

1 **Special issue: Mixed species forests. Risks, resilience and management**

2 **Mixing has limited impacts on the foliar nutrition of European beech and Scots pine trees**

3 **across Europe**

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24

25 **Abstract**

26 Tree species-mixing has been suggested as one option to counteract the adverse effects of global change
27 on tree mineral nutrition, yet the effect of mixing on nutrient availability remains poorly documented.

28 We therefore analyzed the current foliar nutrient (N, P, K, Ca, Mg) quantities and *ilr* balances (isometric
29 log transformed ratios between elements or groups of elements) for 261 European beech and 248 Scots
30 pine trees from 15 sites, each consisting of one beech-pine mixed stand and the respective monocultures,
31 across a gradient of environmental conditions in Europe. We hypothesized an overall positive effect of
32 mixing on tree foliar nutrient content, and that this mixing effect would be stronger on nutrient-poor sites.
33 Using linear mixed models and multivariate linear regression models, we first tested for the effects of
34 species (beech/pine) and composition (pure/mixed) across all sites; we then investigated whether the
35 species-mixing effect was related to site fertility.

36 The nutrient composition of beech leaves and pine needles differed significantly for all *ilr* balances. For
37 both species, significant mixing effects were detected for some nutrients and *ilr* balances; those effects,
38 however, could not be consistently related to contrasted nutrient composition between species. For most
39 nutrients and *ilr* balances, the mixing effect was influenced by the site nutritional status, but the pattern
40 differed from expectation: absence or minor differences between monocultures and mixtures at the lower
41 end of the chemical fertility gradient, and maximum differences in rich soils.

42 The contrasting foliar nutrient composition of pine and beech trees and the site nutrient status only partly
43 explained the mixing effects on tree mineral nutrition. Our results claim for a better understanding of
44 nutrient-related mechanisms associated with complementarity and points towards the need to further
45 expand the existing frameworks to account for the multivariate nature of tree nutrition.

46

47 **Keywords** Foliar nutrition – Complementarity – Species mixture – *Fagus sylvatica* L. – *Pinus*
48 *sylvestris* L.

49

50 **1. Introduction**

51 Many recent studies underline that tree mineral nutrition is deteriorating in Europe under global
52 environmental changes (Jonard et al. 2015, Peñuelas et al. 2013). For instance, negative trend in foliar P
53 concentration were found for *Fagus sylvatica* (L.) and *Pinus sylvestris* (L.) along with negative Ca and Mg
54 trends in the case of the first species and negative S trend for the latter one (Jonard et al. 2015). While
55 concentrations of some elements are declining, N foliar concentrations tend to increase for several species
56 (Jonard et al. 2015). Such opposite trends induce foliar imbalances which, in turn, may impact the ability
57 of forests to provide ecosystem services such as biomass production (Oren and Schulze 1989). In that
58 context, species-mixing could be an efficient management tool to improve stand nutrition through the
59 potential positive effects of interspecies interactions on the availability, uptake or use efficiency of
60 nutrients (Ammer 2019, Forrester and Bauhus 2016, Nickmans et al. 2015, Rothe and Binkley 2001).

61 Nutrient *availability* is influenced by many processes. Richards et al. (2010) reported several ways by which
62 tree species mixture may improve nutrient supply, including enhanced mineralisation, reduced loss of
63 nutrient through leaching and erosion, increased rate of N₂ fixation or weathering, and increased quality,
64 quantity and decomposition rates of the leaf litter. Species-mixing effects on nutrient *uptake* are often
65 associated with differences in physiological, morphological or phenological characteristics between the
66 tree species involved in the mixture (Forrester and Bauhus 2016). An example is fine root overyielding in
67 tree mixtures resulting from belowground niche differentiation for species with contrasting rooting traits

68 (Leuschner et al. 2001). Through improvement of soil resources, species-mixing could also induce a
69 reallocation of carbon to aboveground biomass, resulting in a higher *nutrient use efficiency* compared to
70 pure stands (Epron et al. 2013, Forrester et al. 2006). All the processes listed above have to be seen in the
71 light of the biogeochemical niche hypothesis (Urbina et al. 2017). The hypothesis states that, because of
72 differences in growth and nutrient use strategies, co-occurring species use mineral elements in different
73 proportions, which leads to species-specific stoichiometry and associated stoichiometric flexibility. Such
74 differences in biogeochemical niches between coexisting species allows for reduced competition and
75 nutrient use optimization at the community level (Urbina et al. 2017).

76 It is frequently thought that admixing tree species has beneficial impacts on tree nutrition. In temperate
77 forests, those effects have mostly been studied for mixed stands of coniferous and broadleaved trees
78 (Brown 1992, Rothe and Binkley 2001, Thelin et al. 2002). Such expected impacts result from differences
79 in biogeochemical cycle, nutrient demand and nutrient use between conifers and broadleaves. For
80 instance, nitrogen mineralization and nitrification fluxes are generally lower in coniferous than in
81 broadleaves stands, an effect commonly attributed to more acidic conditions, higher C:N ratio of organic
82 matter, and higher amounts of inhibiting and/or recalcitrant compounds under conifers (Augusto et al.
83 2015). The same mechanisms could also hold true for other nutrients such as P and Ca because their
84 availability is partly related to organic matter mineralization. Additional processes involve distinct impacts
85 of coniferous vs. broadleaved trees on atmospheric deposition and mineral weathering (Augusto et al.
86 2015). Differences in nutrient demand and nutrient use between coniferous and broadleaves species also
87 leave room for beneficial impacts of admixing these species. Such differences include a generally higher
88 nutrient concentration in litter and fine roots for broadleaves, and longer nutrient residence time for
89 conifers (Augusto et al 2015).

90 The impact of admixing coniferous and broadleaf species on tree nutrition has been highlighted in previous
91 studies. For instance, Brandtberg (2001) found that admixing birch species (*Betula pendula* Roth and

92 *Betula pubescens* Ehrh.) and Norway spruce (*Picea abies* (L.) Karst) increased P and K concentrations of
93 spruce needles (but did not influence N, Ca, Mg or Mn concentrations). Thelin et al. (2002) reported higher
94 P and K concentration in Norway spruce needles when mixed with beech, birch or oak (*Quercus robur* L.
95 and *Quercus petraea* (Matt.) Liebl.). Brown (1992) found that interspecific differences in tissue
96 concentration of N and possibly P, were key factors responsible for contrasting types of mixture effects
97 (positive, compensatory and negative) on Norway spruce when mixed with Scots pine, black alder and
98 sessile oak, respectively. However, other studies failed to find any significant differences in foliar nutrient
99 concentrations between pure and mixed stands of conifers and broadleaved species (Heinsdorf 1997,
100 Magh et al. 2018, Neft and Stangl 1985, Rothe et al. 2003). An explanation for those inconsistent results is
101 the dependency of species-mixing effects to environmental conditions (Ratcliffe et al. 2017). Building on
102 the framework developed by Forrester and Bauhus (2016), effects of mixing on nutrition are expected to
103 increase along a gradient of decreasing nutrient availability provided that species interactions improve the
104 availability, uptake or resource use efficiency of limiting nutrients. In accordance with this pattern, Magh
105 et al. (2018) found that in the absence of any water limitation, there was a facilitative effect of silver-fir on
106 N nutrition of beech in N-limited sites, but an antagonistic effect in N-rich soils. However such a simple
107 framework might not be sufficient to explain the full range of patterns for nutrient related mixing effects
108 along environmental gradients. First, it does not account for non linearity of the relationship between
109 resource availability and mixing effects. For instance, while their focus was on the mixing effect on drought
110 exposure, de Streel et al. (2019) highlighted tipping points along a nearly similar environmental gradient.
111 Second, the above framework does not take into account simultaneous limiting factors. Sardans and
112 Peñuelas (2007), however, showed that species-mixing effects on foliar nutrition could be influenced by
113 site water status through its impact on nutrient uptake.

114 The assessment of tree nutrition is largely based on analyzing single foliar nutrient concentrations or
115 contents (i.e. amount of a nutrient in a given number of leaves/needles), as well as ratios between pairs

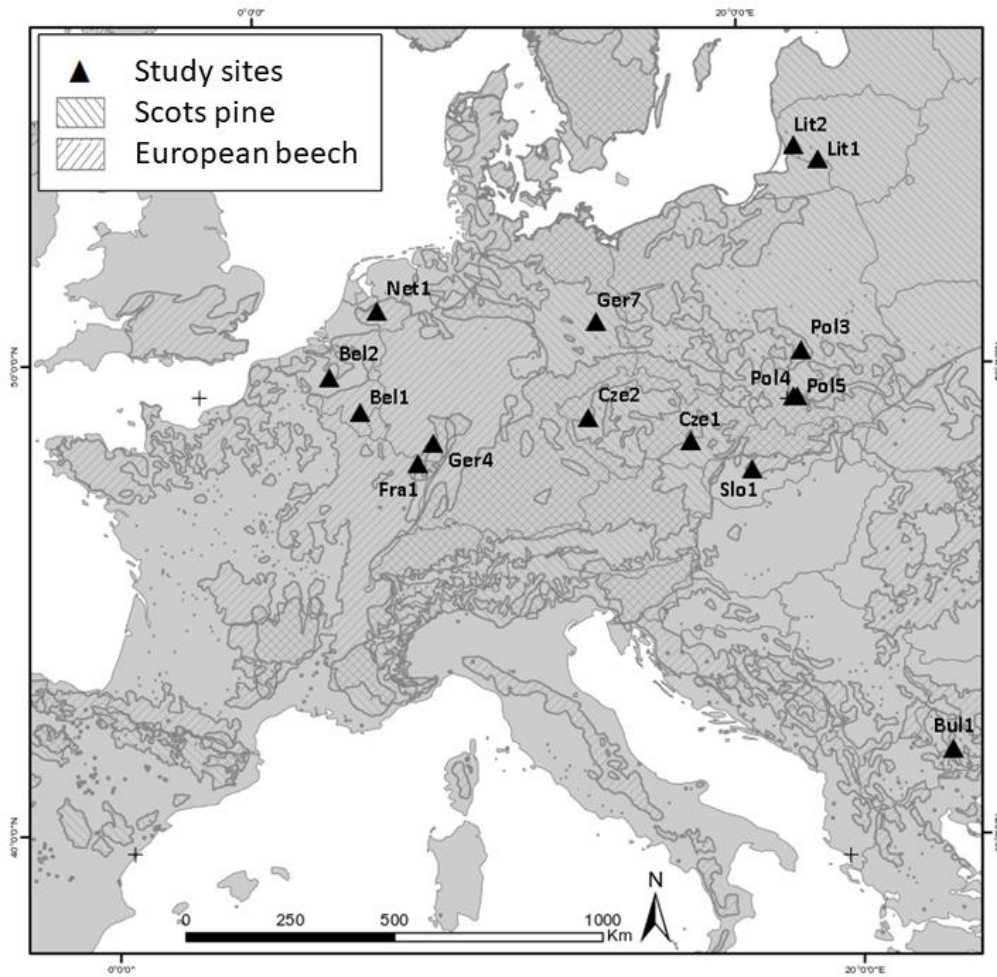
116 of elements (e.g. Jonard et al. 2015). This approach yields valuable information about mineral elements
117 that are essential to plant functioning and to their metabolism, and the reported values can easily be
118 compared with thresholds (Mellert and Göttlein 2012). More recently, a complementary method, where
119 nutrients are subjected to isometric log-ratio (*ilr*) transformation (see Materials and methods below), has
120 been proposed to better account for the interactions between nutrients while limiting biases related to
121 using raw compositional data (Egozcue et al. 2003). Since its development, this method was successfully
122 applied in different contexts. For instance, Parent et al. (2013b) used it to analyse the mineral plasticity of
123 cloudberry and Parent et al. (2013c), Modesto et al. (2014) and Hájek et al. (2014) used it to study the
124 nutrition of Mango trees, maize or mosses, respectively. Recently, Collin et al. (2016) used the *ilr* method
125 to study the nutritional characteristics of Sugar Maple (*Acer saccharum* Marsh.) and Red Maple (*Acer*
126 *rubrum* L.) under varying proportions of conifers, while Nowaki et al. (2017) used this approach to assess
127 the response of irrigated tomato crops to P fertilization.

128 The objective of this paper was to assess the impact of mixing broadleaves and conifers on tree nutrition.
129 Focusing on the widespread European beech - Scots pine mixture, we sampled the foliage of 261 beech
130 (*Fagus sylvatica* L.) and 248 pine (*Pinus sylvestris* L.) trees at 15 sites distributed over Europe, each of which
131 contained a triplet of plots including a mixture and the monocultures of each species. We used the *ilr*
132 approach to obtain a synthetic and non redundant characterization of the overall nutritional signature, in
133 complement to the classical nutrient content/concentration approach.

134 We hypothesized that differences in foliar nutritional characteristics between species would result in
135 significant mixing effects on tree nutrition. We further hypothesized that mixing effects would be
136 dependent on site nutritional status.

137 **2. Materials and methods**

138 **2.1. Study area and site/stand characteristics**



139
140 *Figure 1 Distribution of the 15 EuMIXFOR beech-pine triplets used in the present study and distribution of European beech and*
141 *Scots pine according to EUFORGEN (www.euforgen.org).*

142 This study is based on data from 15 sites each consisting of pure stands of pine and beech and a mixed
143 stand of both species (Figure 1), with all three of them located in similar conditions. The sites have been
144 established under the COST Action FP1206 EuMIXFOR (European Network on Mixed Forests). This network
145 covers a large gradient of environmental conditions and site nutrient status (Figure 2; Tables A1 and A2)
146 within the overlapping natural ranges of pine and beech. Elevations varies between 20 and 1190 m a.s.l.;

147 mean annual precipitation sum (P) ranges from 520 to 1175 mm and mean annual temperature (T) from
148 6 to 10.5 °C. The stands are mostly even-aged and mono-layered. A summary of stand characteristics is
149 provided in Table A3. In the mixtures, the percentage of basal area represented by scots pine ranged from
150 25 % to 69 %; total basal area ranged from 16 to 79 m² ha⁻¹ and stand age from 45 to 130 years. Stand size
151 ranged from 0.025 to 0.73 ha for pure beech stands, from 0.025 to 1.55 ha for pure pine stands and from
152 0.045 to 0.462 ha for mixed stands (Table A3). No silvicultural activities had been conducted in the stands
153 during the preceding decade. A standard protocol for tree data collection (diameters, heights of trees and
154 crown bases) and tree coring was applied. The full measurement protocol was described in details by Heym
155 et al. (2017).

156

157 **2.2. Leaf collection, pretreatment and analysis**

158 Across the 15 sites-network, 509 trees (261 European beech and 248 Scots pine) were sampled. In each
159 site and plot (pure beech, pure pine, mixed beech-pine), 5 to 10 trees were sampled per species (i.e. beech
160 or pine in the pure stands, beech and pine in the mixed stand) among the dominant trees during the
161 summer 2016 (second part of the vegetation period, before autumnal colouring) for a total of 24 to 40
162 sampled trees per site. Current year leaves/needles were collected from several branches located in the
163 upper third of the crown, resulting in one sample per tree. Leaves/needles were dried at 40°C until
164 constant weight was achieved for the nutrient analysis. To determine the dry mass, a subsample was
165 additionally oven-dried at 70°C (50 leaves/500 needles). Foliar N concentrations were measured using the
166 dry combustion method with a Flash Analyzer (Thermo Finnigan Flash EA 1112 elemental analyser) and
167 the other elements (P, K, Ca and Mg) were determined by ICP spectrometry (Varian 720 E-S) after digestion
168 with HNO₃ in a microwave oven (Milestone UltraWAVE, Germany). All analyses were done on a per tree
169 basis.

170

171 **2.3. Tree nutrition assessment**

172 For each tree, the foliar nutrient composition was characterized by single nutrient (N, P, K, Ca, Mg)
173 concentrations and contents, as well as by isometric log transformed ratios (*ilr*) between elements or
174 groups of elements (Table 1).

175 *Nutrient concentrations and contents.* Nutrient concentrations in current-year foliage were first used to
176 rank each site by comparison to species-specific reference thresholds (Mellert and Göttlein 2012). Though
177 such thresholds don't account for nutrient interactions (e.g. Marschner 2011), they are useful as baseline
178 values to assess the nutrient status (e.g. Jonard et al. 2015). For all subsequent analyses, we used nutrient
179 contents instead of nutrient concentrations to account for differences in leaf/needle dry mass resulting
180 from contrasting growing conditions which would impact foliar nutrient concentrations through
181 'concentration' / 'dilution' effects (Binkley and Fisher 2013, Jarrell and Beverly 1981). The content of each
182 individual nutrient in 50 leaves or 500 needles was calculated by multiplying its concentration by the
183 corresponding leaf or needle dry mass.

184 *Isometric log transformed ratios (ilr).* The chemical composition of a plant tissue is a closed system in that
185 all constituting nutrients are constrained to sum to 100%. Data corresponding to quantitative descriptions
186 of the parts of a closed system are defined as "compositional data" (Collin et al. 2016). Among others, the
187 components of such system are not independent. Indeed, if the value of one component increases, the
188 value of at least one other component has to decrease in order to keep the sum constant. Therefore,
189 specific compositional data analyses techniques such as the isometric log-ratio techniques have been
190 developed. Despite its limited use for the assessment of tree foliar nutrition, the *ilr* transformation
191 technique has been proved useful for analysing plant nutrient composition (Hájek et al. 2014, Modesto et

192 al. 2014, Parent et al. 2013b, Parent et al. 2013c). In addition to constraining the system to 100%, it also
193 accounts for the physiological interactions between nutrients (Collin et al. 2016, Nowaki et al. 2017).

194 The *ilr* transformation technique gives information about the relative amounts of elements or groups of
195 elements, allowing D-1 orthogonal (geometrically independent) balances to be produced for a D part
196 composition. D is the number of measured elements plus a filling value (Fv). The filling value corresponds
197 to the difference between the unit or scale of measurement (e.g. 100%) and the sum of all measured
198 elements (Parent et al. 2013). In this context, a balance, hereafter referred to as *ilr* balance, is defined as
199 the relationship between group of parts (groups composed of one or several nutrients) (Egozcue et al.
200 2005). A closure operation is applied to the resulting matrix of compositional data (Aitchison 1986), which
201 computes the constant sum of components as follows (Egozcue and Pawlowky-Glahn 2005):

$$202 \quad S^D = C(c_1, c_2, \dots, c_D) = \left[\frac{c_1 k}{\sum_{i=1}^D c_i}, \frac{c_2 k}{\sum_{i=1}^D c_i}, \dots, \frac{c_D k}{\sum_{i=1}^D c_i} \right] \quad (\text{Eq. 1})$$

203 Where S^D is a vector of D components adding up to a constant k (e.g. 100%), C is the closure operator, c_i is
204 the i^{th} part of a composition of D components.

205 The *ilr* transformation is then applied to the closed data. This system of lineary independent ratios is called
206 Sequential Binary Partition (SBP) and describes the D-1 orthogonal balances between nutrients or groups
207 of nutrients (Parent et al. 2013). The SBP of a D-elements composition is a (D-1) x D matrix where columns
208 correspond to the parts of the composition and rows to the *ilr* balances. In the SBP, elements labelled "+1"
209 correspond to balances numerators, elements labelled "-1" correspond to balance denominators and
210 elements labelled "0" are not part of the balance in question. This SBP is defined *a priori*, for instance,
211 based on user knowledge. In our case, the SBP (Table 1) is based on prior knowledge of plant nutrition
212 (Collin et al. 2016, Marschner 2011, Parent et al. 2013). The first partition contrasts all measured elements
213 (N, P, K, Ca, Mg) to the filling value (Fv). The second partition contrasts Mg and Ca to K, P and N ([Mg, Ca|K,
214 P,N]). The sub-compositions are then divided into [K|P, N], [P|N] and [Mg|Ca]. [P|N] is representative of

215 the Redfield ratio which is an indicator of the balance between processes associated with the synthesis of
 216 N-rich proteins and the synthesis of P-rich r-ARN respectively (Ågren 2004, Loladze and Elser 2011).
 217 [Mg|Ca] is an indicator of the geographical position and soil mineralogy (Walworth and Summer 1988).

218 *Table 1 Sequential Binary Partition (SBP) of foliar nutrients of beech and pine trees based on prior knowledge of nutrient*
 219 *interaction in higher plants and number of components in the (+) et (-) groups (n^+ and n^- , respectively).*

Balance [-1 subset +1 subset]	SBP parts						n^+	n^-
	N	P	K	Ca	Mg	Fv		
[Fv N, P, K, Ca, Mg]	1	1	1	1	1	-1	5	1
[Mg, Ca K, P, N]	1	1	1	-1	-1	0	3	2
[K P, N]	1	1	-1	0	0	0	2	1
[P N]	1	-1	0	0	0	0	1	1
[Mg Ca]	0	0	0	1	-1	0	1	1

220
 221 Once the SBP was defined, we used the following equation to calculate the *ilr* balances (Egozcue and
 222 Pawlowsky-Glahn 2005):

223
$$ilr_j = \sqrt{\frac{n_j^+ n_j^-}{n_j^+ + n_j^-}} \ln \frac{g(c_j^+)}{g(c_j^-)} \quad (\text{Eq. 2})$$

224 *ilr_j* corresponds to the j^{th} isometric log-ratio, n_j^+ and n_j^- are the number of components in the (+) et (-)
 225 groups, $g(c_j^+)$ and $g(c_j^-)$ are the geometric means of groups (+) and (-). $\sqrt{(n_j^+ n_j^-)/(n_j^+ + n_j^-)}$ corresponds
 226 to the orthogonal coefficient of the j^{th} balance defined in the SBP. *ilr* balances are conventionally presented
 227 in the form [components in denominator|components in numerator] because log ratios become more
 228 negative as the denominator increases and hence, the balance leans to the left as in algebra where

229 negative numbers are located on the left side of the zero (Nowaki et al. 2017, Parent et al. 2013b). Thus,
 230 an increase in *ilr* balance can be attributed to either a decrease in the left part of the balance or an increase
 231 in the right part of the balance. For instance, the [P|N] balance is defined by $\sqrt{\frac{1}{2}} \ln \frac{N}{P}$. As a consequence,
 232 an increase in N or a decrease in P concentration will lead to increased *ilr* balance value and the [P|N]
 233 partition leans to the right (Collin et al. 2016, Parent et al. 2013b).

234

235 **2.4. Statistical analyses**

236 To test the differences in foliar *ilr* balances, foliar nutrient concentration and foliar nutrient content
 237 between species, we conducted Student's t-Test. By informing about the biogeochemical niche of the
 238 studied species, those analyses are used to interpret the species-mixing effects.

239 In order to test whether stand composition (pure vs. mixed), species identity (beech vs. pine) and their
 240 interaction influenced the current *ilr* balance or the foliar nutrient content, we fitted linear mixed models
 241 on each *ilr* balance or nutrient content, considering site and stand nested within site as random factors

242

$$243 \quad Y_{hij_s} = \beta \times E_{hij_s} + a_s(0, \sigma_s^2) + a_{j|s}(0, \sigma_j^2) + \varepsilon(0, \sigma_\varepsilon^2) \quad (\text{Eq. 3})$$

244

245 Y_{hij_s} represents the *ilr* balance/content of interest for the i^{th} individual (tree) in the j^{th} stand (pure or mixed)
 246 and the s^{th} site for species h (pine or beech), β is the vector of the fixed effect parameters (species identity,
 247 stand composition and their interaction), E is the matrix of the predictors of the fixed effects, a_s is the
 248 random factor characterized by the inter-site variance σ_s^2 , $a_{j|s}$ is the random factor characterized by the
 249 inter-stand variance within a same site σ_j^2 and ε is the error term of variance σ_ε^2 . In addition, linear mixed
 250 effect models using site as random factor were used to test the significance of the difference between
 251 pure and mixed stands within each species for each *ilr* balance or nutrient content.

252

253 We then investigated the relationship between the species-mixing effect on tree nutrition, and the site
254 nutritional status. To do so, we computed the difference between the *ilr* balance/nutrient content per tree
255 and the corresponding average value per site, for each species separately, and used this index as the
256 response variable. Stand composition (pure vs. mixed), site nutritional status and their interaction were
257 used as explanatory variables in multivariate linear regression models (eq. 4):

258

$$259 \quad Y_{ijs} - \bar{Y}_s = a + \beta \times E_{js} + \varepsilon(0, \sigma_\varepsilon^2) \quad (\text{Eq. 4})$$

260

261 Y_{ijs} represents the *ilr* balance/content of interest for the i^{th} individual (tree) in the j^{th} stand (pure or mixed)
262 and the s^{th} site for each species separately, and \bar{Y}_s is the average value of the *ilr* balance/content of either
263 beech or pine trees across the pure and mixed stands in site s . α is the intercept, β is the vector of the
264 parameters, E is the predictor matrix (stand composition, site nutritional status and their interaction) and
265 ε is the error term of variance σ_ε^2 . Considering that the *ilr* balance [Fv | Mg, Ca, K, P, N] contrasts five mineral
266 elements of major importance for tree growth and functioning against all other elements, we used its
267 average value per site and species as an index of site nutritional status for all *ilr* balance models. For the
268 foliar nutrient content models, we used the average value of the corresponding nutrient content per site
269 and species as the index of site nutritional status. In all cases, the variables used as indicators of site
270 nutritional status were centered. The contrasting average values of either *ilr* balances or nutrient contents
271 between sites resulted in empirical gradients of nutritional status.

272

273 Dependent variables of the mixed and multivariate linear models were normally distributed with the
274 exception of the nutrient content used in the mixed models which were therefore ln-transformed.

275 All statistical analyses were conducted using the R software, version 3.4.1 (R Core Team, 2019).

276 Compositional data analyses were conducted using the “composition” package (Van den Boogaart et al.
277 2019). The closure operation was conducted using the “acomp” function while the *ilr* transformation
278 operation was done on the closed data space with the “ilr” function. Mixed models were fitted with the
279 package “nlme” (Pinheiro et al. 2017).

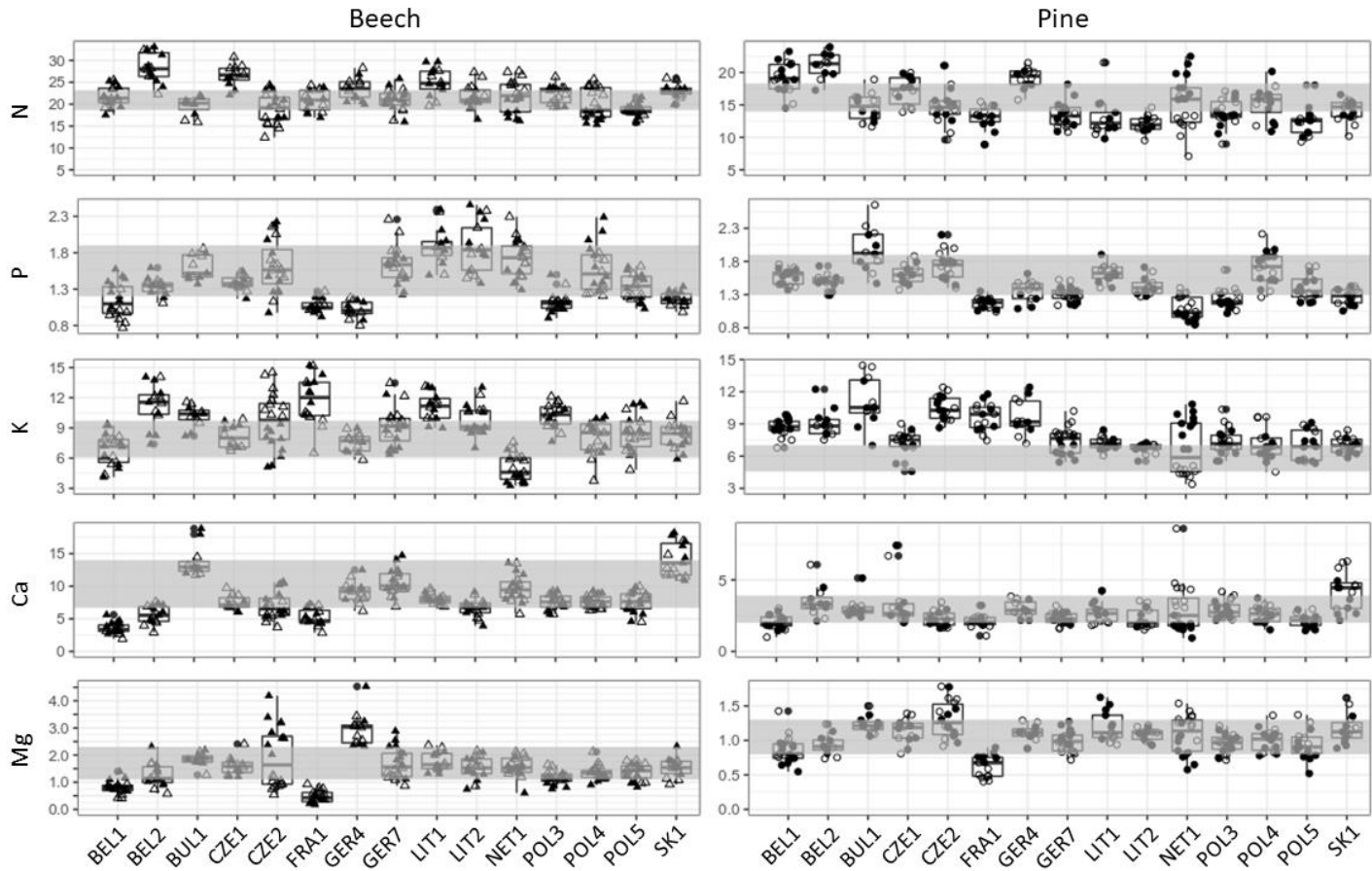
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281 **3. Results**

282 **3.1. Sites nutrient status**

283 The range of foliar nutrient concentrations within and among sites for beech and pine, along with the
284 corresponding optimum range (Mellert and Göttlein 2012), is presented in Figure 2. For both species and
285 most combinations of sites and nutrients, there was a high variability in foliar concentrations. Also, the
286 studied sites spanned a large range of tree nutrient status (Tables A1 and A2).

287 With the exception of K for pine, all nutrients were found to be deficient for at least one combination of
288 site and species (Table A2). While nutrition appeared to be non limiting for neither beech nor pine in two
289 sites (BUL1, CZE1), some sites showed deficiencies for more than one nutrient. For beech, BEL1 and FRA1
290 were deficient for the same three nutrients (P, Ca, Mg); for pine, two different nutrient deficiencies were
291 observed in FRA1 (P, Mg), LIT2 (N, Ca) and POL3 (N, P). At some sites, the same nutrients appeared to be
292 limiting for both beech and pine (Ca in BEL1; P and Mg in FRA1). Two sites (GER7 and LIT1) displayed no
293 deficiencies for beech while deficiencies were observed for pine (P and N limitation in GER7; N limitation
294 in LIT1).



295

296 *Figure 2 Foliar nutrient concentration (mg.g⁻¹) of N, P, K, Ca and Mg in beech leaves (left) and current-year pine needles (right) in*
 297 *the different sites. For each combination of site and species, box plots are used for concentrations across stands (pure, mixed).*
 298 *Individual tree nutrient concentrations are represented by triangles (beech) or dots (pine); filled and open symbols denote pure*
 299 *and mixed stands, respectively. The grey banner indicates the optimum range of foliar nutrient concentrations as defined by*
 300 *Mellert and Göttlein (2012). Y axis may differ between species.*

301

302 3.2. Foliar nutrient composition of beech and pine

303 As shown in Table 2, beech leaves and current year pine needles differed significantly for all investigated
 304 *ilr* balances. The [Fv|Mg, Ca, K, P, N], [K|P, N], [P|N] and [Mg|Ca] balances were higher in beech leaves,
 305 while the [Mg, Ca|K, P, N] was higher in current year pine needles. While P concentrations were similar
 306 for both species, the concentrations of all other nutrients were higher in beech leaves compared to current

307 pine needles; the relative difference between species was most pronounced for Ca, followed by Mg and N
 308 concentrations, and then K (Table 2).

309 *Table 2 Summary of the foliar nutrient composition of beech and current year pine needles. Mean ilr balances (1) and nutrient*
 310 *concentrations (mg.g⁻¹) (2) across all stands and sites (standard errors in parentheses). For each foliar variable (ilr balances or*
 311 *nutrient concentrations) means within a column that are followed by the same letter do not differ at P < 0.05 (Student's t-Tests).*

(1)	[Fv Mg, Ca, K, P, N]		[Mg, Ca K, P, N]		[K P, N]		[P N]		[Mg Ca]	
Beech	-4.83a	(±0.01)	0.77a	(±0.03)	-0.36a	(±0.02)	1.94a	(±0.01)	1.22a	(±0.02)
Pine	-5.17b	(±0.01)	1.34b	(±0.02)	-0.44b	(±0.01)	1.66b	(±0.01)	0.66b	(±0.01)

(2)	N		P		K		Ca		Mg	
Beech	21.9a	(±0.2)	1.4a	(±0.0)	9.0a	(±0.2)	8.1a	(±0.2)	1.5a	(±0.0)
Pine-	15.1b	(±0.2)	1.4a	(±0.0)	8.1b	(±0.1)	2.7b	(±0.1)	1.0b	(±0.0)

312

313

314 **3.3. Stand composition effect on *ilr* balances and nutrient content**

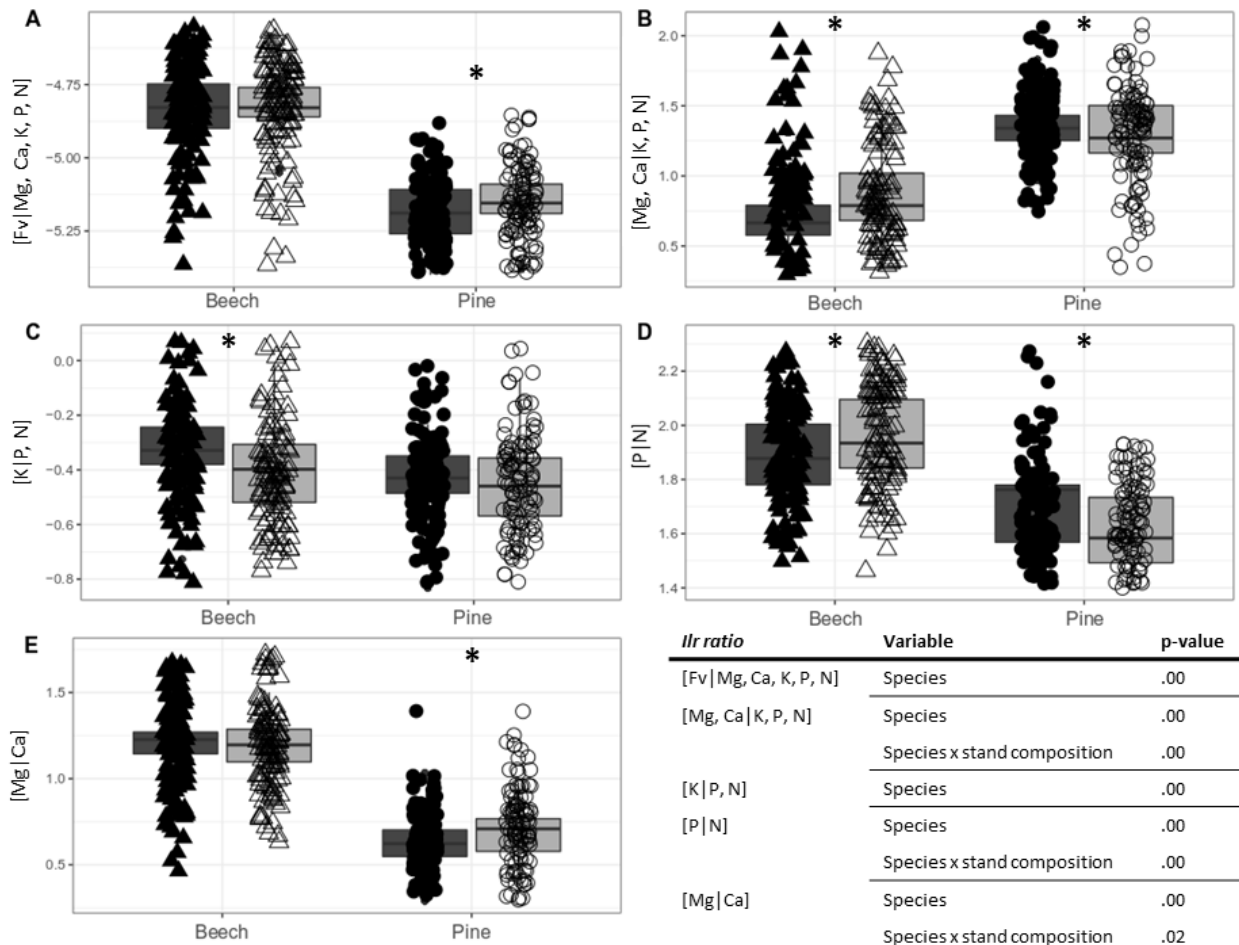
315 Species identity had a significant effect on all *ilr* balances (Figure 3).

316 For all investigated *ilr* balances, at least one species displayed a significant species-mixing effect (Figure 3).

317 For beech, [Mg, Ca|K, P, N] and [P|N] balances were higher in mixed stands than in monocultures while
 318 the opposite was true for [K|P, N]. For pine, mixed stands displayed higher [Fv|Mg, Ca, K, P, N] and
 319 [Mg|Ca] balances and lower [Mg, Ca|K, P, N] and [P|N] balances than monocultures.

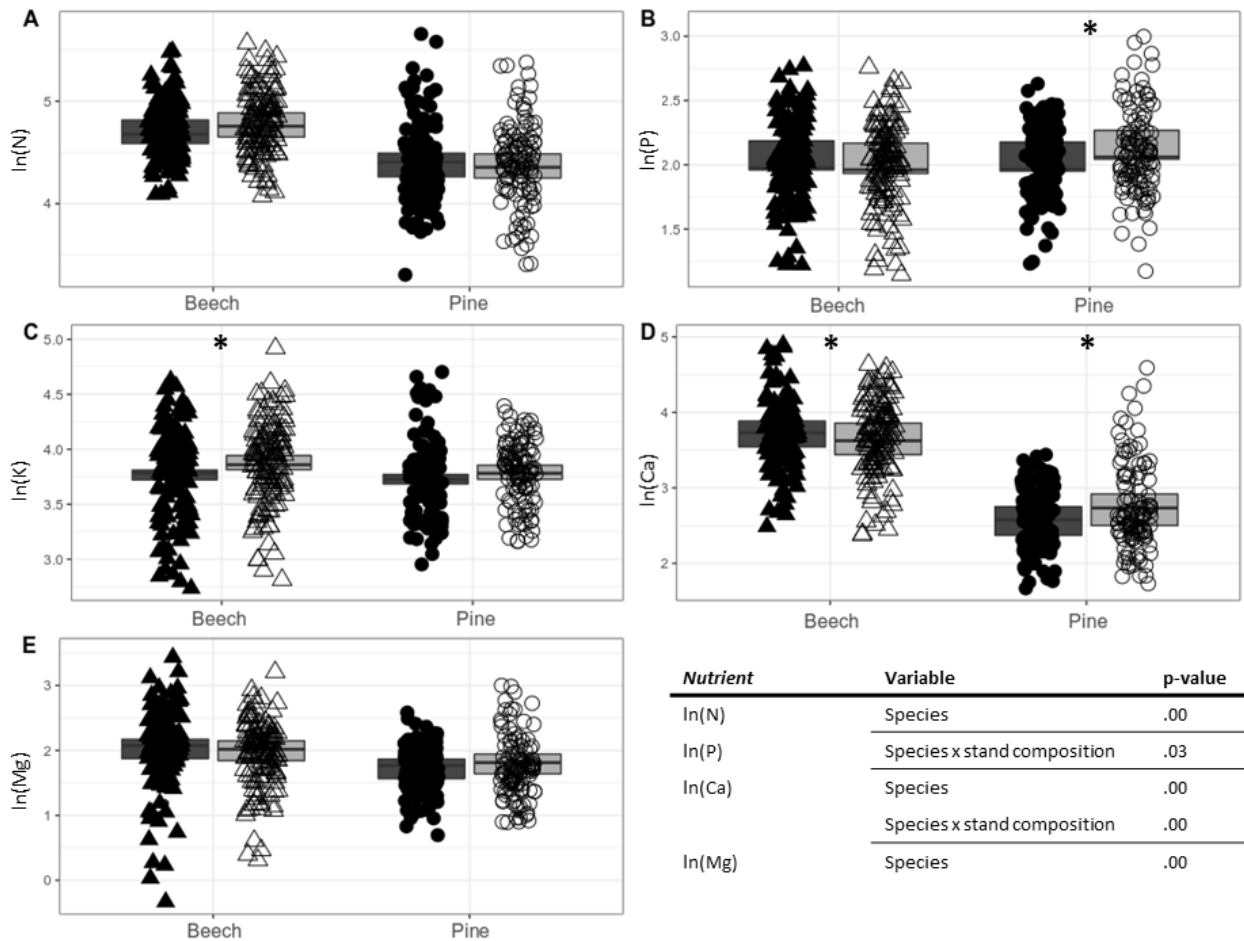
320 For three *ilr* balances ([Mg, Ca|K, P, N], [Mg|Ca] and [P|N]), the effect of stand composition differed
 321 between species (Figure 3; significant species × stand composition interaction), with two *ilr* balances

322 showing opposite effects of mixing: compared to monocultures the [Mg, Ca | K, P, N] and [P | N] balances in
 323 mixed stands were higher for beech and lower for pine.



324
 325 *Figure 3 Beech and pine foliar ilr balances (boxplots and individual (tree) observations) as a function of species identity*
 326 *(beech/pine - current-year needles) and stand composition (pure/mixed) across all sites. The individual tree observations are*
 327 *depicted by black filled triangles (beech) or dots (pine); filled and open symbols denote pure and mixed stands, respectively. Black*
 328 *stars indicate significant differences between pure and mixed stands of a same species at $P < 0.05$ from linear mixed models*
 329 *adjusted on each species and each ilr balance separately. The statistical significant effects of the mixed models (Eq. 3) are shown*
 330 *in the enclosed table, while the detailed statistics for all ilr balance models are given in Table A4.*

332 Looking at nutrient contents (Figure 4), mixing increased P contents in pine and K contents in beech;
 333 compared to the monocultures, Ca contents of the mixed stands were lower in beech and higher in pine.
 334



335
 336 *Figure 4 Beech and pine foliar nutrient contents (boxplots and individual observations as a function of species identity*
 337 *(beech/pine - current-year needles) and stand composition (pure/mixed) across all sites. The individual tree observations are*
 338 *depicted by triangles (beech) or dots (pine); filled and open symbols denote pure and mixed stands, respectively. Black stars*
 339 *indicate significant differences between pure and mixed stands of a same species at $P < 0.05$ from linear mixed models adjusted*
 340 *on each species and each nutrient content separately. The statistically significant effects of the mixed models (Eq. 3) are shown in*
 341 *the enclosed table, while the detailed statistics for all models are given in Table A5.*

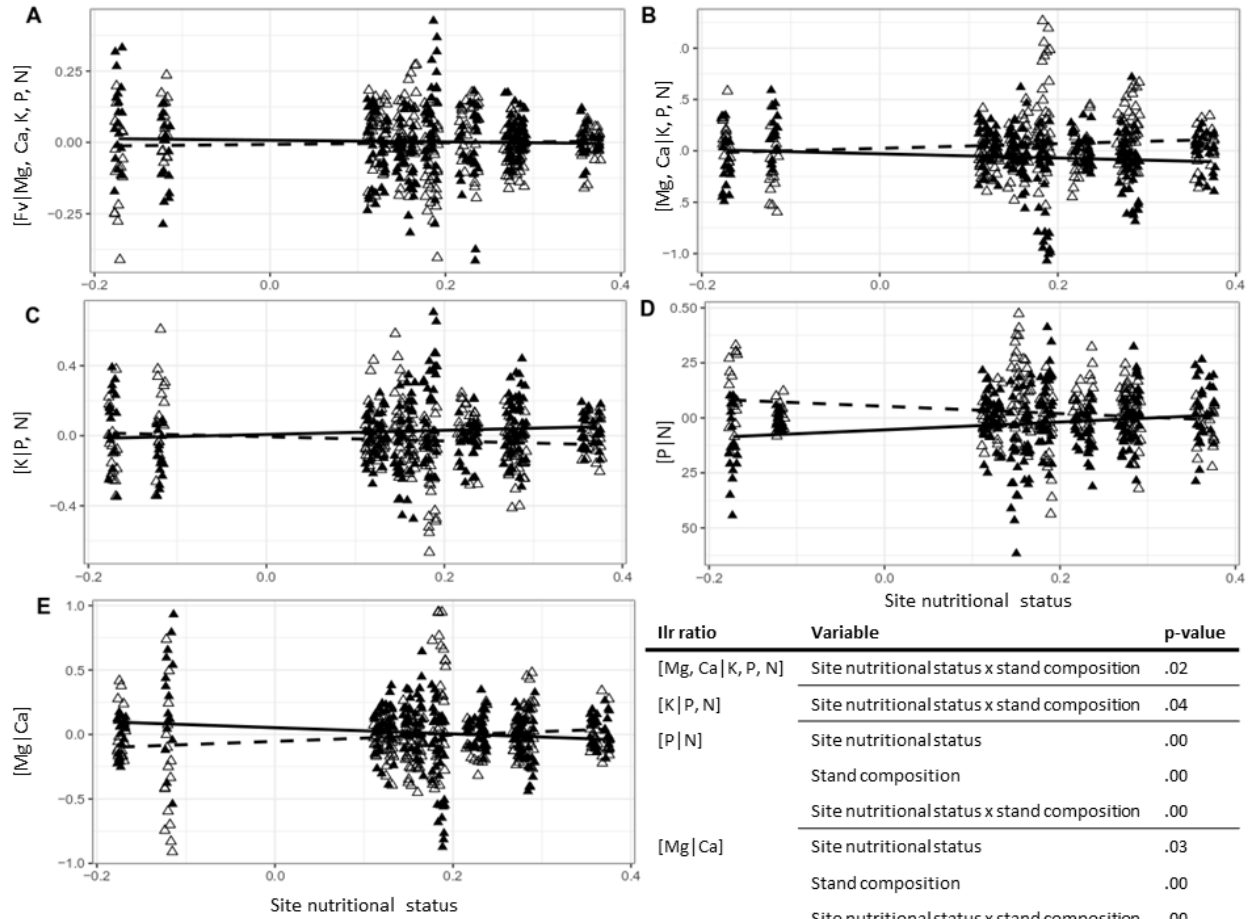
343 **3.4. Impact of site nutrient status on species mixing effect**

344 The effects of species-mixing effect on *ilr* balances along the gradient of site nutritional status for beech
345 and pine are displayed in Figure 5 and 6, respectively.

346 For beech, the [Mg|Ca] and [P|N] balances were significantly affected by the site nutrient level, the stand
347 composition, and their interaction. For both *ilr* balances, the difference between pure and mixed stands
348 was highest at the lower end of the gradient where mixed stands were associated with higher [P|N] and
349 lower [Mg|Ca] balances, compared to pure stands. For the [Mg,Ca|K,P,N] and [K|P,N] balances, only the
350 interaction of site and stand was significant, indicating that the site nutrient status had a different impact
351 in pure and in mixed stands. No significant effect was detected for the [Fv|elements] balance.

352 For pine, the [Fv|elements] balance was significantly higher in mixed than in pure stands, and the effect
353 was constant along the soil fertility gradient. The interaction of site and stand was significant for the
354 [K|P,N] and [Mg|Ca] balances, showing significantly different responses to site nutrient status of pure and
355 mixed stands.

356



357

358

Figure 5 Difference between individual ilr balances of beech trees and the corresponding average value per site as a function of

359

site nutritional status and stand composition. The ilr balance differences are represented by filled (pure stands) or open triangles

360

(mixed stands). The lines represent the predictions for the pure (solid line) and mixed stands (dashed line), using Eq. 4. For all ilr

361

balances, the site nutritional status is quantified by the site-average species-specific [Fv | Mg, Ca, K, P, N] ilr balances. Significant

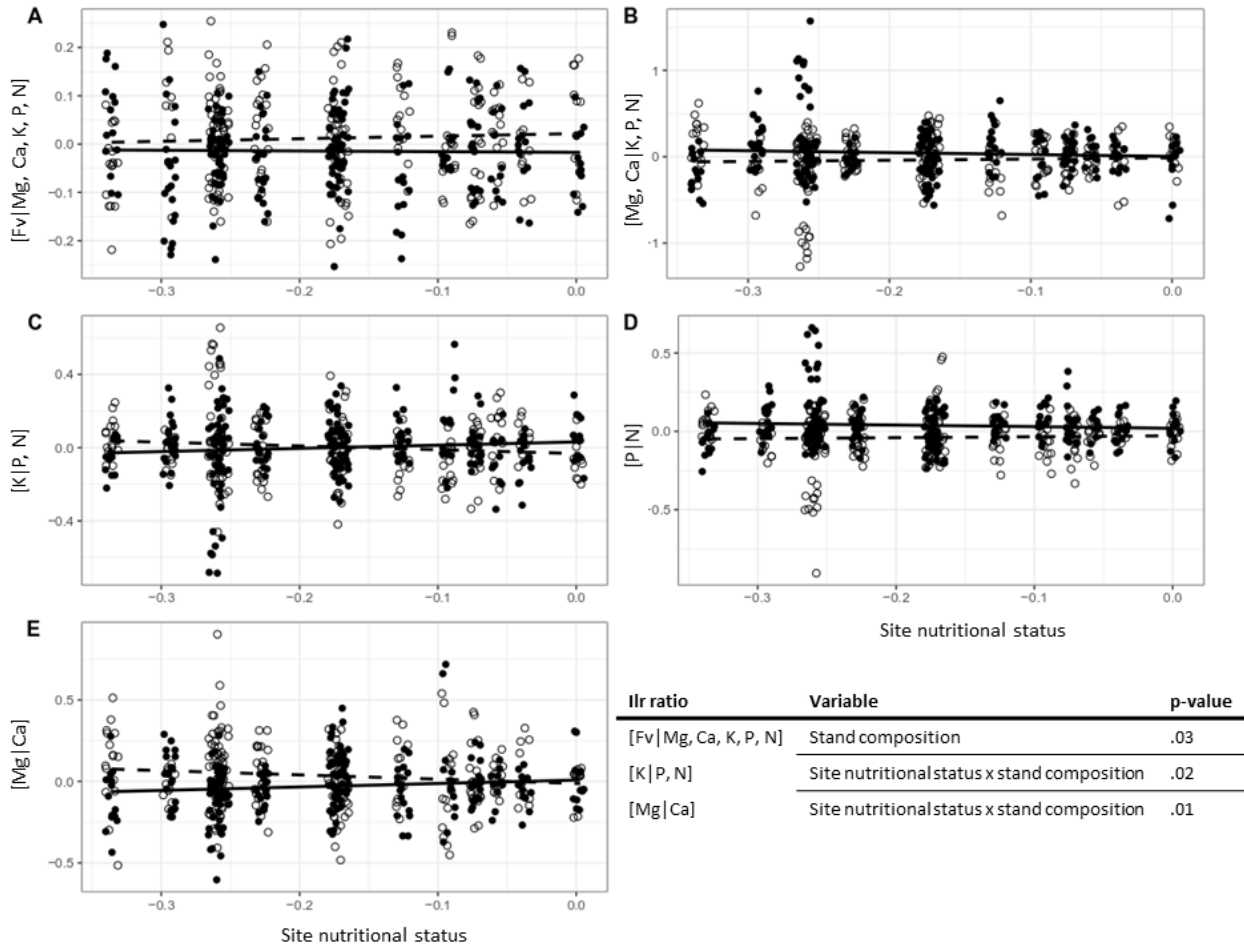
362

effects of the linear models (Eq. 4) are shown in the enclosed table, while the detailed statistics for all ilr balance models are

363

given in Table A6. In the table, default value for stand composition is "pure".

364

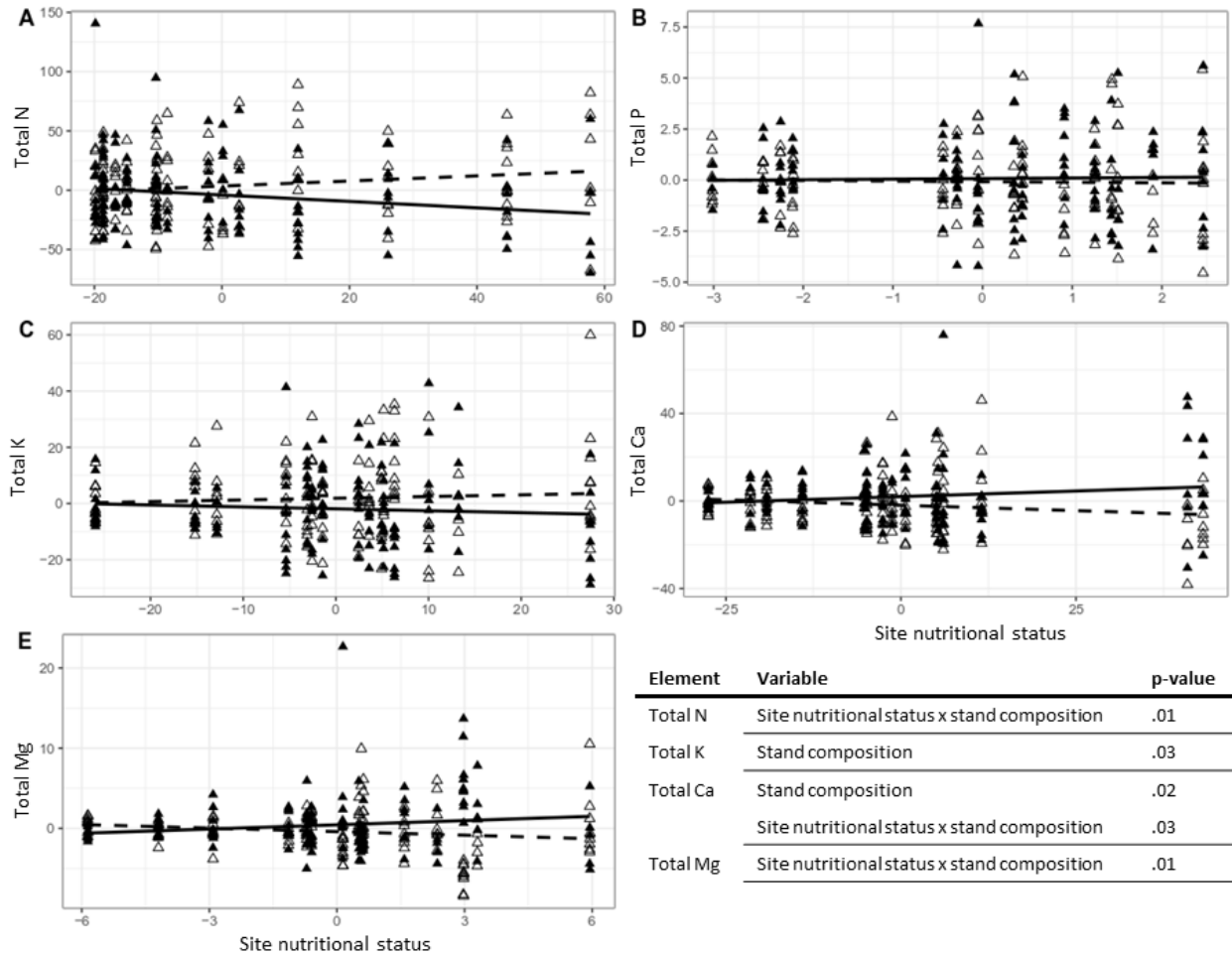


365
 366 *Figure 6* Difference between individual ilr balances of pine trees and the corresponding average value per site as a function of site
 367 nutritional status and stand composition. The ilr balance differences are represented by filled (pure stands) or open dots (mixed
 368 stands). The lines represent the predictions for the pure (solid line) and mixed stands (dashed line), using Eq. 4. For all ilr
 369 balances, the site nutritional status is quantified by the site-average species-specific [Fv|Mg, Ca, K, P, N] ilr balance. Significant
 370 effects of the linear models (Eq. 4) are shown in the enclosed table, while the detailed statistics for all ilr balance models are
 371 given in Table A6. In the enclosed table, default value for stand composition is “pure”.

372
 373 Looking at the nutrient contents of beech leaves (Fig. 7), stand composition had a significant effect on both
 374 K and Ca. The foliar K contents were higher in the mixed stands irrespective of site fertility; by contrast,
 375 the effect of stand composition on Ca differed along the gradient, with lower Ca contents in mixtures on
 376 nutrient-rich sites and almost similar contents for both stand types at the lower end of the gradient. For N

377 and Mg, the effect of site differed between pure and mixed stands, with the strongest difference among
378 stand types observed at the richer sites. There was no main stand composition effect, yet the associated
379 p-value was quite close to 0.05 (Table A7). P contents did not respond to either site nutrient status nor
380 stand composition. For pine nutrient contents (Fig. 8), P and Mg showed a similar pattern. There was an
381 overall significant site and stand composition effect, yet the effect of an increased site fertility was
382 negative in the pure stands and positive in the mixtures. For K and Ca, the effect of site also differed
383 between pure and mixed stands. For K, the interaction was associated with a significant site effect and to
384 opposite effects of stand composition at both extremities of the gradient (higher K content in mixed stand
385 at the lower end of the gradient and lower K content in mixed stand at the higher end of the gradient).
386 The Ca contents of pine needles were on average significantly higher in the mixed stands, yet the stand
387 composition effect was more pronounced as soil fertility increased. N contents in pine needles did not
388 change significantly with either stand composition nor site fertility.

389



390

391 *Figure 7 Difference between individual foliar content of beech trees and the corresponding average value per site as a function of*

392 *site nutritional status and stand composition. The nutrient content differences are represented by filled (pure stands) or open*

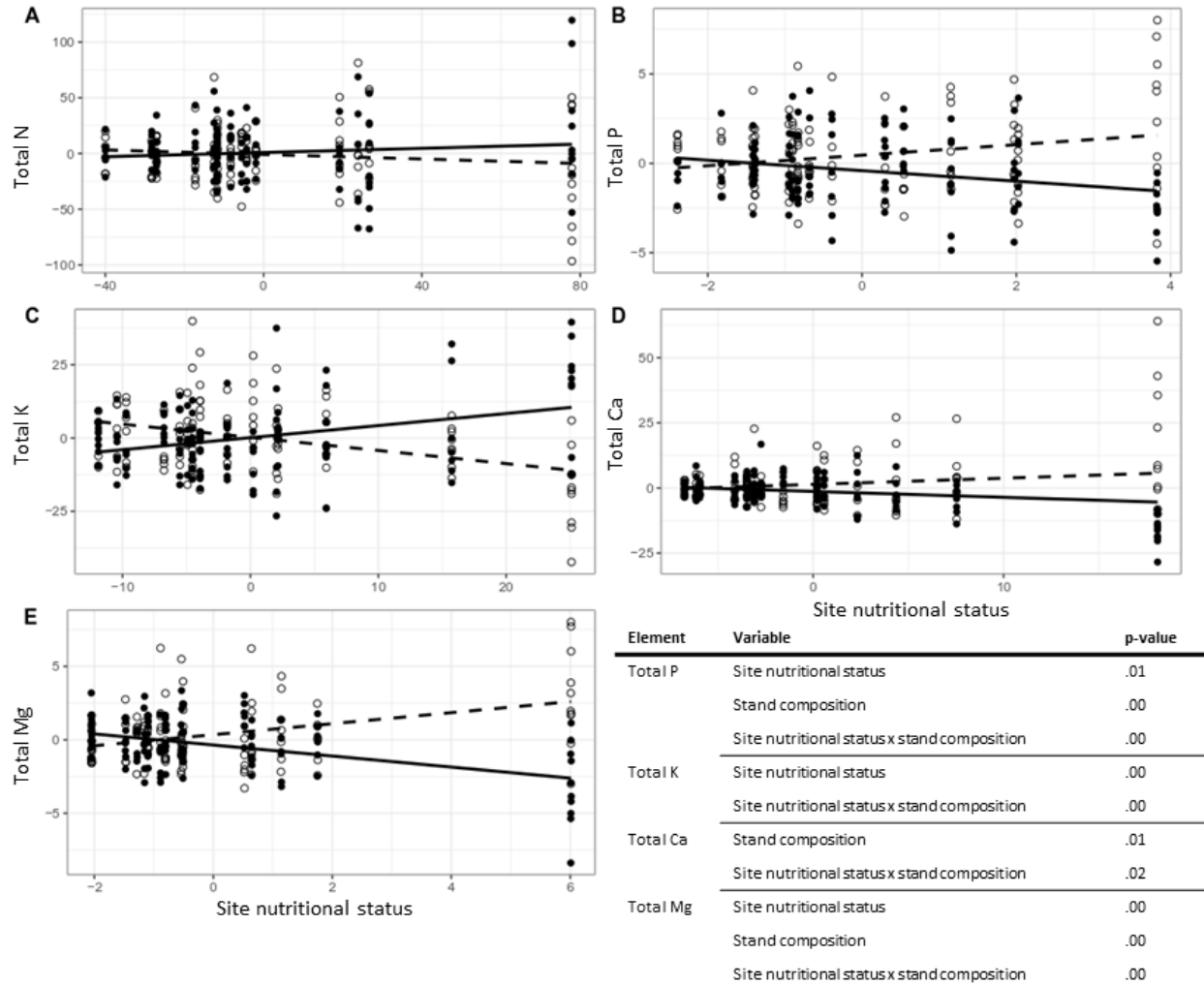
393 *triangles (mixed stands). The lines represent the predictions for the pure (solid line) and mixed stands (dashed line), using Eq. 4.*

394 *For all elements, the site nutritional status is quantified by the site-average species-specific nutrient content. Significant effects of*

395 *the linear models (Eq. 4) are shown in the enclosed table, while the detailed statistics for all nutrient content models are given in*

396 *Table A7. In the enclosed table, default value for stand composition is “pure”.*

397



398

399 *Figure 8 Difference between individual foliar content of pine trees and the corresponding average value per site as a function of*
 400 *site nutritional status and stand composition. The nutrient content differences are represented by filled (pure stands) or open*
 401 *dots (mixed stands). The lines represent the predictions for the pure (solid line) and mixed stands (dashed line), using Eq. 4. For*
 402 *all elements, the site nutritional status is quantified by the site-average species-specific nutrient content. Significant effects of the*
 403 *linear models (Eq. 4) are shown in the enclosed table, while the detailed statistics for all nutrient content models are given in*
 404 *Table A7. In the enclosed table, default value for stand composition is “pure”.*

405

406 4. Discussion

407 4.1. Foliar *ilr* balances and individual nutrient concentrations in beech and pine

408

409 Our results showed that beech leaves and current year pine needles have distinct foliar nutrient
410 signatures. To start with, it is essential to remember that as a result of the computation procedure and
411 notation (Eq. 2), the *ilr* balances behave differently than traditional raw elemental ratios. An increase in *ilr*
412 balance means that either the left part (the balance denominator) has decreased, and/or the right part
413 (the balance numerator) has increased. In this respect, the higher [P|N] and [Mg|Ca] balances in beech
414 leaves compared to pine needles are consistent with the corresponding average nutrient concentrations
415 (Table 2). The increased [K|P,N] in beech leaves can mainly be attributed to the much higher N
416 concentrations in beech leaves compared to pine needles, given the limited difference in K concentrations
417 and the similarity of P concentrations. The significantly higher [Mg,Ca|K,P,N] balance in pine needles can
418 primarily be attributed to the strongly reduced Ca concentrations of pine needles compared to beech
419 leaves, and to a lesser extent, to the lower Mg concentrations of pine needles. Finally, the higher [Fv|Mg,
420 Ca, K, P, N] value in beech leaves is consistent with the higher concentrations of most measured nutrients
421 in beech leaves (Table 2), and their significantly lower Fv value (data not shown); the latter is composed
422 of unmeasured macro- and micro-elements. We are not able to point out those elements, which are
423 involved in this difference.

424 The differences in average nutrient concentrations between beech and pine irrespective of stand type
425 (pure/mixture) were in close agreement with literature data (González de Andrés et al. 2019, Mellert and
426 Göttlein, 2012), confirming the distinctly higher N, Ca and Mg concentrations in beech leaves compared
427 to current year pine needles (Liu et al. 2006, Sardans et al. 2015). According to Sardans et al. (2015) such
428 distinct foliar signatures probably result from long-term differentiation of metabolic and physiological
429 functions and morphology leading to species-specific optimal elemental composition according to the
430 biogeochemical niche hypothesis.

431 The contrasted elemental composition and stoichiometries thus indicates differences in nutritional
432 characteristics between species. Because distinct characteristics between co-occurring species are

433 necessary to entail complementarity, such differences could lead to reduced competition (Sardans et al.
434 2015) and improved nutrition in mixed stands.

435

436 **4.2. Overall mixing effect on *ilr* balances and nutrient content**

437 When significant, the stand composition effect resulted in increased nutrient contents in the mixed stands
438 compared to the monocultures in all (P and Ca in pine, K in beech) but one case (Ca in beech) (Figs. 3 & 4).
439 The positive effect of beech on pine Ca and the corresponding negative effect of pine on beech Ca could
440 be related to the strong difference in nutrient Ca concentration between beech leaves and pine needles
441 (Table 2), resulting in changed nutrient inputs through litterfall. Differences in litterfall between species
442 have been shown to influence soil nutrient availability (Fassnacht and Gowerr 1999; Polyakova and Billor
443 2007; Vesterdal 1998). However, such a difference in nutrient composition between leaves and needles
444 could not explain the other synergistic effects. Indeed, pine P content and K beech content increased in
445 mixed stands compared to pure stands even though foliar P concentrations are similar for both species
446 and K foliar concentrations are lower in pine than in beech. In addition, no stand composition effect could
447 be detected for some nutrients e.g., for nitrogen contents in pine needles, despite the much higher N
448 concentrations in beech leaves compared to pine. Previous studies have found contrasting results on the
449 impact of species mixing on tree foliar nutrition depending on the element considered. Some studies are
450 consistent with our results. For instance, Brandtberg (2001) and Thelin et al. (2002) found P and K
451 concentrations in foliage of conifers to increase if broadleaves were present. Berger et al. (2009) found a
452 tendency to higher Ca concentrations in mixed stands compared to pure stands for coniferous species and
453 lower concentrations for beech stands. On the other hand, our findings differed from some other studies.
454 For instance, Magh et al. (2018) hypothesized broadleaves N acquisition to be lower in mixed stands due
455 to a negative impact of the more recalcitrant conifer litter on the mineralization rates of organic matter.

456 However, we did not find lower amounts of N in beech foliage in mixed stands compared to pure stands.
457 One possible explanation is that soil N availability is also strongly influenced by other environmental
458 drivers than litter quality (Anderson and Domsch 1993, Brown 1992). Such simultaneous influence of
459 several factors and their interaction on species-mixing effects on foliar are not well studied yet.

460 Altogether, our results suggest that differences in foliar nutrient concentrations between associated
461 species and the corresponding change in nutrient inputs, are not strong enough to be used as predictors
462 of overall species-mixing effects. Obviously, additional factors should be considered to explain the large
463 variability in species-mixing effects between sites. This is in line with the results from Rothe et al. (2003).
464 In their study, foliar nutrient pools were not always good predictors of nutrient availability. Several other
465 mechanisms such as the effect of species mixing on litter production through increased canopy packing
466 (Barbeito et al. 2017; Jucker et al. 2015; Pretzsch 2014; Scherer-Lorenzen et al. 2007) or species mixing
467 effect on rates of litter decomposition (Joly et al. 2016; Jonard et al. 2008) can be expected to also come
468 into play although we were not able to properly assess their relevance for and contribution to our findings.

469 More specifically, beech is expected to have a positive impact on pine nutrition through improvement of
470 the humus layer, the upper mineral soil and exploitation of nutrients from deeper soil horizons (Chodzicki
471 1934, Pretzsch et al. 2015). The higher nutrient content, lower C:N ratio and lower amount of recalcitrant
472 compound in beech litter compared to pine (Rumberger et al. 2004, Zhong and Makeschin 2004) could
473 lead to reduced soil acidity and richer humus which will in turn improve nutrient availability in mixed
474 stands compared to pure pine stands (Collin et al. 2016, González de Andrés et al. 2017). In addition, the
475 distinct potential rooting pattern of both species (heart-shaped fine root profile for beech vs. peak of fine
476 root biomass in the most superficial soil layers for pine; Prévosto and Curt 2004) can lead to higher
477 combined root occupancy in mixture and thus, more complete belowground exploitation at least in some
478 sites (González de Andrés 2017).

479 In addition to the observed effect on selected nutrient contents, mixing beech and pine changed nutrient
480 balances. Two *ilr* balances significantly differed between pure and mixed stands for both pine and beech
481 trees: [Mg,Ca|K,P,N], [P|N], yet in an opposite direction (Fig. 3). The three other *ilr* balances were also
482 affected by stand composition, but only for one species. To our knowledge, few studies analyzed the
483 influence of species mixing on tree foliar nutrition using elements balances. Thelin et al. (2002) found the
484 “classical” ratio P:N to be higher in Norway spruce needles when mixed with deciduous species (beech,
485 birch or oak), which is coherent with our finding. The absence of a species mixing effect on beech [Mg|Ca]
486 is coherent with the findings of Colin et al. (2016) who used isometric log-ratios to study the nutritional
487 characteristics of sugar maple and red maple in different contexts (pure stands vs. mixed with conifers;
488 increasing soil acidity levels). They explained the absence of effects by a simultaneous decrease in Mg and
489 Ca with increased conifer proportions. In contrast to our results, however, Collin et al. (2016) also found a
490 decrease in the [Fv|Mg, Ca, K, P, N] balance with increasing proportions of conifers, which they attributed
491 to the effect of increased soil acidity on Al and Mn availability (higher ionic activity in more acidic
492 conditions); as those elements were not measured in the leaves, they were expected to increase
493 correspondingly the Fv value. The dependency of those mechanisms on local factors, like the forest
494 management strategy, the acidifying potential of the species involved or the initial soil acidity could explain
495 contrasting results between studies. Importantly, the change in *ilr* balance suggests that mixing can modify
496 tree nutrition by altering some key nutrient balances, well beyond changing specific individual nutrients.
497 This suspected complementarity with regard to mineral nutrition could be, at least in part, the cause of
498 the increased productivity highlighted on the same network of pine and beech by Pretzsch et al. (2015) as
499 proposed by Burkhardt and Tham (1992), Frivold and Kolström (1999), and Kelty (1992). However, our data
500 set does not allow to properly explore the relationships between tree/stand productivity and nutrition,
501 due to the limited number of sites compared to the potentially high number of processes involved (cf. e.g.
502 Forrester and Bauhus, 2016 for a comprehensive review on those processes).

503

504 **4.3. Mixing effect as a function of site nutrient status**

505 Regarding the impacts of mixing on nutrient contents, there was a close agreement between the results
506 of the overall analysis (Eq. 3) and the analysis where the nutrient gradient was explicitly accounted for (Eq.
507 4); the only exception to this agreement was Mg under pine for which the stand composition effect was
508 not significant in the overall analyses (p -value = 0.13). By contrast, we found more discrepancies between
509 the two approaches for the *ilr* balances. A possible explanation could be related to the index we used to
510 characterize the gradient of nutrient availability for the latter approach ([Fv|Mg, Ca, K, P, N] for all *ilr*
511 balances). We deliberately selected Fv|[Mg, Ca, K, P, N] over the traditional site index (SI) in our case as
512 the latter integrates all site factors, and not only nutrients (Brandl et al. 2014). The use of other indicators
513 such as foliar nutrient remobilization (Achat et al. 2018), soil pH or soil exchangeable cation pools in future
514 studies should help improve our understanding of the dependency of species-mixing effect on site
515 chemical fertility. It is also important to note that, for beech, the sites were not homogeneously distributed
516 along the [Fv|Mg, Ca, K, P, N] gradient (Fig. 5); the two sites at the lower end of the gradient could have
517 had a strong influence on the slope for some models (see the [P|N] balance for instance). One way to cope
518 with this would be to select alternative indicators of soil fertility (see above), and/or to increase the
519 sampling to fill-in the current gaps along the gradient.

520 Focusing on the contents, we expected the mixing effect to increase with a decrease in nutrient availability
521 according to the framework proposed by Forrester and Bauhus (2016). However, in all cases where we
522 found a significant overall stand composition effect (K and Ca for beech, Fig. 7; P, Ca, Mg for pine, Fig. 8),
523 we actually observed the opposite, with only minor or no difference between monocultures and mixtures
524 on the poorest sites, and maximum differences on richer sites.

525 The limited stand composition effect or even the lack of any significant mixing effect in low-nutrient sites
526 could first be explained by the co-occurrence of several nutrient limiting factors. Indeed, the uptake of a
527 nutrient is not only determined by its availability but also by the availability of other nutrients and their
528 interactions, and by environmental conditions such as soil pH or water availability (Marschner 2011;
529 Wilkinson 2000). Simultaneous limitations of distinct nutrients have been observed for both beech and
530 pine stands at several locations (for instance, site FRA1 displayed P and Mg deficiencies for both species
531 and BEL1 displayed Ca deficiency for both species and Mg deficiency for beech) and could originate from
532 the coexistence of several factors limiting nutrient availability. It might be advocated that even in the case
533 where the availability of one limiting nutrient has increased as a result of mixing, this would not translate
534 into an increased foliar content as far as another constraint limiting nutrient availability (e.g. limited water
535 availability) would still be present.

536 A second hypothesis to explain our results is related to all those situations where species interactions may
537 not improve the availability of the limiting nutrient. This could be due to the intrinsic pool of traits present,
538 where inter-specific differences in physiology, phenology or morphology do not improve nutrient
539 availability. Indeed, traits of the species involved in the mixture are of primary importance to determine
540 whether this mixture benefits from improvement of resources availability, uptake or use-efficiency
541 (Ammer 2019). We expected from an association of species with such distinct foliar traits as beech and
542 pine (for instance in term of leaf life span), to induce differences in foliar nutrient composition compared
543 to monocultures. However, such differences in traits may not systematically influence the availability,
544 uptake or resource-use efficiency of nutrients and thus, the foliar nutrient composition. Another possible
545 explanation is the existence of factors limiting the expression of any potential complementarity
546 interaction. Jucker et al. (2014) highlighted such a mechanism in Mediterranean mixed forests. They found
547 that species mixing improved stand growth through complementary light use but that this effect was
548 severely reduced if water resources became limiting. Considering the interspecific differences in rooting

549 patterns between beech and pine (Prévosto and Curt 2004) as a candidate process to increase nutrient
550 availability in mixed species stands compared to monocultures (Forrester and Bauhus 2016), any
551 constraint that would limit the potential development of the rooting system (e.g. anoxic conditions;
552 Kozłowski 1986) would alter this potential spatial stratification.

553 The largest differences in foliar nutrient contents between monocultures and mixed-species stands were
554 observed on the richer sites, with synergistic effects in all cases except for Ca in beech (see previous
555 section). The improved tree nutrient composition of the mixtures at the higher end of the fertility gradient
556 was unexpected, as those situations reflect normal ranges or even a surplus of the nutrient in question.
557 However, an adequate supply of one given nutrient does not exclude potential limitations of others,
558 whereas their availability could in turn limit the uptake of the target nutrient. If mixing removes this
559 constraint, this could explain why a synergistic effect of mixture occurred. An additional explanation is an
560 increased competition in pure stands on the nutrient rich sites due to increased growth, not compensated
561 for by the higher nutrient availability. Bravo-Oviedo et al. (2006) for instance, found that faster tree growth
562 in richer sites lead to increased competition and higher individual tree mortality. In this context, species-
563 mixing could reduce resource competition, leading to an increased species-mixing effect on richer sites.

564

565 **5. Conclusion**

566 Considering the impact of mixing on single nutrient contents, we mostly found either no effect or
567 synergies, with only one case where nutrient contents were lower in mixed stands (antagonism for Ca in
568 beech). Our results showed that the contrasting foliar nutrient compositions of pine and beech trees alone
569 were unable to explain the mixing effects on tree nutrition, which suggests that other mechanisms than a
570 simple change of nutrient inputs through litterfall are at play. Far beyond changing the foliar contents of

571 single individual nutrients, mixing was found to significantly alter key nutrient balances in both beech
572 leaves and current year pine needles.

573 Our analysis also revealed unexpected patterns of complementarity as a function of nutrient availability,
574 with only minor or no differences between monocultures and mixtures at the lower end of the nutritional
575 gradient, and maximum differences on nutrient rich soils. This clearly points to further expanding the
576 existing frameworks to account for the multivariate nature of tree nutrition. In this paper, we made a first
577 attempt in that direction by working on both contents and *ilr* balances. Further work is needed to
578 disentangle the underlying processes, using a combination of field work and modeling approaches. This
579 claims for the establishment of controlled standardized experiments across Europe to test for specific
580 mechanisms.

581

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588

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593

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