season (Arcos and
377 Oro, 2002). For this reason, the discard-dependent individuals could contribute to a
378 decline in the population when discards are no longer available, as they may move to
379 other areas in the Mediterranean in search of higher food availability (Bicknell *et al.*,
380 2013).

381 The NW Mediterranean is known to be a hotspot for clupeiform spawning (sardine,
382 anchovy; natural prey for the Audouin's gull) (Maynou, 2008). If these lipid-rich fish
383 species are equally exploited by the local fisheries after the implementation of the ban, it
384 may be difficult for the fish stocks to recover, and thus making it difficult to meet the
385 demands for both natural-feeding gulls (and other species of the seabird community) and
386 the fishers. For this reason, it is of utmost importance to continue monitoring the seabird
387 populations (as well as the commercial fish stocks) once the ban is implemented, to take
388 appropriate measures that will minimize the impacts. If gulls are not able to overcome the
389 sudden shortage of discard availability, this dependence on discards will have been an
390 ecological trap (Schlaepfer *et al.*, 2002), as gulls will be forced to use sub-optimal habitats
391 with less food availability.

392 Finally, this study provides some information of interest to policy makers. We showed
393 that not all the areas in the study zone were of the same importance (as interactions
394 occurred in specific places around the colony) and that not all times of day were equally
395 important (as the magnitude of the interaction is not constant throughout the day). Our
396 approach allowed us to characterise the interactions between the Audouin's gull and
397 fishing boats. Moreover, this methodology can be useful to assess other aspects of
398 seabird-fishery interactions. Bycatch is a key issue regarding these interactions, as it
399 produces 500.000 seabird deaths worldwide annually (Rodriguez *et al.*, 2019), especially
400 by longliners (Anderson *et al.* 2011). The use of spatiotemporal buffers can help to
401 identify interaction hotspots, and to relate them to bycatch data, allowing for informed
402 policy proposals, well-designed marine protected areas, and other measures to protect
403 seabirds from this mortality focus. This could be of particular interest given the discard
404 ban, as many birds could start feeding near longliners, likely resulting in an increase in
405 mortality due to these interactions (Laneri *et al.* 2010). Moreover, for deeper insights into
406 how the interactions are produced, other complementary methods could be used, such as
407 on-board surveys or the deployment of bird-borne cameras (Votier *et al.*, 2013). These
408 methodologies would allow an evaluation of the interactions from another point of view
409 and would improve our understanding of these interactions. By using these methods, other
410 features about the interactions could be assessed, such as how fish are taken, or the nature
411 of interactions with other scavenging seabird species, therefore providing more
412 ecosystemic and community-based assessments.

413

414

415

416 **ACKNOWLEDGEMENTS**

417 We thank Santiago Bateman, Albert Cama, Andreia Dias, Víctor García-Matarranz,
418 Manuel García-Tarrasón, Oscar González, Lluís Jover and Jordi Prieto for their help in
419 carrying out this study. To the wildlife personnel at the Ebro Delta Natural Park (Julia
420 Piccardo, Albert Bertolero, David Bigas, Francesc Vidal, Antoni Curcó, Carles Ibàñez
421 and Jordi Martí Aledo) for the use of their facilities during fieldwork. Silvia Revenga
422 through the Centro de Seguimiento de Pesca (Secretaría General de Pesca del Ministerio
423 de Agricultura, Alimentación y Medio Ambiente) kindly provided the VMS data.
424 Fieldwork was funded under the LIFE+ INDEMARES (LIFE 07NAT/E/00732) and
425 MCyI project CGL2008- 05448-C02/BOS. Partial funding was obtained from projects
426 CGL2013-45099-P (Ministerio de Economía y Competividad) and CGL2016-08963-R
427 (AEI/FEDER, UE). JO-C was supported by a Department Collaboration grant
428 (18CO1/006033) from the MEyFP. FR was funded by “Ministerio de Economía, Industria
429 y Competitividad” (“Subprograma Juan de la Cierva-Incorporación” -IJCI-2015-24531).

430 **REFERENCES**

- 431 Abelló, P., Arcos, J. M., and Gil de Sola, L. 2003. Geographical patterns of seabird
432 attendance to a research trawler along the Iberian Mediterranean coast. 67(Suppl.
433 2):69-75
434
- 435 Aebischer, N. J., Robertson, P. A., and Kenward, R. E. 1993. Compositional Analysis of
436 Habitat Use From Animal Radio-Tracking Data. *Ecology*, 74: 1313–1325.
- 437 Aguzzi, J., Sbragaglia, V., Tecchio, S., Navarro, J., and Company, J. B. 2015. Rhythmic
438 behaviour of marine benthopelagic species and the synchronous dynamics of
439 benthic communities. *Deep Sea Research Part I: Oceanographic Research Papers*,
440 95: 1–11.
- 441 Anderson, O. R. J., Small, C. J., Croxall, J. P., Dunn, E. K., Sullivan, B. J., Yates, O.,
442 and Black, A. 2011. Global seabird bycatch in longline fisheries. *Endangered
443 Species Research*, 14: 91–106.

- 444 Arcos, J., Oro, D., and Sol, D. 2001. Competition between the yellow-legged gull *Larus*
445 *cachinnans* and Audouin's gull *Larus audouinii* associated with commercial
446 fishing vessels: the influence of season and fishing fleet. *Marine Biology*, 139:
447 807–816.
- 448 Arcos, J., and Oro, D. 2002. Significance of nocturnal purse seine fisheries for seabirds:
449 a case study off the Ebro Delta (NW Mediterranean). *Marine Biology*, 141: 277–
450 286.
- 451 Bartumeus, F., Giuggioli, L., Louzao, M., Bretagnolle, V., Oro, D., and Levin, S. A. 2010.
452 Fishery Discards Impact on Seabird Movement Patterns at Regional Scales.
453 *Current Biology*, 20: 215–222.
- 454 Bearhop, S. C. V. S., Fyfe, R., and Furness, R. W. 2008. Temporal and spatial variation
455 in the diet of a marine top predator—links with commercial fisheries. *Marine*
456 *Ecology Progress Series*, 367: 223–232.
- 457 Bécarea, J., Rodríguez, B., Arcos, P., & Ruiz, A. 2010. Técnicas de marcaje de aves
458 marinas
459 para el seguimiento remoto. *Revista de Anillamiento*, 25-26, 29-40.
- 460 Bécarea, J., García-Tarrasón, M., Villero, D., Bateman, S., Jover, L., García-Matarranz,
461 V., Sanpera, C., *et al.* 2015. Modelling terrestrial and marine foraging habitats in
462 breeding Audouin's gulls *Larus audouinii*: timing matters. *PloS One*, 10:
463 e0120799.
- 464 Bicknell, A. W. J., Oro, D., Camphuysen, K. C. J., and Votier, S. C. 2013. Potential
465 consequences of discard reform for seabird communities. *Journal of Applied*
466 *Ecology*, 50: 649–658.
- 467 Blaxter, J. H. S., and Hunter, J. R. 1982. The Biology of the Clupeoid Fishes. *In*
468 *Advances in Marine Biology*, pp. 1–223. Ed. by J. H. S. Blaxter, F. S. Russell,
469 and M. Yonge. Academic Press.
470 <http://www.sciencedirect.com/science/article/pii/S0065288108601406>
471 (Accessed 20 June 2019).
- 472 Bodey, T. W., Jessopp, M. J., Votier, S. C., Gerritsen, H. D., Cleasby, I. R., Hamer, K.
473 C., Patrick, S. C., *et al.* 2014. Seabird movement reveals the ecological footprint
474 of fishing vessels. *Current Biology*, 24: R514–R515.
- 475 Bolnick, D. I., Svanbäck, R., Fordyce, J. A., Yang, L. H., Davis, J. M., Hulsey, C. D.,
476 and Forister, M. L. 2003. The ecology of individuals: incidence and implications
477 of individual specialization. *The American Naturalist*, 161: 1–28.
478
- 479 Borges, L. 2015. The evolution of a discard policy in Europe. *Fish and Fisheries*, 16: 534–
480 540.

- 481 Brooks, M. E., Kristensen, K., Benthem, K. J. van, Magnusson, A., Berg, C. W.,
482 Nielsen, A., Skaug, H. J., *et al.* 2017. glmmTMB Balances Speed and Flexibility
483 Among Packages for Zero-inflated Generalized Linear Mixed Modeling. *The R*
484 *Journal*, 9: 378–400.
485
- 486 Bub, H. 1991. *Bird Trapping and Bird Banding: A Handbook for Trapping Methods All*
487 *Over the World*. Cornell University Press. 340 pp.
- 488 Burger, J., Gochfeld, M., Garcia, E.F.J. & Sharpe, C.J. 2019. Audouin's Gull (*Larus*
489 *audouinii*). In: del Hoyo, J., Elliott, A., Sargatal, J., Christie, D.A. & de Juana, E.
490 (eds.). *Handbook of the Birds of the World Alive*. Lynx Edicions, Barcelona.
491 (retrieved from <https://www.hbw.com/node/53971> on 20 June 2019).
- 492 Calado, J. G., Matos, D. M., Ramos, J. A., Moniz, F., Ceia, F. R., Granadeiro, J. P., and
493 Paiva, V. H. 2018. Seasonal and annual differences in the foraging ecology of two
494 gull species breeding in sympatry and their use of fishery discards. *Journal of*
495 *Avian Biology*, 49.
- 496 Calenge, C. 2006. The package “adehabitat” for the R software: A tool for the analysis of
497 space and habitat use by animals. *Ecological Modelling*, 197: 516–519.
- 498
499 Cama, A., Josa, P., Ferrer-Obiol, J., and Arcos, J. M. 2011. Mediterranean Gulls *Larus*
500 *melanocephalus* wintering along the Mediterranean Iberian coast: numbers and
501 activity rhythms in the species’ main winter quarters. *Journal of Ornithology*, 152:
502 897–907.
503
- 504 Cama, A., Bort, J., Christel, I., Vieites, D., and Ferrer, X. 2013. Fishery management has
505 a strong effect on the distribution of Audouin’s gull. *Marine Ecology Progress*
506 *Series*, 484: 279–286.
507
- 508 Ceia, F., Paiva, V., Fidalgo, V., Morais, L., Baeta, A., Crisóstomo, P., Mourato, E., *et al.*
509 2014. Annual and seasonal consistency in the feeding ecology of an opportunistic
510 species, the yellow-legged gull *Larus michahellis*. *Marine Ecology Progress*
511 *Series*, 497: 273–284.
512
- 513 Ceia, F. R., and Ramos, J. A. 2015. Individual specialization in the foraging and feeding
514 strategies of seabirds: a review. *Marine Biology*, 162: 1923–1938.
515
- 516 Christel, I., Navarro, J., del Castillo, M., Cama, A., and Ferrer, X. 2012. Foraging
517 movements of Audouin’s gull (*Larus audouinii*) in the Ebro Delta, NW
518 Mediterranean: A preliminary satellite-tracking study. *Estuarine, Coastal and*
519 *Shelf Science*, 96: 257–261.
520

- 521 Cury, P. M., Boyd, I. L., Bonhommeau, S., Anker-Nilssen, T., Crawford, R. J. M.,
522 Furness, R. W., Mills, J. A., *et al.* 2011. Global Seabird Response to Forage Fish
523 Depletion--One-Third for the Birds. *Science*, 334: 1703–1706.
- 524 Daleo, P., Escapa, M., Isacch, J. P., Ribeiro, P., and Iribarne, O. 2005. Trophic facilitation
525 by the oystercatcher *Haematopus palliatus* Temminck on the scavenger snail
526 *Buccinanops globulosum* Kiener in a Patagonian bay. *Journal of Experimental*
527 *Marine Biology and Ecology*, 325: 27–34.
- 528 Damalas, D. 2015. Mission impossible: Discard management plans for the EU
529 Mediterranean fisheries under the reformed Common Fisheries Policy. *Fisheries*
530 *Research*, 165: 96–99.
- 531 Egozcue, J. J., Pawlowsky-Glahn, V., Mateu-Figueras, G., and Barceló-Vidal, C. 2003.
532 Isometric Logratio Transformations for Compositional Data Analysis.
533 *Mathematical Geology*, 35: 279–300.
- 534 Essington, T. E., Beaudreau, A. H., and Wiedenmann, J. 2006. Fishing through marine
535 food webs. *Proceedings of the National Academy of Sciences*, 103: 3171–3175.
- 536 FAO. The State of World Fisheries and Aquaculture 2018 - Meeting the sustainable
537 development goals. 2018. Rome. Licence: CC BY-NC-SA 3.0 IGO.
- 538 Galí, M., Levasseur, M., Devred, E., Simó, R., and Babin, M. 2018. Sea-surface
539 dimethylsulfide (DMS) concentration from satellite data at global and regional
540 scales. *Biogeosciences*, 15: 3497–3519.
- 541 García-Tarrasón, M. 2014. Trophic ecology, habitat use and ecophysiology of Audouin's
542 Gull (*Larus audouinii*) in the Ebro Delta. <http://rgdoi.net/10.13140/2.1.4711.5045>
543 (Accessed 5 June 2019).
- 544 Gaston, A. J. 2004. *Seabirds: a natural history*. Yale University Press, New Haven.
- 545 Gonzalez Zevallos, D. R., and Yorio, P. M. 2006. Seabird use of discards and incidental
546 captures at the Argentine hake trawl fishery in the Golfo San Jorge, Argentina.
547 <http://ri.conicet.gov.ar/handle/11336/16764> (Accessed 20 June 2019).
- 548 Granadeiro, J. P., Brickle, P., and Catry, P. 2014. Do individual seabirds specialize in
549 fisheries' waste? The case of black-browed albatrosses foraging over the
550 Patagonian Shelf. *Animal Conservation*, 17: 19–26.
- 551 Guillen, J., Holmes, S. J., Carvalho, N., Casey, J., Dörner, H., Gibin, M., Mannini, A., *et*
552 *al.* 2018. A Review of the European Union Landing Obligation Focusing on Its
553 Implications for Fisheries and the Environment. *Sustainability*, 10: 900.

- 554 Halpern, B. S., Frazier, M., Potapenko, J., Casey, K. S., Koenig, K., Longo, C., Lowndes,
555 J. S., *et al.* 2015. Spatial and temporal changes in cumulative human impacts on
556 the world's ocean. *Nature Communications*, 6: 1–7.
557
- 558 Hardin, G. J. 1960. The competitive exclusion principle. *Science*, 131: 1292–1297.
- 559 Hudson, A. V., and Furness, R. W. 1988. Utilization of discarded fish by scavenging
560 seabirds behind whitefish trawlers in Shetland. *Journal of Zoology*, 215: 151–166.
- 561 Karris, G., Ketsilis-Rinis, V., Kalogeropoulou, A., Xirouchakis, S., Machias, A., Maina,
562 I., and Kavadas, S. 2018. The use of demersal trawling discards as a food source
563 for two scavenging seabird species: a case study of an eastern Mediterranean
564 oligotrophic marine ecosystem. *Avian Research*, 9: 26.
- 565 Landis, J. R., and Koch, G. G. 1977. The measurement of observer agreement for
566 categorical data. *Biometrics*, 33: 159–174.
- 567 Laneri, K., Louzao, M., Martínez-Abraín, A., Arcos, J. M., Belda, E. J., Guallart, J.,
568 Sánchez, A. D., *et al.* 2010. Trawling regime influences longline seabird
569 bycatch in the Mediterranean: new insights from a small-scale fishery.
- 570 Lewison, R. L., Crowder, L. B., Wallace, B. P., Moore, J. E., Cox, T., Zydalis, R.,
571 McDonald, S., *et al.* 2014. Global patterns of marine mammal, seabird, and sea
572 turtle bycatch reveal taxa-specific and cumulative megafauna hotspots.
573 *Proceedings of the National Academy of Sciences*, 111: 5271–5276.
- 574 Louzao, M., Arcos, J. M., Guijarro, B., Valls, M., and Oro, D. 2011. Seabird-trawling
575 interactions: factors affecting species-specific to regional community utilisation
576 of fisheries waste: Seabird-trawler interactions: species versus community.
577 *Fisheries Oceanography*, 20: 263–277.
- 578 Lynam, C. P., Llope, M., Möllmann, C., Helaouët, P., Bayliss-Brown, G. A., and
579 Stenseth, N. C. 2017. Interaction between top-down and bottom-up control in
580 marine food webs. *Proceedings of the National Academy of Sciences of the*
581 *United States of America*, 114: 1952–1957.
- 582 Martínez-Abraín, A., Maestre, R., and Oro, D. 2002. Demersal trawling waste as a food
583 source for Western Mediterranean seabirds during the summer. *ICES Journal of*
584 *Marine Science*, 59: 529–537.
585
- 586 Maynou, F., Olivar, M. P., and Emelianov, M. 2008. Patchiness and spatial structure of
587 the early developmental stages of clupeiforms in the NW Mediterranean Sea.
588 *Journal of Plankton Research*, 30: 873–883.
- 589 Monsarrat, S., Benhamou, S., Sarrazin, F., Bessa-Gomes, C., Bouten, W., and Duriez, O.
590 2013. How Predictability of Feeding Patches Affects Home Range and Foraging
591 Habitat Selection in Avian Social Scavengers? *PLOS ONE*, 8: e53077.

- 592 Navarro, J., Oro, D., Bertolero, A., Genovart, M., Delgado, A., and Forero, M. G. 2010.
593 Age and sexual differences in the exploitation of two anthropogenic food
594 resources for an opportunistic seabird. <https://digital.csic.es/handle/10261/59251>
595 (Accessed 18 December 2019). 11: 2453–2459
596
- 597 Nevitt, G. A., Veit, R. R., and Kareiva, P. 1995. Dimethyl sulphide as a foraging cue for
598 Antarctic Procellariiform seabirds. *Nature*, 376: 680.
- 599 Nevitt, G. A., and Bonadonna, F. 2005. Sensitivity to dimethyl sulphide suggests a
600 mechanism for olfactory navigation by seabirds. *Biology Letters*, 1: 303–305.
- 601 Orians, G. H., and N. E. Pearson. 1979. On the theory of central place foraging. Pages
602 155–177 in *Analysis of Ecological Systems* (D. J. Horn, R. D. Mitchell, and G. R.
603 Stairs, Eds.). Ohio State University Press, Columbus
- 604 Oro, D. 1992. The colony of the Audouin's Gull at the Ebro Delta. *Avocetta*. 16: 98-101
- 605 Oro, D. 1995. Audouin's Gulls *Larus audouinii* Associate with Sub-surface Predators in
606 the Mediterranean Sea. *Kurze Mitteilungen*.
- 607 Oro, D., Jover, L., and Ruiz, X. 1996. Influence of trawling activity on the breeding
608 ecology of a threatened seabird, Audouin's gull *Larus audouinii*. *Marine Ecology*
609 *Progress Series*, 139: 19–29.
- 610 Oro, D. 1999. Trawler discards: a threat or a resource for opportunistic seabirds? In:
611 Adams, N.J. & Slotow, R.H. (eds) *Proc. 22 Int. Ornithol. Congr.*, Durban: 717-
612 730. Johannesburg: BirdLife South Africa.
- 613 Oro, D., Pradel, R., and Lebreton, J.-D. 1999. Food availability and nest predation
614 influence life history traits in Audouin's gull, *Larus audouinii*. *Oecologia*, 118:
615 438–445.
- 616 Oro, D., Bertolero, A., Fernández Chacón, A., Genovart, M., Igual, M., Piccardo, J. 2010.
617 Seguiment científic de la colònia de gavina corsa *Larus audouinii* a la reserva
618 natural parcial de la punta de la Banya, Parc Natural del Delta de l'Ebre.
619 Generalitat de Catalunya-CSIC. Unpublished report.
620
- 621 Oro, D., Genovart, M., Tavecchia, G., Fowler, M. S., and Martínez-Abraín, A. 2013.
622 Ecological and evolutionary implications of food subsidies from humans. *Ecology*
623 *Letters*, 16: 1501–1514.
- 624 Pauly, D. 1998. Fishing Down Marine Food Webs. *Science*, 279: 860–863.
- 625 Phillips, R. A., Xavier, J. C., and Croxall, J. P. 2003. Effects of Satellite Transmitters on
626 Albatrosses and Petrels. *The Auk*, 120: 1082–1090.

- 627 Perthold, J. CatTrack1 - User Manual CatTrack I - GPS Position Logger. 2011
628 http://www.mrlee-catcam.de/BINARY/CatTrack1_User_Manual.pdf.
- 629 Phillips, R. A., Ridley, C., Reid, K., Pugh, P. J. A., Tuck, G. N., and Harrison, N. 2010.
630 Ingestion of fishing gear and entanglements of seabirds: Monitoring and
631 implications for management. *Biological Conservation*, 143: 501–512.
- 632 Pierre, J. P., Abraham, E. R., Richard, Y., Cleal, J., and Middleton, D. A. J. 2012.
633 Controlling trawler waste discharge to reduce seabird mortality. *Fisheries*
634 *Research*, 131–133: 30–38.
- 635 R Development Core Team. R: A language and environment for statistical computing. R
636 Foundation for Statistical Computing, Vienna, Austria. ISBN 3-900051-07-0,
637 URL. <http://www.R-project.org>. 2008.
- 638 Ramírez, F., Afán, I., Davis, L. S., and Chiaradia, A. 2017. Climate impacts on global hot
639 spots of marine biodiversity. *Science Advances*, 3: e1601198.
640
- 641 Rayner, M. J., Hartill, B. W., Hauber, M. E., and Phillips, R. A. 2010. Central place
642 foraging by breeding Cook’s petrel *Pterodroma cookii*: foraging duration reflects
643 range, diet and chick meal mass. *Marine Biology*, 157: 2187–2194.
- 644 Rodríguez, A., Arcos, J. M., Bretagnolle, V., Dias, M. P., Holmes, N. D., Louzao, M.,
645 Provencher, J., *et al.* 2019. Future Directions in Conservation Research on
646 Petrels and Shearwaters. *Frontiers in Marine Science*, 6.
- 647 Schlaepfer, M. A., Runge, M. C., and Sherman, P. W. 2002. Ecological and
648 evolutionary traps. *Trends in Ecology & Evolution*, 17: 474–480.
- 649 Schmitt, S., Pouteau, R., Justeau, D., Boissieu, F. de, and Birnbaum, P. 2017. ssdm: An r
650 package to predict distribution of species richness and composition based on
651 stacked species distribution models. *Methods in Ecology and Evolution*, 8: 1795–
652 1803.
- 653 Stithou, M., Vassilopoulou, V., Tsagarakis, K., Edridge, A., Machias, A., Maniopoulou,
654 M., Dogrammatzi, A. 2019. Discarding in Mediterranean trawl fisheries—a
655 review of potential measures and stakeholder insights. *Maritime Studies*.
- 656 Tasker, M. L., Camphuysen, C. J., Cooper, J., Garthe, S., Montevecchi, W. A., and
657 Blaber, S. J. M. 2000. The impacts of fishing on marine birds. *ICES Journal of*
658 *Marine Science*, 57: 531–547.
- 659 Tyson, C., Shamoun-Baranes, J., Van Loon, E. E., Camphuysen, K. (C J.), and Hintzen,
660 N. T. 2015. Individual specialization on fishery discards by lesser black-backed
661 gulls (*Larus fuscus*). *ICES Journal of Marine Science*, 72: 1882–1891.

- 662 BirdLife International. 2005. Tracking Ocean Wanderers: The Global Distribution of
663 Albatrosses and Petrels. *The Auk*, 122: 1307.
- 664 Votier, S. C., Bearhop, S., Witt, M. J., Inger, R., Thompson, D., and Newton, J. 2010.
665 Individual responses of seabirds to commercial fisheries revealed using GPS
666 tracking, stable isotopes and vessel monitoring systems. *Journal of Applied*
667 *Ecology*, 47: 487–497.
- 668 Votier, S. C., Bicknell, A., Cox, S. L., Scales, K. L., and Patrick, S. C. 2013. A Bird's
669 Eye View of Discard Reforms: Bird-Borne Cameras Reveal Seabird/Fishery
670 Interactions. *PLOS ONE*, 8: e57376.
671
- 672 Weimerskirch, H., Capdeville, D., and Duhamel, G. 2000. Factors affecting the number
673 and mortality of seabirds attending trawlers and long-liners in the Kerguelen area.
674 *Polar Biology*, 23:236-249.
- 675 Wilson, R. P., Grémillet, D., Syder, J., Kierspel, M. A. M., Garthe, S., Weimerskirch, H.,
676 Schäfer-Neth, C., *et al.* 2002. Remote-sensing systems and seabirds: their use,
677 abuse and potential for measuring marine environmental variables. *Marine*
678 *Ecology Progress Series*, 228: 241–261.
- 679 Yorio, P., Quintana, F., Dell'arciprete, P., and González-Zevallos, D. 2010. Spatial
680 overlap between foraging seabirds and trawl fisheries: implications for the
681 effectiveness of a marine protected area at Golfo San Jorge, Argentina. *Bird*
682 *Conservation International*, 20: 320–334.
683
- 684 Zeller, D., Cashion, T., Palomares, M., and Pauly, D. 2018. Global marine fisheries
685 discards: A synthesis of reconstructed data. *Fish and Fisheries*, 19: 30–39.
- 686 Zuur, A., Ieno, E. N., Walker, N., Saveliev, A. A., and Smith, G. M. 2009. *Mixed Effects*
687 *Models and Extensions in Ecology with R. Statistics for Biology and Health.*
688 Springer-Verlag, New York. <https://www.springer.com/gp/book/9780387874579>
689 (Accessed 25 October 2019)

Table 1. Standard deviations for all the averages of the number of positions or individuals in the waveform analysis shown in Figure 3.

Hour Interval	Trawling interaction % positions SD	Purse seining interaction % positions SD	Gulls' % positions weekend SD
0-2	0	0	13.80
2-4	0	9.83	11.90
4-6	0	36.61	7.30
6-8	14.64	25.73	3.08
8-10	6.55	17.61	6.83
10-12	12.28	7.94	8.57
12-14	10.83	0	6.42
14-16	19.43	0	7.74
16-18	21.78	0	6.72
18-20	10.04	0	6.35
20-22	4.41	0	6.08
22-24	0	2.27	15.30

Figure captions:

Figure 1. Distribution of gulls' GPS positions. The area between A and B was under a trawling moratorium.

Figure 2. Daily cycles of boat activity (A), density of trawler positions (B), density of purse seiner positions (C). Shaded area in A shows the area above MESOR.

Figure 3. Daily cycles of interactions between gulls and boats (A), gulls at sea during the weekend (B). Kernel analysis of the interaction locations (C). Shaded area in A and B show the area above MESOR.

Figure 4. Boxplot showing the minimum distance per individual. Ordered by median value.

Figure1.

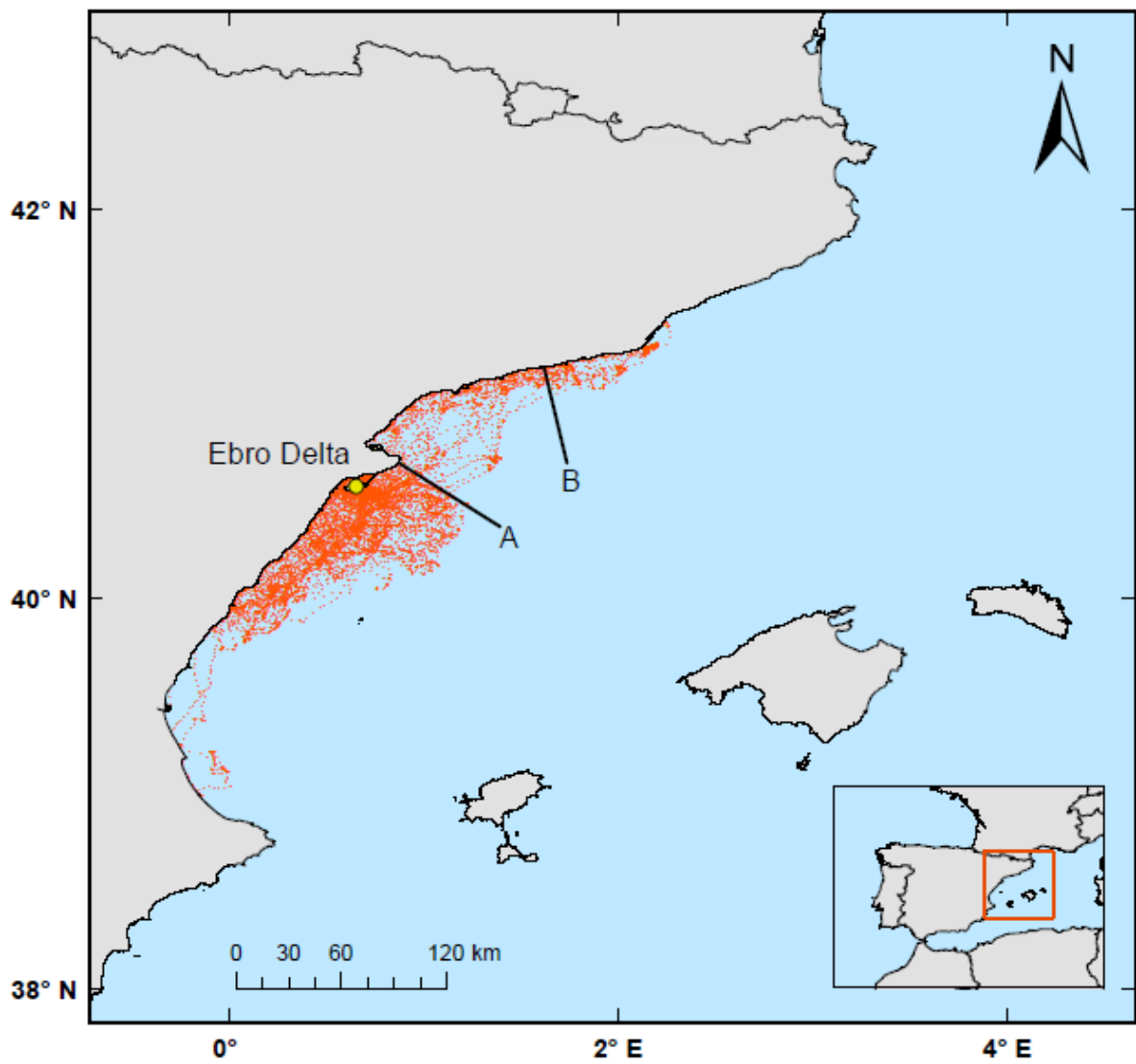


Figure 2.

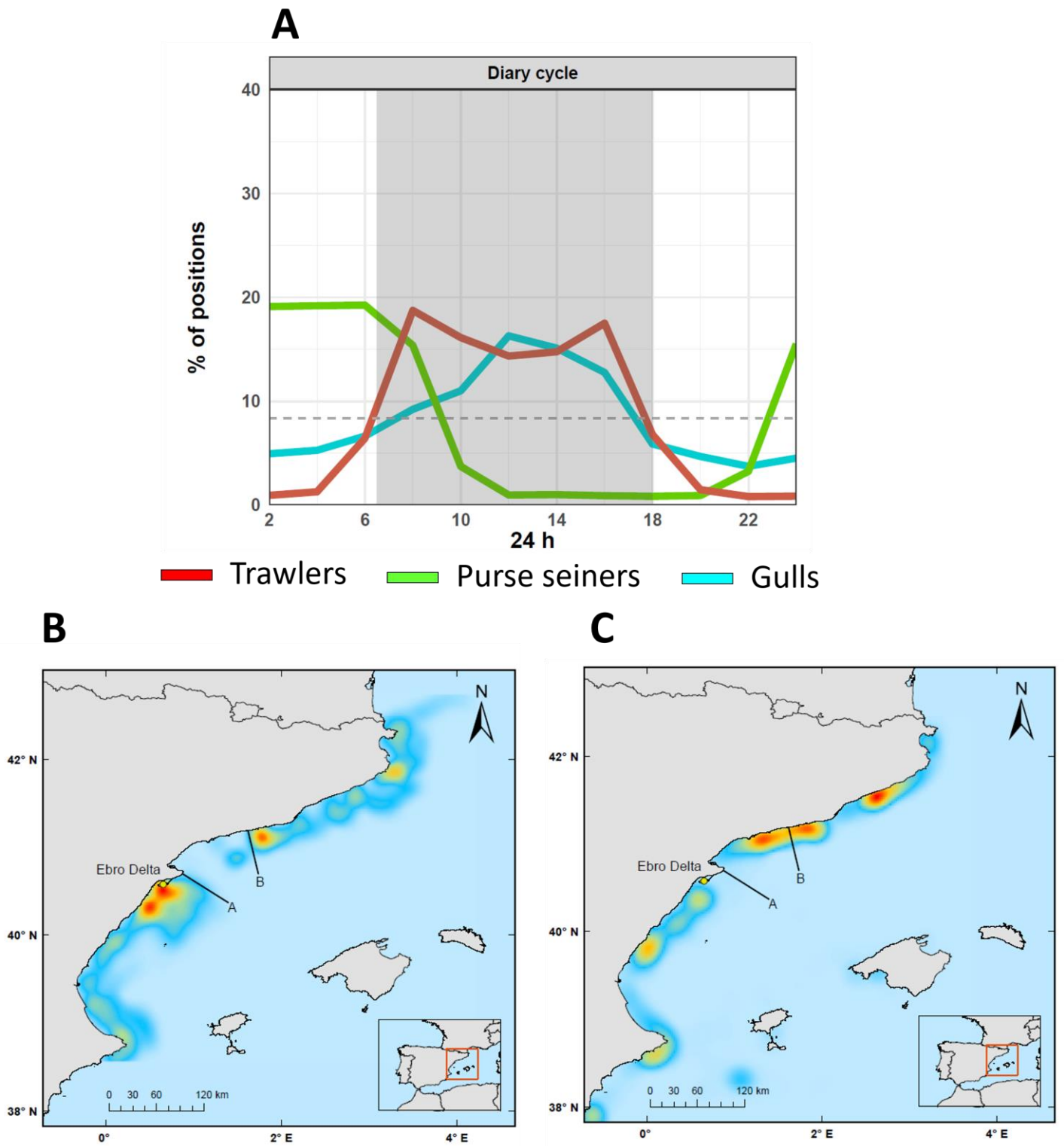


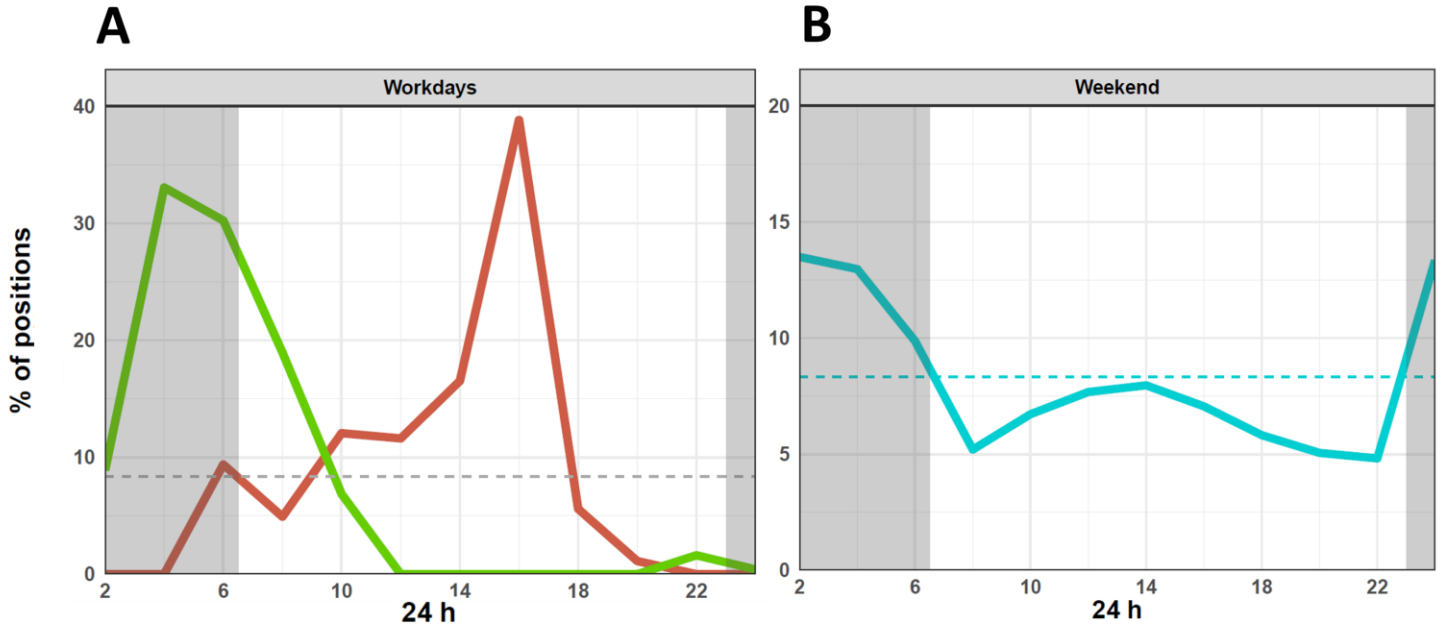
Figure 3

■ Interactions with trawlers

■ Gull's behaviour on weekends

■ Interactions with purse seiners

■ Average number of individuals at sea



C

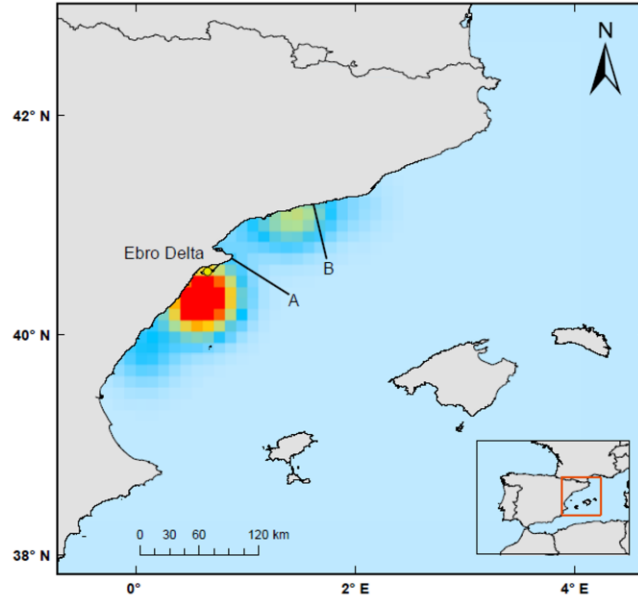


Figure 4.

