| 1 | Do global distribution patterns provide evidence of niche-shift by the introduced | | | |
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| 2 | African dung beetle Digitonthophagus gazella (Coleoptera: Scarabaeidae)? | | | |
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| 4 | Running title: Introduced African dung beetle niche-shift | | | |
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43 Abstract

The introduction of cattle into different regions of the world has led to dung beetle 44 45 relocation programs to reduce pasture fouling and to control dung-breeding flies. One of 46 these beetles, Digitonthophagus gazella (Fabricius, 1787) (Coleoptera: Scarabaeidae), is 47 native to southeast Africa and has been introduced into the Americas, Australia, and 48 New Zealand. Distribution records for this species have been used by several authors to 49 develop climate models of potential future establishment. Recent studies, however, 50 identify D. gazella as a complex comprised of seven species. Taking into account this 51 revision, we developed climate models to identify factors contributing to the ability of 52 this species to establish in many different regions. We compared the climatic conditions 53 of D. gazella in its origin with climatic conditions where the species has or has not 54 established. In Africa, Central, and South America, our results indicate that D. gazella is 55 absent in regions where it could potentially establish. We speculate that its African 56 distribution is limited due to competitive exclusion. The introduction of D. gazella in 57 Central and South America is relatively recent, such that the full extent of its 58 distribution may yet to be realized. In Australia and North America, D. gazella is 59 present in regions not predicted by the native climatic conditions. This discrepancy may 60 reflect a lack of competitive exclusion, phenotypic plasticity, and (or) genetic adaptation. 61 Our analyses suggest that the species has the ability to adapt to new conditions and a 62 wide range of climatic regions that are extremely different from those in the native region. 63 64

65 Key words: climate modelling, colonization, environmental niche, invasive species,

66 natural range, Onthophagini, Scarabaeinae, spatial distribution

67

68 Introduction

69 What are the mechanisms that allow invasive species to expand their distributions into 70 new geographic regions? This has been a persistent and challenging question in 71 biogeographical studies (Crisp & Cook, 2012; Boucher et al., 2014; Pyron et al., 2015) 72 with important implications for the spread (Broennimann et al., 2007; Petitpierre et al., 73 2012; Hill et al., 2017) and management (Hugall et al., 2002) of invasive species. Many species expand into new geographic regions when given the opportunity, often the result 74 75 of human-aided transport (e.g., emerald ash borer Agrilus planipennis Fairmaire, 1888; 76 spotted vinegar fly Drosophila suzukii (Matsumura, 1931); red fire ant Solenopsis 77 invicta Buren, 1972) or, increasingly, of climate change (Stachowicz et al., 2002; 78 McGeoch et al., 2010; Turbelin et al., 2016). Such examples do not require a change in 79 the species' fundamental niches, but rather reflect their establishment into regions for 80 which they are potentially pre-adapted (Simberloff, 1989). Thus, the native range of a 81 species has often been used as a proxy to predict regions of potential future 82 establishment (Broennimann et al., 2007; Boucher et al., 2014). However, the 83 fundamental niche of a species may also allow it to establish in new locations 84 characterized by abiotic or biotic conditions absent in the native range (Prentis et al., 85 2008; Wilson et al., 2009; Blackburn et al., 2011; Keller et al., 2011; Sultan et al., 86 2013). In such cases, predicting distribution solely on a species' native range can 87 underestimate potential future areas of establishment (Urban et al., 2007; Beaumont et 88 al., 2009; Parker et al. 2013). In addition, the ability of a species to expand into new 89 regions can be aided by genetic change, that may alter the boundaries of its fundamental niche to allow it to shift its realized niche (Jezkova & Wiens, 2016; Chapman et al., 90 91 2017). Examples of genetic changes include crop cultivars bred to withstand a wider 92 range of drought, temperature, and saline conditions (Matsui & Omasa, 2002; Olesen et 93 al., 2011). Thus, predictions of where species may establish outside their native ranges 94 are confounded by the extent of knowledge of its native range, its unsuspected 95 physiological attributes and phenotypic plasticity, and potential genetic changes.

Digitonthophagus gazella (Fabricius, 1787) is an African dung beetle species
 (Coleoptera: Scarabaeidae: Scarabaeinae: Onthophagini) that is now widely distributed
 in North America, Central and South America, Australia and the southwest Pacific, with

99 records from Japan (see Table 1). Its range expansion has been aided by redistribution 100 programs established to accelerate the degradation of cattle dung on pastures. Adult 101 beetles arrive at fresh dung pats to remove and bury dung in tunnels beneath the pat. 102 This behaviour removes dung pats as breeding sites for parasites and arthropods 103 affecting livestock (Blume & Aga, 1978; Fincher et al., 1983; Doube et al., 1991). 104 Additionally, buried dung restores soil nutrients and the tunnels increase soil 105 bioturbation, soil aeration, and water percolation (Nichols et al., 2008). 106 Digitonthophagus gazella displays broad tolerance to different ecosystems and regions, 107 particularly non-forested cattle pastures in tropical, subtropical and semi-arid regions 108 (Montes de Oca & Halffter, 1998). It also has a high dispersal rate (Seymour, 1980; 109 Barbero & López-Guerrero, 1992; Kohlmann, 1994; Álvarez et al., 2009; Noriega et al., 110 2011) and a relatively short generation time (Lee & Peng, 1981; 1982; Floate et al., 111 2015). Because of these attributes, D. gazella has become a dominant member of the dung beetle community in many locations (Doube et al., 1991; Lobo & Montes de Oca, 112 113 1997; Aidar et al., 2000; Marchiori et al., 2003; Koller et al., 2007; Matavelli & 114 Louzada, 2008). It has also become dominant in dung beetle literature. Génier & Davis 115 (2017) report the existence of at least 1120 scientific publications that reference the

116 study of this dung beetle species.

117 Models that predict regions of potential establishment for D. gazella are of 118 interest for two reasons. First, they are useful as a tool to aid future programs of 119 intentional release. Second, they can be used to identify regions where native dung 120 beetle faunas may be at risk by possible competitive interactions with D. gazella. 121 Duncan et al. (2009) used climatic parameters from its native range to predict the 122 occurrence of D. gazella in Australia at sites where the species is known to occur. 123 However, their model indicated discrepancies between the species' predicted and actual 124 Australian distributions. Medina (2016) compared climatic parameters from D. gazella's native range with that of its global invaded range. His results showed a high 125 126 degree of overlap between the two sets of parameters, suggesting that the climatic niche 127 of the species is stable between native and invaded regions. Floate et al. (2017) used 128 bioclimatic models to predict the potential distribution of D. gazella and two other 129 Onthophagini species in North America. Results for D. gazella indicated that it has 130 likely already reached the North American limits of its potential distribution.

131 Models previously developed to predict the distribution of *D. gazella* may be 132 compromised by new taxonomic insights. In their recent revision of the genus, Génier & 133 Moretto (2017) identified D. gazella as a complex of cryptic species with Indo-African 134 distribution. The species now specifically recognized as D. gazella has a restricted 135 native range from central to southern Africa and it is the only member of the species 136 complex that has been widely introduced outside of Africa (Génier & Moretto, 2017). 137 Thus, one of the main results of this revision is that the native range of D. gazella is 138 actually more restricted than previously thought. In addition, the name D. gazella has 139 been misapplied and is a synonym of a strictly Indomalayan species, D. catta 140 (Fabricius, 1787). To preserve current usage and stability in nomenclature an 141 application is currently under consideration by the International Commission of 142 Zoological Nomenclature requesting that the name D. gazella be retained for the widely 143 distributed African species (Génier & Krell, 2017: Case 3722). Our assumption in this 144 work is that the application will be successful.

145 In the current study, we revisited previous models that predict sites of potential 146 establishment for *D. gazella* based on climatic parameters from its native range. 147 Whereas these previous models used native site localities for *D. gazella* and other 148 members of the species complex, we used only the native site localities for D. gazella as 149 reported in Génier & Moretto (2017). Specifically, we asked whether D. gazella has 150 successfully invaded regions where climate models predict establishment to be unlikely. 151 A negative answer would indicate that the modelled realized niche of D. gazella is the 152 same in both its native and expanded range as concluded by Medina (2016). A positive 153 answer would suggest a shift in the realized niche with the establishment of the species 154 in regions outside its native range. If the latter is the case, then future research would be 155 warranted to verify and identify underlying mechanisms enabling this niche shift.

156

157 Materials and methods

158 Species data

159 We compiled a database of 2582 unique locality records for *D. gazella* in both its native

160 and invaded ranges, which represent a larger data set for this species compared to

- 161 previous studies (see Appendix S1). Records in this database were obtained from Génier
- 162 & Moretto (2017) and reviewed by François Génier, who checked male genitalia of at

- 163 least one specimen from each collecting event from 27 countries around the world (for 164 more details see Génier & Moretto, 2017). This examination determined that the species 165 redistributed outside Africa is D. gazella and does not include any other members of the 166 D. gazella species complex (Génier & Davis, 2017). Additional records for D. gazella 167 were obtained from the scientific literature, unpublished records provided by colleagues, 168 and from specimen labels in private and public collections. Each locality was associated 169 with its longitude and latitude (in decimal degrees) following the WGS 84 coordinates 170 reference system. The database included locality information for specimens collected in 171 51 countries across six big regions around the world (Africa -19.6%, Asia -0.3%, 172 North America – 26.8%, Central America and the Caribbean – 3.3%, South America –
- 173 21.9%, and Oceania 28.1%).
- 174

175 Environmental predictors

176 A total of 27 environmental variables at a resolution of 2.5 minutes (approximately 25 177 km²) were used as predictors: five topographic (elevation, elevation diversity, slope, 178 slope diversity, and aspect diversity), one edaphic (soil or edaphic diversity), and 21 179 climatic (see Appendix S2). Elevation was calculated from a digital elevation model 180 (DEM) downloaded from the USGS EROS Data Center (www.eros.usgs.gov). This 181 DEM was used to calculate both a slope and an aspect map, and to estimate the diversity 182 of slopes and aspects in a 7 x 7 pixel window using the classic Shannon diversity index. 183 In the same manner as that for topographic variables, we used the world map of soil 184 orders from the U.S. Natural Resource Conservation Service (www.nrcs.usda.gov) to 185 calculate the diversity of soils. Finally, the 19 bioclimatic variables of WorldClim were 186 also used (see www.worldclim.org; Hijmans et al., 2005), adding aridity and 187 continentality as calculated by Valencia-Barrera et al. (2002).

188

189 Modelling procedure

- 190 Georeferenced localities in the native range of *D. gazella* were used at a resolution of
- 191 2.5 minutes to identify regions of the globe with abiotic conditions similar to those in
- 192 the native range or potential distribution (*i.e.*, a provisional image of the inhabited
- 193 localities when the set of contingent factors preventing the colonization of suitable
- 194 regions have not affected the distribution; see Lobo, 2016). To identify regions of
- 195 potential *D. gazella* establishment outside the native range, we accounted for three main

196 disadvantages of the procedures frequently used to estimate these distributions: i) the 197 lack of reliable absence data, ii) the use of complex modelling techniques, and iii) the 198 arbitrary selection of a geographical extent. We excluded the use of pseudo-absences 199 selected at random from the studied area that are generally used when reliable absence 200 data is lacking (Hastie & Fithian, 2013; Iturbide et al., 2018). We also establish an 201 appropriate geographic area of analysis, which is fundamental to correctly selecting the 202 appropriate predictors of distribution (VanDerWal et al., 2009; Barve et al., 2011; 203 Acevedo et al., 2012), in order to generate reliable models and assessments of their 204 accuracy (Lobo et al., 2008; Hijmans, 2012; Somodi et al., 2017).

205 Thus, assuming that populations in the native range maintain a certain degree of 206 genetic connectivity (*i.e.*, geographic barriers do not isolate populations) we firstly 207 delimited the continuous distributional extent or geographical background (GB) of D. 208 gazella in its native range (i.e., the area colonized by the species in which the different 209 populations are probably interconnected). Doing so minimized the effect of contingent 210 events and factors limiting the dispersal of the species (Acevedo et al., 2012) because 211 the area used to estimate the most explanatory environmental conditions is assumed to 212 be one that is accessible to the species. Given that drainage basins constitute natural 213 landform units, the GB area was delimited as the one composed by the river basins with 214 presence observations in the native area that, in turn, enables the connection of all the 215 available occurrences (all the selected basins must be connected). The watershed 216 information provided by the WaterBase project (www.waterbase.org) was used for this 217 purpose, which includes a hierarchical coding system to recognize river basins of 218 different levels. This information is managed by the ModestR software (García-Roselló 219 et al., 2013; 2014) in order to select the minimum level of river basins with occurrences 220 that generate a contiguous and connected area. Subsequently, the most relevant 221 predictor variables were selected within this GB area to enhance the detection of 222 preferred environmental conditions within the inhabited basins.

To identify variables with the highest predictive value, we firstly excluded those with a Variance Inflation Factor (VIF) lower than five. In statistics, VIF quantifies the multicollinearity of predictors. This step eliminates twelve variables from further consideration because they are correlated with the other variables (see Appendix S2). The remaining 15 variables were then screened to identify those with the highest capacity to discriminate between the environmental conditions in the presence cells

229 against those prevailing in the GB area. An Instability Index, included within the 230 ModestR software, was estimated by dividing each continuous predictor into bins and 231 compares the relative frequency of the cells with presence data against those of GB for 232 each bin (Guisande et al., 2017). The explanatory variables showing greatest percentage 233 of contribution to the Instability Index are assumed to be those that have the highest 234 capacity to discriminate the cells of occurrence in the selected region (Guisande et al., 235 2006, 2017). As a result of this screening, six additional variables were eliminated from 236 consideration (see Appendix S2). The remaining nine variables were retained: (i) 237 precipitation of the wettest month, (ii) precipitation of the warmest quarter, (iii) 238 precipitation of the driest month, (iv) precipitation seasonality, (v) continentality (i.e., 239 the degree to which the climate of a region exemplifies that of the interior of a 240 landmass), (vi) mean diurnal range of temperature, (vii) isothermality, (viii) mean 241 temperature of the wettest quarter, and (ix) edaphic diversity. The so selected 242 explanatory variables are not intended to be those that are causally and micro-243 environmentally linked to the occurrence and abundance of the studied species. These 244 true causal variables can change spatially, may be difficult to detect in absence of 245 ecophysiological knowledge, and should reflect the environmental requirements of all 246 development stages of an insect species. Given the frequent correlation between 247 different environmental variables, the followed process of selection only aims to 248 diminish the number correlated variables and reject those with a reduced capacity of 249 being explanatory for the target species at the examined resolution. Thus, it is assumed 250 that at least some of the selected predictors direct influence demographic processes that 251 are important for the species (Austin, 1980).

252 The values of the nine selected climatic predictors were used to generate a map 253 of potential distribution according to a simple multidimensional envelope or generalized 254 intersection procedure (see Jiménez-Valverde et al., 2011; Booth et al., 2014). This 255 binary suitability map was generated by estimating the maximum and minimum 256 environmental values in cells of species presence for each selected environmental 257 variable. For each variable, we assume that the species may colonize areas showing 258 conditions within and up to these extremes at which native occurrences are observed. 259 Therefore, the map would show the potential world distribution based on similar 260 environmental conditions to those existing in the observed area of occurrence (potential 261 distribution). To asses continuous suitability representation of this potential distribution,

the formerly selected environmental variables were used to calculate the scale-invariant Mahalanobis distance (MD; Farber & Kadmon, 2003), thus measuring the environmental distance between the observations of *D. gazella* in its native range and all the remaining world cells. The upper quartile MD value in the native range has been considered the cut-off value to discriminate the most suitable localities within this potential distribution.

268 The potential global distribution derived from native occurrences was considered 269 to represent a global geographical projection of the realized niche of the species, given 270 the used predictors and the available observations. Derived in such a way, the potential 271 distribution cannot be subjected to classic validation procedures because reliable 272 absences are lacking, and also because occurrence data cannot be used to validate 273 potential distributions (Lobo, 2016); reliable absences can be obtained in favourable 274 locations predicted as presences due to the role played contingent factors, biotic 275 interactions or dispersal limitations. The geographic area currently occupied by D. 276 gazella (realized distribution) can only be estimated using known observations (Figure 277 1). Thus, an α-shape procedure was applied (Pateiro-López & Rodríguez-Casal, 2011) 278 to generate non-convex polygons by incorporating discontinuities in species 279 distributions. The overlap of the so obtained potential and realized distributions allow us 280 to identify three types of areas outside the GB area: i) environmentally suitable areas 281 that have been successfully invaded (suitable invaded areas - SIAs), ii) environmentally 282 suitable areas not yet invaded (SNIAs), and iii) environmentally unsuitable areas that have been successfully invaded (UIAs). All of the analyses were conducted using 283 284 ModestR (García-Roselló et al., 2103; 2014; www.ipez.es/ModestR) and STATISTICA 285 packages (StatSoft Inc. v12.0).

286

287 Results

The potential distribution derived from the multidimensional enveloping procedure (Figure 2) suggests that several regions of the Americas, Africa, and Australia have environmental conditions similar to those existing in the native range of *D. gazella*. Suitable conditions also occur in Madagascar and in north of its current reported distribution in Africa, mostly avoiding arid and tropical areas. Environmentally

- favourable conditions are predicted to occur in a sizeable part of South and Central
- 294 America, northern and eastern Australia, and even southern India and Indochina. In total,

- the potential distribution of *D. gazella* could encompass around 27.8 million km². Of 295 296 this area, approximately 20.2 million km^2 (72.7%) still remains un-colonized, but this 297 percentage corresponds mainly to a priori suitable areas in Africa and South America. 298 About half (47.6%) of the presence cells (Table 2) appear under *a priori* 299 environmentally favourable conditions but a slightly higher number of cells (52.4%) 300 occur in unfavourable areas (n=567). Digitonthophagus gazella seems to have invaded 301 favourable areas (i.e., SIAs) with conditions falling within extremes similar to those 302 experienced in the native area (see maximum and minimum values in Table 2), but with 303 different average values. Invaded localities have slightly higher edaphic diversity and 304 isothermality values (the range of temperatures throughout a day in comparison to 305 annual variations is higher). Above all, they have (i) a higher mean diurnal range of 306 temperature, (ii) a higher mean temperature during the wettest quarter, (iii) a 307 significantly greater precipitation during the driest month, (iv) the wettest month, and (v) 308 the warmest quarter, and (vi) a lower seasonality in monthly precipitation (Table 2).
- 309 The overlap of realized and potential distributions (Figure 3) indicates that large 310 regions of North America and central Australia have been invaded without harbouring 311 the *a priori* suitable climatic conditions of the native occurrences, based on climate 312 matching from the native range (southeast Africa, around 6 million km²). The 313 worldwide invaded localities where climatic characteristics are outside the favourable 314 climatic values of the native observations (UIAs) seem to show consistently 315 significantly higher average values of continentality and precipitation during the driest 316 month (Table 2). In addition, these invaded localities indicate that the extreme 317 conditions reached by some occurrences greatly exceed those existing in the native area. 318 This occurs for all of the selected variables, but especially for precipitation-related 319 variables. Invasive D. gazella populations seem to inhabit wetter, warmer, and more 320 continental areas than populations in their native range (Table 2). When mapping the 321 particular climatic conditions of the invaded localities that are not shared with those 322 present in the native area, there is no suitable 2.5 minute cell in Africa (e.g., only in 323 some equatorial and Malagasy areas it is possible to observe rain values higher than 60 mm/m² during the driest month, and mean temperatures over 30°C during the wettest 324 325 quarter can only be experienced in the Sahel and the Horn of Africa). 326
- 327 Discussion

328 In this study we analyse the distribution and the associated probable niche shift of D. 329 gazella, an invasive species introduced into several regions around the world. We base 330 our analysis on recent clarification of its native distribution using all the available 331 information in order to evaluate the magnitude and tendency of the environmental niche 332 shift experienced by this species. Our results indicate that about half of the distribution 333 records for D. gazella are centered on regions with differing climatic conditions to its 334 native range in Africa. These results are consistent with a substantial shift in the realized 335 environmental niche of the species. As a result, D. gazella is the most widespread dung 336 beetle in tropical and subtropical pastures around the globe. The current study indicates 337 that D. gazella has successfully established itself in geographic regions that could not be 338 entirely predicted based on environmental conditions in its native range. The global 339 invasion of D. gazella seems to have occurred in regions predicted a priori to be both 340 suitable and unsuitable for establishment. A combination of processes may explain this 341 observation, (i) relaxation of competition, (ii) ecological adaptation, (iii) physiological 342 plasticity, (iv) genetic variability, (v) anthropic disturbance and introduced exotic 343 resources, and (vi) short-term evolutionary processes.

344

345 **Competitive interactions**

346 In its native range, D. gazella populations co-occur with seven closely-related species 347 that occupy the same functional guild (Génier & Moretto, 2017). These latter species 348 are absent in all of the regions into which D. gazella has been introduced (i.e., Australia, 349 North America, Central America, and South America). Evaluation of potential negative 350 interactions between D. gazella and other native dung beetles (i.e., competition, spatial 351 displacements, or local extinctions), is generally inconclusive, especially in saturated 352 dung beetle assemblages (Noriega et al., 2017). Although reports from USA, Mexico, 353 West Indies, and Brazil suggest that introduced D. gazella have caused a decline of 354 native species (Howden & Scholtz, 1986; Montes de Oca & Halffter, 1995; Young, 355 2007; Ivie & Philips, 2008; Filho et al., 2017), this is disputed by other findings from 356 the USA, Mexico, and Nicaragua (Lobo & Montes de Oca, 1994; Howden & Howden, 357 2001; Horgan, 2008; Tiberg & Floate, 2011). However, further findings from the USA 358 and Brazil report clear changes in abundance structure of assemblages (Fincher et al., 359 1986; Filho et al., 2017) or dominance of D. gazella in terms of abundance in 360 assemblages of the USA and Australia (Edwards, 2003; Kaufman & Wood, 2012).

361

362 Ecological adaptation

363 Another interesting research matter to explore is the study of potential differences 364 between native and invasive populations of D. gazella that are living outside the limits 365 of the predictive range areas. Under a global climate change scenario it is highly 366 probable that D. gazella will be favoured (in some specific areas) and will be able to 367 colonize higher elevations, northern and southern latitude areas that could be warmer in 368 the future, taking in consideration regional precipitation changes. However, additional 369 studies of the interactions between a changing climate and biological restrictions are 370 needed (Simberloff, 2000). Studies of other taxa suggest that the potential adaptation of 371 invasive species to novel and extreme conditions is strongly connected to genetic 372 plasticity (Sexton et al., 2002; Geng et al., 2007). Distribution programs for D. gazella 373 began in the 1960s (Bornemissza, 1970; Edwards, 2007), providing a 50 year period of 374 culturing the species. A short generation time combined with multiple generations per 375 year (Tyndale-Biscoe, 1990; Floate et al., 2015) could facilitate genetic changes 376 between native and redistributed populations. Molecular analyses seem warranted to 377 assess the potential genetic differences between local ecotypes, regional, and 378 intercontinental populations (the different climate strains referred to by Edwards, 2007). 379 Further molecular studies (*i.e.*, epigenetic analyses of recognized heat/cold tolerance 380 genes) are needed to compare genetic differences between native and worldwide-381 introduced populations and identify probable genetic changes experienced by invaders. 382 This suggestion receives some support from Whipple et al. (2012) who report high 383 genetic diversity between South African and Puerto Rican populations of D. gazella 384 with genetic differentiation best explained by geographic isolation in the absence of 385 evidence for inbreeding depression.

386

387 Potential areas of colonization by *D. gazella* around the world

Based on conditions occurring in its native range, our results suggest that vast areas of the Americas, Africa, Australia, India, and Indochina have environmentally favourable conditions for *D. gazella*, making the invasion of new areas highly probable. More than 27.8 million km² (almost three times the size of USA) could be part of the potential distribution area of the species, following a very conservative estimation. These new areas to be invaded (especially those that are different from the native conditions and 394 were not predicted to be suitable) expand the potential range conditions (showing a 395 higher continentality, a higher edaphic diversity, a higher isothermality, a highest 396 diurnal range and wettest quarter temperature, a highest precipitation of the warmest 397 quarter, of the driest, and wettest month, and a lower precipitation seasonality). This 398 invadable area could be much higher if the climatic conditions in the invaded areas now 399 detected as unsuitable are considered as favourable, thus increasing the potential ability 400 of the species to invade new regions with more extreme climatic conditions.

401 The study of the colonization limits of D. gazella is important, particularly in 402 terms of latitude, elevation, and physiological restrictions. The highest elevational 403 records are around 1500 m a.s.l. in the native range (Génier & Moreto, 2017) versus 404 1920 m a.s.l. in the Americas (Anduaga, 2004), although high elevation sites typically 405 range from 1200-1500 m a.s.l (Vidaurre et al., 2008; Noriega et al., 2010). The northern 406 latitudinal limit in North America corresponds to Missouri - USA (38° N; Floate et al., 407 2017). The southern latitude limit in South America is in Chillan – Chile (36° S). Its 408 establishment in Victoria - Australia at 36° S (Edwards, 2003) and parts of New 409 Zealand $(34^{\circ} - 47^{\circ} \text{ S})$ (Forgie et al., 2018) demonstrate the ability of *D. gazella* to 410 survive in more southerly latitudes. Floate et al. (2015) report that D. gazella does not 411 have an obligatory diapause and that the minimum temperature for egg development is 412 18°C (with an average egg-to-adult development time of 52.2 days at 22 °C, a 413 development threshold of 18°C equates to 209 degree days). Tyndale-Biscoe (1990) 414 suggests that the species overwinter in Australia in both adult and larval stages. This 415 combination of factors could limit the northernmost extent of its range. Some authors 416 think that the species has reached the maximum extent of its potential distribution in 417 North America (Floate et al., 2017). However, it is possible that the species could 418 expand its actual distribution even further to colonize new areas, including regions that 419 are outside of the actual potential distribution zone that we have established here. 420 Given its membership in a "complex of seven morphologically almost 421 indistinguishable species" (Génier & Moretto, 2017), careful examination of all 422 specimens recovered outside of Africa is encouraged to establish species identity. 423 Examination of male genitalia support the contention that the species recovered outside 424 of Africa (e.g., American and Australian specimens) is D. gazella, but we do not 425 exclude the possibility that other species in this complex may have been relocated off of

426 the continent due to many different introduction programs. For example, Maes, et al.

427 (1997) reported the presence of the species in the Natural Reserve of Bosawas (north-428 eastern of Nicaragua). However, in that paper they mention the small size of the species 429 (0.5 cm, the average size of *D. gazella* is around 1 cm) and also the collection of just 430 one specimen in an area without human and cattle activity (primary rain forest) that is 431 very unusual for this species. For these reasons, we believed that this initial report 432 represented a probable confusion with Onthophagus gazellinus Bates, 1887, a native 433 species that has a similar overall shape and colour pattern but that it is smaller than D. 434 gazella and inhabits rainforest. However, we have now the confirmation of the presence 435 of D. gazella in Nicaragua (A. Solís pers. comm.).

436

437 Environmental developers and barriers to colonization

438 The speed with which D. gazella has established in different suitable invaded areas 439 (SIA's) has been variable (Barbero & López-Guerrero, 1992; Kohlmann, 1994; Noriega 440 et al., 2011). This partially may reflect different levels of competition from native 441 species. Establishment success would be hindered in regions where the functional guild 442 of D. gazella (i.e., crepuscular-nocturnal fast tunnelers in grassland environments) in 443 cattle dung is already occupied by native species. Conversely, the absence of native 444 species in this functional guild may explain the dominance of D. gazella in Central 445 America (and possibly South America) (Horgan, 2008). Even if there are no studies that 446 evaluate the importance of biotic interactions as potential barriers (*e.g.*, competitors, 447 predators, parasitoids) to the invasion of D. gazella, we suggest that the main reason for 448 the species' absence in some potential areas of Africa is due to biotic interactions (*i.e.*, 449 resource competition) with other very similar species that would act as an ecological 450 barrier to its expansion. The existence of seven neighbouring species could prevent the 451 expansion of the species, particularly as they occur within the northern and southern 452 limits of the distribution of D. gazella (Génier & Moretto, 2017). Another possible 453 element that could affect the distribution pattern of D. gazella outside its native range is 454 reduced predation pressure. Doube (1991) suggested that the low numbers of 455 individuals of *D. gazella* in African populations in comparison to those recorded in 456 Australia (North Queensland) might be related to the frequent predation on pats by 457 termites that are common in African grasslands but lack functional equivalents in 458 Australia.

459 One of the most evident barriers to the distribution of the species in many 460 regions is vegetation cover. Digitonthophagus gazella possesses a broad distribution 461 range occupying many different ecosystems: savannas, pastures, and several different 462 altered and open landscapes, especially associated with livestock (Noriega et al., 2017). 463 In its native range, *D. gazella* strongly favour grassland (75%) over woodland (18%) 464 and shaded thickets (7%) with a bias to finer-grained rather than coarser-grained soils 465 (77% sandy clay loam: 23% deep sand, n=402; Davis 1996). It appears unable to invade 466 and colonize forest areas (Lobo & Montes de Oca, 1994; Noriega et al., 2010; Davis et 467 al., 2014; Noriega et al., 2017). However, Matavelli & Louzada (2008) report that in 468 Brazil D. gazella has been able to invade inter-Amazonian grasslands that are 469 geographically isolated from cattle pastures by an extensive barrier of rainforest. 470 Although some of these introductions could result from individuals hitchhiking rides 471 within cattle dung during cattle transportation from one pasture to another. In the case 472 of inland zones of Australia, Floate et al. (2017) suggested that irrigation programs in 473 dry regions could allow D. gazella to colonize and expand into these new areas that 474 were previously not suitable for the species. However, some of the new localities 475 outside its predicted range are almost entirely non-irrigated (SOE, 2017), so something 476 different is facilitating this expansion. Vidaurre et al. (2008) mention that it could be 477 difficult for the species to invade dry inter-Andean valleys and below Puna regions 478 because of climatic conditions. However, in Colombia the species has been recovered in 479 an inter-Andean valley (Noriega, 2016) in a very dry environment (Tatacoa desert, 30° 480 C, 60% humidity, and less than 1000 mm annually), which implies the high plasticity of 481 this species to adapt to extreme conditions.

482 It is almost certain that the success of *D. gazella* in many regions is due to the 483 conversion of forests into grasslands on which cattle have been introduced. It is to be 484 expected that within a future scenario of an increase in anthropogenic disturbance (e.g., 485 deforestation, conversion of forestland to agricultural systems, and the spread of 486 grassland areas for cattle production), the establishment and invasion of D. gazella will 487 be favoured. These new open-areas create suitable habitat conditions, having available 488 food resources that the species needs. Most of the new country and establishment 489 records of the species in Central and South America are attributed to cattle breeding 490 intensification and cattle movement between countries (Noriega, 2002; Vidaurre et al., 491 2008; Noriega et al., 2010; 2017; Pablo-Cea et al., 2017). Besides, it is very interesting

that the native range of *D. gazella* in Africa matches to the more diverse areas of the
native distribution of bovids (Deshler, 1963; Heywood, 2010). Furthermore, in some
regions of Australia the presence of *D. gazella* is used as a clear indicator of disturbance
of native woodlands (Gollan et al., 2010). In addition, Wallace & Richardson (2005)
and Cave (2005) found that *D. gazella* is able to adapt to city environments and can use
dung of domestic animals (dogs), favouring the potential expansion of the species from
rural to urban areas.

499 It is also crucial to study more about the natural history and ecology of this 500 species in order to assess differences in its reproductive rate and phenology between 501 populations in its native and expanded ranges. Given that the main reason for their 502 introduction to several regions is the "assumed" capacity of D. gazella to work as an 503 excellent dung removal recycler and also act as an efficient biological control agent for 504 flies and gastrointestinal helminths (Doube et al., 1991; Miranda et al., 2000), we know 505 very little about the ecosystem services that this species performs in invaded 506 communities. It would be very important to validate and quantify the functional role of 507 this species in the introduced regions.

508 In conclusion, if we place together all the environmental, climatic, and biotic 509 elements that we know might restrict the spread and invasion of D. gazella, the main 510 geographical barriers could be: 1) vegetation cover (secondary and primary forests; such 511 as in the entire Amazon region), 2) elevation (most of the records are below 2000 m 512 a.s.l.), 3) extremes in average annual temperatures (*i.e.*, more than 30 C°, or less than 6 C°), 4) extremes in average annual precipitation (*i.e.*, more than 4400 mm, or less than 513 514 90 mm), 5) latitude (between 40-45° N and 40-45° S), and 6) saturated assemblages with 515 similar functional species (crepuscular-nocturnal fast tunnelers). However, the 516 surprising large differences between the SIA's and the UIA's suggest that the species 517 has the ability to adapt to new conditions that are extremely different from those that 518 exist in the original African native region.

519 Finally, one evident conclusion of this study is the crucial relevance of a correct 520 taxonomical identification and the importance of using an extensive and accurate 521 locality database of the native and actual distributional ranges for the species in question. 522 Using incomplete or wrong data that do not correspond to the native range, incorrect 523 localities, or mistaken data because of a poor taxonomic classification could generate a 524 wrong estimation. For example, using the data of other species that were included 525 erroneously in the "D. gazella" complex (Génier & Moretto, 2017) could generate a lot 526 of noise. This may explain discrepancies between results of the current study and those 527 of Medina (2016). Results of our model identify a greater than expected capacity for D. 528 gazella to establish in a wide range of climatic regions. This discrepancy may reflect a 529 lack of competitive exclusion, phenotypic plasticity, genetic adaptation, and (or) 530 weaknesses in the climate model. Genetic and developmental studies comparing 531 geographically isolated populations of D. gazella will help clarify the factors underlying 532 this species' distribution. 533 534 Acknowledgements 535 We want to thank Owen Olfert, Ross Weiss, John Feehan, Xu Hao, Shinya Kawai, 536 Kazuo Yamazaki, Yuichi Oba, Stephane de Greef, and Keiju Minatani for providing 537 valuable data. The data from New Zealand are courtesy of the Dung Beetle Release 538 Strategy Group (DBRSG) and Landcare Research Manaaki Whenua. JAN was 539 supported by a Colciencias PhD scholarship. 540 541 References 542 Acevedo P, Jiménez-Valverde A, Lobo JM & Real R (2012) Delimiting the 543 geographical background in species distribution modelling. Journal of 544 Biogeography 39: 1383–1390. 545 Aidar T, Koller WW, Rodrigues SR, Correa AM, da Silva JCC, Balta OS, de Oliveira 546 JM & Oliveira JL (2000) Besouros coprófagos (Coleoptera: Scarabaeidae) 547 coletados em Aquidauana, MS, Brasil. Anais da Sociedade Entomológica do 548 Brasil 29: 817-820. 549 Álvarez MC, Damborsky MP, Bar ME & Ocampo FC (2009) Registros y distribución de la especie afroasiática Digitonthophagus gazella (Coleoptera: Scarabaeidae: 550 551 Scarabaeinae) en Argentina. Revista de la Sociedad Entomológica Argentina 68: 552 373-376. 553 Anderson JR & Loomis S (1978) Exotic dung beetles in pasture and rangeland 554 ecosystems. California Agriculture 32: 31-32. 555 Anduaga S (2004) Impact of the activity of dung beetles (Coleoptera: Scarabaeidae: 556 Scarabaeinae) inhabiting pastureland in Durango, Mexico. Environmental 557 Entomology 33: 1305–1312.

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947 Figure legends

948

- Figure 1 World occurrences of *Digitonthophagus gazella* used in this study (2.5 minute
 cells). Native occurrences are those records in mainland Africa with the exception of
- 951 those from Madagascar and the Comoros islands.

unsuitable not invaded (UNIAs without colour).

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953 Figure 2 Worldwide potential distribution of Digitonthophagus gazella using the data 954 of the selected climatic variables from its native occurrences and a multidimensional 955 enveloping procedure. The four regional maps, displayed in greater detail, show the 956 potential distribution of *D. gazella* and the observed occurrences (black points). The 957 colour gradient reflects the Mahalanobis distance (MD) from the conditions existing in 958 the native range. The upper MD quartile value of the native range represents the most 959 suitable localities within the potential distribution ($\mathbf{v} = 5.27$). Uncoloured areas lie 960 outside the potential distribution of the species. 961 962 Figure 3 Mapping of the current and potential distribution of *Digitonthophagus gazella* 963 showing the extent and location of the environmentally suitable invaded areas 964 (equivalent to the native range in mainland Africa; SIAs in blue), the suitable not yet 965 invaded areas (SNIAs in orange), unsuitable invaded areas (UIAs in green), and

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Table 1 List of introductions and/or first records of *Digitonthophagus gazella* by continent and country around the world outside its native distribution (Génier & Moretto, 2017)

| Continent | Country | Year | References (chronologically) |
|-----------|------------------------------------|-------|--|
| Asia | Japan ¹ | 1978 | Okajima & Araya, 2012 |
| Africa | Madagascar | 1885 | Génier & Moretto, 2017 |
| | France (Mayotte) | 1958 | Génier & Moretto, 2017 |
| | Comoros | 1958 | Génier & Moretto, 2017 |
| North | US (Hawaii) | 1957 | Legner, 1978; Markin & Yoshioka, 1998; Génier & Moretto, 2017 |
| America | US | 1972 | Blume & Aga, 1978; Anderson & Loomis, 1978; Fincher et al., 1983; Downie, |
| | | | 1984; Blume, 1985; Hunter & Fincher, 1985; Fincher et al., 1986; Barbero & |
| | | | López-Guerrero, 1992; Vulinec & Eudy, 1993; MacRae & Penn, 2001; Génier & |
| | | | Moretto, 2017 |
| | Mexico | 1981 | Fincher et al., 1983; Zunino & Halffter, 1988; Rivera-Cervantes & García-Real, |
| | | | 1991; Barbero & López-Guerrero, 1992; Thomas, 1993; Kohlmann, 1994; Lobo |
| | | | & Montes de Oca, 1994; 1997; Montes de Oca & Halffter, 1998; Morón et al., |
| | | | 1998; Deloya, 2000; Montes de Oca, 2001; Morales et al., 2004; Génier & |
| | | | Moretto, 2017 |
| Central | Guatemala | 1987 | Kohlmann, 1994 |
| America | El Salvador | 2016 | Pablo-Cea et al., 2017 |
| | Nicaragua | <1996 | (A. Solís pers. comm.) |
| (Greater | Cuba, Jamaica, Dominican Republic, | 1990 | Ivie & Philips, 2008; Génier & Moretto, 2017 |

| Antilles) | Puerto Rico (US) | | |
|-----------|--|-------|--|
| | St Croix (US), Anguilla (UK), St. Kitts, | 1992 | Huchet ,1992; Kohlmann, 1994; Ivie & Philips, 2008; Génier & Moretto, 2017 |
| (Lesser | Antigua, Montserrat (UK), Guadeloupe | | |
| Antilles) | and Marie-Galante (France), Martinique | | |
| | (France), St Vincent, Union, Grenada | | |
| South | Colombia | 1995 | Noriega, 2002; Noriega et al., 2006; Rivera & Wolff, 2007; Navarro et al., 2009; |
| America | | | Noriega et al., 2011; Noriega, 2016; Noriega et al., 2017 |
| | Venezuela | 1996 | Gámez et al., 1997; Noriega et al., 2017 |
| | French Guiana | 2008 | Boilly & Vaz-de-Mello, 2013 |
| | Brazil | ~1990 | Nacimiento et al., 1990; Bianchin et al., 1998; Koller et al., 1999; Aidar et al., |
| | | | 2000; Miranda et al., 2000; Marchiori, de Oliveira & Linhares, 2001; 2003; |
| | | | Schiffler et al., 2003; Scheffler, 2005; Koller et al., 2007; Matavelli & Louzada, |
| | | | 2008; Génier & Moretto, 2017 |
| | Peru | 1999 | Noriega et al., 2010 |
| | Bolivia | 1990 | Vidaurre et al., 2008 |
| | Chile (Easter Island) | 1988 | Ripa & Rodríguez, 1990; Ripa et al., 1995 |
| | Chile | 1988 | Ripa & Rodríguez, 1990; Ripa et al., 1995 |
| | Paraguay | 1998 | Ruiz, 2000 |
| | Argentina | 2006 | Ocampo & Ruiz-Manzanos, 2008; Álvarez et al., 2009 |
| Oceania | Australia | 1967 | Waterhouse, 1974; Génier & Moretto, 2017; Reid et al., 2018 |
| | Papua New Guinea | 1973 | Génier & Moretto, 2017 |
| | | | |

| New Zealand | 2013 | ERMA, 2011; Forgie et al., 2018 |
|---------------|------|--|
| New Caledonia | 1978 | Gutiérrez et al., 1988; Génier & Moretto, 2017 |
| Vanuatu | 1978 | Gutiérrez et al., 1988; Génier & Moretto, 2017 |
| Fiji | 1978 | Génier & Moretto, 2017 |

¹ Any specimen from this locality (Okinawa island, Japan) has been studied by F. Génier or by any of the co-authors. Further studies need to check if this record belongs to *D. gazella*.

Table 2 Mean values of the selected climatic variables (\pm 95% confidence intervals) for the 2.5-minute cells with occurrence observations of *Digitonthophagus gazella* in native and invaded areas. Within the invaded area, SIAs are climatically suitable and UIAs are climatically unsuitable according to conditions prevailing in the native distribution areas. The last two numbers separated by a hyphen represent minimum and maximum climatic values. Temperatures are in decimal degrees and precipitation in mm

| Climatic variables | Native | Invaded | |
|--------------------------------------|----------------------|---------------------------|---------------------------|
| | | SIAs (suitable) | UIAs (unsuitable) |
| Number of cells | 292 | 516 (47.64%) | 567 (52.35%) |
| Continentality | 9.05±0.65, 0-28 | 9.23±0.62, 0-28 | 22.93±1.32, 0-66 |
| Edaphic diversity | 0.57±0.05, 0-1.73 | $0.74 \pm 0.04, 0 - 1.70$ | $0.62 \pm 0.04, 0 - 1.97$ |
| Isothermallity | 58.55±0.65, 49-85 | 61.72±0.79, 50-84 | 51.83±0.92, 26-92 |
| Mean diurnal temperature range | 134.78±2.79, 62-172 | 118.74±1.87, 63-172 | 125.16±2.20, 61-190 |
| Mean temperature wettest quarter | 234.16±2.90, 96-286 | 254.79±2.04, 96-284 | 248.15±4.25, 33-327 |
| Precipitation of the driest month | 7.39±1.11, 0-59 | 17.42±1.28, 0-59 | 29.75±2.71, 0-244 |
| Precipitation seasonality | 83.59±2.33, 16-129 | 73.91±2.30, 27-126 | 62.64±2.75, 8-136 |
| Precipitation of the warmest quarter | 292.47±12.38, 25-884 | 407.03±13.79, 71-807 | 359.23±24.04, 3-1403 |
| Precipitation of the wettest month | 158.64±7.70, 39-465 | 218.74±8.61, 47-458 | 182.23±10.96, 17-648 |



Figure 1 World occurrences of *Digitonthophagus gazella* used in this study (2.5 minute cells). Native occurrences are those records in mainland Africa with the exception of those from Madagascar and the Comoros islands.



Figure 2 Worldwide potential distribution of *Digitonthophagus gazella* using the data of the selected climatic variables from its native occurrences and a multidimensional enveloping procedure. The four regional maps, displayed in greater detail, show the potential distribution of *D. gazella* and the observed occurrences (black points). The colour gradient reflects the Mahalanobis distance (MD) from the conditions existing in the native range. The upper MD quartile value of the native range represents the most suitable localities within the potential distribution ($\checkmark = 5.27$). Uncoloured areas lie outside the potential distribution of the species.



Figure 3 Mapping of the current and potential distribution of *Digitonthophagus gazella* showing the extent and location of the environmentally suitable invaded areas (equivalent to the native range in mainland Africa; SIAs in blue), the suitable not yet invaded areas (SNIAs in orange), unsuitable invaded areas (UIAs in green), and unsuitable not invaded (UNIAs without colour).