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4 **IMPROVING SOIL SIMAZINE DISSIPATION THROUGH AN ORGANIC**  
5 **AMENDMENT INOCULATED WITH *TRAMETES VERSICOLOR***

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19 **Running title: Dissipation of simazine by organic amendments**

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21

## 22 **ABSTRACT**

23 The aim of this research was to investigate the effect of applying sewage sludge combined with  
24 wheat crop residue as an organic amendment on the dissipation rate of simazine spiked at 2 and 20  
25 mg kg<sup>-1</sup> in an Andisol soil from southern Chile. Changes in some soil enzymes related to soil quality  
26 were measured by spectrophotometry, simazine dissipation rates were measured by gas  
27 chromatography, and biomass production in the contaminated soil were evaluated. Results of this  
28 study indicated that application of the organic amendment inoculated with *Trametes versicolor*  
29 enabled a decrease in negative effects of the herbicide on soil enzymatic activities and a reduction  
30 in final concentrations of simazine (~80% at both doses). The simazine half-life time was reduced  
31 from 14 to 10 days and from 36 to 15 days for doses of 2 and 20 mg kg<sup>-1</sup>, respectively. These  
32 findings prove that the combined strategy of biostimulation and bioaugmentation using these  
33 residues can be effectively used to reduce residue pesticides in soils, mainly by increasing the  
34 microbiological activity, thus improving simazine dissipation in an Andisol soil.

35

36 **Keywords:** Biodegradation; Bioremediation; Pesticides; Soil biochemical properties

37

## 38 **1. INTRODUCTION**

39

40 Agriculture is one of the most important productive activities in the world, being vital for solving  
41 the global food demand, which increases annually due to a sustained increase in population. To  
42 improve crop yield and productivity, the addition of compounds like herbicides is a normal practice  
43 for crop development (Salem et al., 2017). However, the constant and excessive addition of  
44 herbicides can negatively impact the environment on different levels, mainly through the  
45 progressive accumulation and persistence of xenobiotics in natural environments (Hashmi et al.,  
46 2017; Pinochet et al., 2018). The herbicide simazine (2-chloro-4,6-bis(ethylamino)-s-triazine) is a  
47 widely used selective, systemic s-triazine, which has been broadly applied to control broad-leaved  
48 weeds and annual grasses that affect various crops (Cheng et al., 2017). Such widespread use  
49 affects non-target terrestrial and aquatic environments (Martínez-Iñigo et al., 2010). Arias-Estévez  
50 et al. (2008) pointed out that only a minor fraction of an applied herbicide reaches the target  
51 (<0.1%), whereas the remaining fraction contributes to environmental pollution. At these

52 microsites, the mobility of herbicides in bulk soils depends on processes such as retention, transport  
53 and degradation (Jensen et al., 2018).

54 The use of soil microorganisms for soil bioremediation is a convenient and promising tool that has  
55 been applied to reduce the adverse effects of xenobiotics in soils (Gao et al., 2014). In this context,  
56 white-rot fungi are strong candidates to be applied in soil environmental cleanup due to their ability  
57 to synthesize nonspecific selective enzymes (laccases, lignin peroxidase and manganese  
58 peroxidase) that could be useful in dissipating soil contaminants (Camacho-Morales et al., 2017;  
59 Coelho-Moreira et al., 2018). Hence, bioaugmentation strategies using these fungi have been  
60 reported to accelerate the onset of degradation, improve soil health in heavy metals contaminated  
61 soils and protect the native microbiological communities against adverse effects (Arriagada et al.,  
62 2010).

63 Several organic residues are used in agroforestry systems, including crop residues such as wheat  
64 straw and sewage sludge. Both types of residue could be an excellent raw material to apply in some  
65 biostimulation strategies (Almonacid et al., 2015; Kumari et al., 2018). Furthermore, the addition  
66 of an organic amendment as a nutrient source for endemic microorganisms improves the physical,  
67 chemical and biological properties of soil, mainly through the increase in overall microbiological  
68 activity promoted by the availability of organic and inorganic nutrients (Almonacid et al., 2015).

69 In this study we hypothesized that sewage sludge combined with wheat crop residues, used as an  
70 organic amendment, can be used as a bioaugmentation-biostimulation strategy to promote simazine  
71 dissipation in an Andisol soil. Additionally, *Trametes versicolor* in combination with the organic  
72 amendment can contribute to increasing simazine dissipation rates and improving soil health. For  
73 this purpose, a simazine-contaminated Andisol soil was supplemented with *T. versicolor* inoculated  
74 in an organic amendment comprising sewage sludge and wheat straw residues. Soil enzyme  
75 activities were assessed to evaluate biochemical changes, and GC-ECD analyses were used to  
76 quantify the simazine dissipation rate.

77

## 78 **2. MATERIALS AND METHODS**

79

### 80 **2.1. Simazine**

81

82 High purity simazine (99%; Sigma-Aldrich, San Luis, MO) was used in the experiments. Soils  
83 were spiked with 10 ml of simazine diluted in acetone to perform contamination at doses of 2 and  
84 20 mg kg<sup>-1</sup> (estimated field dose and 10X field dose, respectively).

85

## 86 **2.2. Soil characteristics**

87

88 The collected soil was an Andisol from southern Chile belonging to the Freire family (38°50'S and  
89 72°35'W; medial, mesic, Typic Placudands) with a silty loam texture (CIREN, 2002). Soil samples  
90 were collected from the surface layer (0–20 cm), air dried, sieved through a 2 mm mesh, and  
91 characterized according to the methods described in Sadzawka *et al.* (2004). Briefly, the organic  
92 matter content was measured using the Walkley-Black method. The pH was measured in soil  
93 suspensions with deionized water (1:2.5 w:v). Cation exchange capacity (CEC) was calculated  
94 from the total exchangeable bases (Mg, Ca, K, and Na extracted by 1 M ammonium acetate at pH  
95 7.0) analyzed by flame atomic absorption spectrophotometry.

96

## 97 **2.3. Organic residues and saprophytic fungi**

98

99 Wheat straw corresponded to crop residues from the Araucania Region and the stabilized sewage  
100 sludge was collected from municipal wastewater plant ESSAL S.A., Osorno, Chile. Chemical  
101 characterization of the residues were: i) 43 % cellulose, 30 % hemicellulose, 9 % lignin, C:N ratio  
102 87, pH 5.5, 46 % carbon, total N 0.5 %, total P content 1.5 mg kg<sup>-1</sup>, Cu content 2.5 mg kg<sup>-1</sup> and Zn  
103 content 5.8 mg kg<sup>-1</sup> for wheat straw; and ii) 1.96 % cellulose, 11.0 %, hemicellulose, 1.3 % lignin,  
104 C:N ratio 8.5, pH 12.81 % organic matter, 46 % carbon, total N 0.6 %, total P content 19500 mg  
105 kg<sup>-1</sup>, Cu content 113 mg kg<sup>-1</sup> and Zn content 399 mg kg<sup>-1</sup> for sewage sludge.

106 Saprophytic fungi *T. versicolor* (A-1369) was obtained from the culture collection of the Centro  
107 de Investigaciones Biológicas in Madrid. The fungal inoculum was prepared as follows: active  
108 mycelia plugs of *T. versicolor* from 7-day-old culture on potato dextrose agar were placed in 400  
109 mL of potato dextrose broth and incubated at 30 °C and 200 rpm for 15 days. Then, 30 mL aliquots  
110 of inoculum were used to start growth in 250 mL Erlenmeyer flasks containing 40 g of sterilized  
111 organic residues (10 g of wheat straw and 30 g of sewage sludge), which was the best combination  
112 for enhancing enzyme production (Almonacid *et al.*, 2015). The inoculated substrate was incubated

113 for 2 weeks in darkness and then incorporated into the soil microcosms (40 g of inoculated substrate  
114 mixed with 460 g of soil). The moisture was monitored gravimetrically and periodically adjusted  
115 by adding distilled water.

116

#### 117 **2.4. Amendment microcosms**

118

119 Experiments to evaluate the simazine dissipation rate, enzymatic activities and effects on soil  
120 microbial activity were performed in eight different microcosms. Each microcosm consisted of a  
121 glass pot with 500 g of soil, at approximately 60 % of water holding capacity (WHC), supplemented  
122 with the organic amendment and/or simazine (Table 1). Soil and the organic amendment were  
123 mixed and then preincubated for 2 weeks at 25°C in darkness prior to applying the contamination  
124 dose. After that, the simazine solution was slowly spiked over the microcosms. The simazine  
125 concentrations were controlled at 2 and 20 mg kg<sup>-1</sup>. Samples were taken at days 0, 1, 5, 15 and 30  
126 after simazine application for enzymatic determinations and simazine content.

127

#### 128 **2.5. Soil biochemical determinations**

129

130 Soil biochemical analyses were performed on the different eight microcosms at days 0, 1, 5, 15 and  
131 30 after simazine application as follows: i) Soil acid phosphatase activity was determined using *p*-  
132 nitrophenyl phosphate as an orthophosphate monoester analogue according to Sannino and  
133 Gianfreda (2001); ii)  $\beta$ -glucosidase activity was measured according to Tabatabai and Bremner  
134 (1969); iii) Total microbial activity was measured by monitoring FDA hydrolysis (Adam and  
135 Duncan, 2001); and iv) Manganese peroxidase (MnP) activity was measured by mixing 1 g of soil,  
136 2.0 mL of sodium tartrate (0.1 M pH 5.0), 2.0 mL of H<sub>2</sub>O<sub>2</sub> (0.1 mM) and 2.0 mL of MnSO<sub>4</sub> (0.1  
137 mM). Samples were incubated at 25 °C for 30 min and measured at an absorbance of 420 nm. All  
138 activities were assayed in triplicate and reported on a dry soil basis.

139

#### 140 **2.6. Determination of residual simazine**

141

142 Ten grams of soil (60% of its WHC) were submitted to simazine extraction by shaking the soil and  
143 20 mL of acetone for 60 min at 150 rpm, followed by sonication for 30 min. The resulting

144 suspension was transferred to 50 mL falcon tubes and centrifuged at 5000 rpm for 10 min. One mL  
145 of supernatant was passed through activated florisisil (2 g) and sodium sulfate (1 g) columns with 5  
146 mL of acetone. Samples were lyophilized and concentrated to 1 mL. Samples were filtered and  
147 analyzed for simazine quantifications. The concentrations of simazine in extracts were analyzed  
148 with a Shimadzu (Shimadzu Corp., Kyoto, Japan) gas chromatograph coupled with an electron  
149 capture detector (GC-ECD) using a RTX-5 column (Restek Corp., Bellefonte, PA). The column  
150 was programmed from an initial temperature of 60 °C for 1 min to 140 °C at a rate of 12 °C min<sup>-1</sup>,  
151 held for 1 min, and then ramped at a rate of 8 °C min<sup>-1</sup> to 240 °C with a final hold time of 4 min.  
152 The detector and injector were maintained at 320 °C and 250 °C, respectively. The injector was in  
153 splitless mode for detection. To determine the detection limits, aliquots of soil were spiked with  
154 simazine. Recovery percentages were 83.25 ± 1.70 for S2; 81.19 ± 7.83 for SR2; 80.74 ± 3.21 for  
155 SRT2; 96.89 ± 3.88 for S20; 94.68 ± 4.82 for SR20; and 91.98 ± 1.24 for SRT20. The first-order  
156 rate of degradation and the DT50 (time required for 50% of the initial dose of pesticide to be  
157 degraded) of each compound in each soil were determined with the following equations:

158 
$$Ct = Co \cdot e^{-kt} \quad t_{1/2} = \frac{\ln 2}{k}$$

159 Where Ct is the concentration of pesticide remaining in soil (mg kg<sup>-1</sup>) after t (days), Co is the initial  
160 concentration of pesticide (mg kg<sup>-1</sup>), and k is the rate of degradation (day<sup>-1</sup>) (Swarcewicz and  
161 Gregorczyk, 2013).

162

## 163 **2.7. Greenhouse experiments**

164

165 *Solanum lycopersicum* was used as the test plant. Seeds were surface sterilized with NaClO for 15  
166 min, thoroughly rinsed with sterilized distilled water. Four weeks after germination, ten uniform  
167 seedlings were transplanted to individual 1 L pots containing the remaining soil to evaluate the  
168 presence of simazine and/or to evaluate the toxicity of the degradation products. The plants were  
169 grown in a greenhouse for 45 days and the dry shoot and root biomass were measured.

170

## 171 **2.8. Statistical analyses**

172

173 All results were analyzed by one-way ANOVA. Means and standard errors of ten replicates were  
174 calculated for enzymatic activity, simazine dissipation and shoot and root biomass. Statistical

175 significance was set at  $p < 0.05$ . All statistical tests were conducted using the R software (R Core  
176 Team 2018; <https://www.R-project.org>).

177

### 178 **3. RESULTS**

179

#### 180 **3.1. Soil enzyme activities**

181

182 Results for acid phosphatase,  $\beta$ -glucosidase, Mn peroxidase and FDA hydrolysis are shown in  
183 Table 2. FDA hydrolysis was similar for all treatments at the beginning of the experiments. The  
184 higher values were obtained at day five ( $p < 0.05$ ) for both doses. The simazine application caused  
185 a decrease in the enzymatic activity at day 1, mainly due to the change induced by the presence of  
186 the herbicide. Higher activity values were obtained in treatments with the organic amendment  
187 inoculated with *T. versicolor* (SRT, SRT2, and SRT20) (Table 2).

188 Acid phosphatase activity was similar in the control treatments (S, SR, and SRT). The results also  
189 showed little variations induced by the addition of simazine in the control treatments S, S2 and  
190 S20. Higher values of enzymatic activity were noted in the treatments with the organic amendment  
191 (SR2, SR20, SRT2, and SRT20) (Table 2).

192  $\beta$ -glucosidase was higher in treatments with the organic amendment (SR, SR2, SR20, SRT2, SRT2,  
193 and SRT20) at days 0, 1, and 5. However, treatments S2 and S20 showed results similar to  
194 treatments with the residues after day 5. At day 15, results were similar for all treatments. In  
195 general, values of enzymatic activity decreased over time, as in the control (S, SR, and SRT). In  
196 treatments with the organic amendment,  $\beta$ -glucosidase activity was higher on the first day ( $p <$   
197  $0.05$ ). In soils with simazine the enzymatic activity increased significantly, which was higher in  
198 S20, SR20 and SRT20 at day 5.

199 MnP activity increased at day 5 for all treatments, with the highest values for treatments SR2 and  
200 SRT2. After the peak, the values decreased. At day 30 values were similar to the control (S, SR,  
201 and SRT) for both doses.

202

#### 203 **3.2. Simazine removal from the soil microcosm**

204

205 Results of simazine degradation are shown in Fig. 1A for the estimated field dose ( $2 \text{ mg kg}^{-1}$ ) and  
206 in Fig. 1B for the 10X field dose ( $20 \text{ mg kg}^{-1}$ ). The recovery percentages of simazine by GC ECD  
207 were approximately 83.25 % for the treatment S2, 81.19 % for SR2 and SRT2, 90.59 % for S20,  
208 and 94.08 % for SR20 and SRT20. For treatments with the estimated field dose  $2 \text{ mg kg}^{-1}$  (Fig.  
209 1A), results showed that simazine degradation started at day 1 with no significant differences.  
210 Degradation of simazine at day 5 was higher in treatments with the residue (inoculated or not) than  
211 the control soils (S2) ( $p < 0.05$ ). After that, the degradation processes were slow, but continued. At  
212 day 30, total removed simazine was approximately 78.79 % for the control treatment S2;  
213 approximately 82.22 % for SR2; and 88.62 % for SRT2. For treatments with simazine at  $20 \text{ mg kg}^{-1}$   
214 (Fig. 1B), the degradation kinetic was similar to the field dose. The results show that the  
215 dissipation process also began within 24 hours, even though the differences were not significant  
216 until day 5, where there were clear differences ( $p < 0.05$ ) between treatments with the residue  
217 (SR20 and SRT20) and the control treatment S20, which had a higher simazine content. At day 30,  
218 the end of the analysis, the removed simazine was about 43.75 % for the control treatment (S20),  
219 approximately 62.28 % for treatment SR20 and approximately 73.90 % for SRT20.

220 For simazine persistence, a first-order kinetic was used, and the estimation of half-life times is  
221 reported in Table 3. The half-life times were higher in treatments without organic amendment (S2  
222 and S20). In particular, for  $2 \text{ mg kg}^{-1}$ , the half-life values were reduced from 13.6 to 12.0 days for  
223 the biostimulation strategy (SR2), and from 13.6 to 9.8 days with the organic amendment  
224 inoculated with *T. versicolor* (SRT2). Similarly, in treatments with  $20 \text{ mg kg}^{-1}$  the DT50 values  
225 were reduced from 35.9 to 21.3 days with the organic amendment (SR20) and from 35.9 to 15.8  
226 days with the organic amendment inoculated with *T. versicolor* (SRT20).

227

### 228 **3.3. Greenhouse experiments**

229 After 45 days of culture of *S. lycopersicum* in the spiked soils, there were statistical differences  
230 between controls and treatments without the organic amendment (S2 and S20) (Fig. 2). In general,  
231 we showed a tendency to increase plant biomass in treatments with the organic amendments, in  
232 particular aerial dry weight ( $p < 0.05$ ). Furthermore, the application of simazine to the soils induced  
233 a decrease in the production of biomass, which was stronger in treatments with the higher simazine  
234 dose (Fig. 2). These negative effects were lower when the organic residue was biotransformed by

235 *T. versicolor*, showing biomass production similarly to that obtained in the control treatments (S;  
236 Fig. 2)

237

#### 238 **4. DISCUSSION**

239

240 Our study showed that the application of simazine induces several changes in soil enzymatic  
241 activities, especially in the FDA hydrolysis. FDA hydrolysis occurs by the action of lipase, esterase,  
242 protease and hydrolase enzymes, hydrolyzing nonspecifically fluorescein diacetate. Thus, FDA  
243 hydrolysis measures the direct action of soil microorganisms (Casucci et al., 2003). In general the  
244 higher values of FDA hydrolysis in treatments SRT, SRT2, and SRT20 are caused by *T. versicolor*,  
245 which has been characterized as a fungus with a strong extracellular enzymatic apparatus, and is a  
246 suitable candidate for use in bioremediation strategies (Bastos and Magan, 2009).

247 We also showed negative effects of simazine on the phosphatase activity. Acid phosphatase plays  
248 an important role in soils, specifically in the conversion of organic phosphorous to bioavailable  
249 forms for plants and microorganisms (Huang et al., 2003). When the organic amendment was  
250 incorporated, the overall acid phosphatase activity was improved, establishing an effective strategy  
251 for improving the soil biochemical properties of the simazine-contaminated soil.

252 Our study tested whether simazine affects the  $\beta$ -glucosidase activity, mainly by stimulating native  
253 soil microorganisms, which respond to stress and thus raise the level of enzyme production.  
254 Glucosidase activity is important for organic matter decomposition, specifically in the hydrolysis  
255 of  $\beta$ -glucosidase bonds of large carbohydrate chains present in lignocellulosic residues (Han and  
256 Chen, 2008).

257 In the case of MnP, our results showed that this enzyme can play an important role in early  
258 degradation stages of the herbicide, improving the microbiological activity. MnP activity is one of  
259 the enzymes involved in the degradation of soil pollutants, because the enzyme is directly involved  
260 in the oxidation of chemicals compounds (Pizzul et al., 2009). Cea et al. (2010) showed a positive  
261 correlation between the activity of this enzyme and the degradation of pentachlorophenol as a result  
262 of a bioaugmentation strategy using *Anthracoxyllum discolor* in an Andisol soil. These results  
263 agree with those reported in our study, where *T. versicolor* is a strong candidate to promote  
264 biodegradation of the herbicide.

265 Our study showed simazine removal from the soil microcosms. The differences in final simazine  
266 concentrations in treatments S2 and S20 (without biostimulation and bioaugmentation) are  
267 explained by the presence of native soil microorganisms, which were able to degrade the herbicide  
268 under the conditions of this study. These degradation rates increased when the soil incorporated  
269 the organic amendment (biostimulation strategies) in treatments SR2 and SR20, because the  
270 organic amendment promoted the development and enzymatic activity of certain strains of  
271 microorganisms, improving the simazine dissipation rates. On the other hand, when the organic  
272 amendment inoculated with *T. versicolor* was used (treatments SRT2 and SRT20), the dissipation  
273 rates increased due to the metabolic activity of *T. versicolor*. The incorporation of the organic  
274 amendment inoculated with *T. versicolor* makes it possible to obtain free nutrient sources for  
275 natural soil microbiota. The main form of herbicide degradation involves soil microbiological  
276 communities (Van Eerd et al., 2003). In addition, pesticide adsorption to organic matter and clays  
277 also plays a fundamental role in decreasing the amount of herbicide available in the soil solution  
278 (Đurović et al., 2009). Bastos and Magan (2009) studied the behavior of *T. versicolor* in a soil  
279 contaminated with high levels of atrazine, finding optimal degradation rates, classifying this fungus  
280 as a candidate for use in bioremediation strategies of soils contaminated with s-triazine herbicides.  
281 Similar results were reported by Morgante et al. (2012), who found simazine degradation by soil  
282 microorganism strains similar to those used in our study, showing microbial degradation rates of  
283 simazine, which is the main degradation mechanism of simazine under natural conditions.  
284 Likewise, the use of a bioaugmentation technique with a strain of *Pseudomonas* sp. can reduce the  
285 half-life times of the herbicide in soil. In our study, we showed that soil preincubation with an  
286 organic amendment (inoculated or not) can improve microbial responses to simazine contamination  
287 and improve the soil quality (Leskovar and Othman, 2018). This preincubation caused an increase  
288 in the enzymatic activity and improved the soil conditions in order to obtain better simazine  
289 degradation rates and a lower half-life time.

290 Our study showed that the presence of simazine in the soil affects the growth of *Solanum*  
291 *lycopersicum*, especially in the soil treatments with the highest residual simazine. Although the  
292 biomass production was higher in the amendment treatments, we showed that the non-amendment  
293 soil also can allow plant development, in spite of the lower biomass obtained, which agrees with  
294 the natural simazine dissipation rate obtained in treatments without the organic amendment. The  
295 fact that some microorganisms can use simazine as a nitrogen source for growth is a relevant factor

296 that contributes to simazine dissipation and could also help reduce ground water contamination  
297 through leaching (Chris Wilson et al., 2011; Dinamarca et al., 2007). These natural processes are  
298 possible due to natural microorganisms, which can promote the mineralization of xenobiotic  
299 compounds to less contaminating forms. However, these processes are slower in non-amendment  
300 soils and require more time to complete the dissipation to less contaminating forms such as NH<sub>3</sub>  
301 and CO<sub>2</sub> (Morgante et al., 2012), producing smaller plants than in amendment soils. With respect  
302 to amended microcosms, the presence of organic matter improves the interchangeable sites in the  
303 soil and therefore induces lower pesticide availability in the soil soluble fractions (Flores et al.,  
304 2009). Furthermore, adding fungal strains to the soil as a bioaugmentation strategy is an effective  
305 technique for improving biological and chemical soil properties. The use of bioaugmentation to  
306 degrade simazine using bacterial strains has been reported (Flores *et al.*, 2009), but little is known  
307 about the ability of fungal strains to dissipate simazine residues themselves. The effect on plant  
308 growth promotion of *S. lycopersicum* may be an indirect process in which the fungus produces  
309 different extracellular enzymes (Nakatani et al., 2010), degrading both the pollutant and the organic  
310 residues; a process that indirectly stimulates the proliferation of certain bacterial and/or fungal  
311 strains involved in the degradation of simazine, improving the soil conditions to promote plant  
312 growth (Morgante et al., 2012). Organic amendment, either natural or inoculated, can be an  
313 effective strategy to apply in bioremediation processes, because it contains available organic  
314 carbon, mineral elements essential for the growth of microorganisms and plants (Almonacid et al.,  
315 2015). Previous research has described these techniques as addressing the degradation of  
316 xenobiotic compounds (Ghazali et al., 2004; Suja et al., 2014). In this study, we tested two  
317 bioremediation techniques: biostimulation and bioaugmentation, and both can be effectively used  
318 to contribute to improving simazine dissipation rates in the Andisol soil.

319

## 320 **5. CONCLUSIONS**

321

322 This study showed that *T. versicolor* contributes to simazine dissipation. In addition, the use of a  
323 combined biostimulation and bioaugmentation strategy with *T. versicolor* can improve simazine  
324 dissipation rates, demonstrating the potential of this technique to design bioremediation strategies  
325 to recover simazine-contaminated soils. Also, we saw negative effects of simazine on the growth  
326 of *Solanum lycopersicum* plants and the enzymatic activities of simazine-treated soils; however,

327 these effects can be reduced by the use of the biostimulation and bioaugmentation strategy  
328 performed.

329

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331

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334

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**Table 1:** Description and abbreviations of the microcosm treatments performed in the experiments.

Simazine dose	Treatments		
0	Soil ( <b>S</b> )*	Soil + Residues ( <b>SR</b> )*	Soil + Residues + <i>T.versicolor</i> ( <b>SRT</b> )
2	Soil ( <b>S2</b> )*	Soil + Residues ( <b>SR2</b> )*	Soil + Residues + <i>T.versicolor</i> ( <b>SRT2</b> )*
20	Soil ( <b>S20</b> )*	Soil + Residues ( <b>SR20</b> )*	Soil + Residues + <i>T.versicolor</i> ( <b>SRT20</b> )*

434 \* Abbreviations of the treatments. (**S**) soil; (**S2**) soil plus simazine at 2 mg kg<sup>-1</sup>; (**S20**) soil plus  
 435 simazine at 20 mg kg<sup>-1</sup>; (**SR**) soil plus organic residues; (**SR2**) soil plus organic residues and  
 436 simazine at 2 mg kg<sup>-1</sup>; (**SR20**) soil plus organic residues and simazine at 20 mg kg<sup>-1</sup>; (**SRT**) soil  
 437 plus organic residues + *T. versicolor*; (**SRT2**) soil plus organic residues + *T. versicolor* and  
 438 simazine at 2 mg kg<sup>-1</sup>; (**SRT20**) soil plus organic residues + *T. versicolor* and simazine at 20 mg  
 439 kg<sup>-1</sup>.

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**Table 2:** Enzymatic activities in the different soil microcosms throughout time. Data are means  $\pm$  standard error. Statistical differences were investigated for each day between treatments. The same letter is not significantly different according to Tukey's multiple range test ( $p < 0.05$ )

Enzyme	Time (days)	S <sup>a</sup>	S2 <sup>b</sup>	S20 <sup>c</sup>	SR <sup>d</sup>	SR2 <sup>e</sup>	SR20 <sup>f</sup>	SRT <sup>g</sup>	SRT2 <sup>h</sup>	SRT20 <sup>i</sup>
<b>Acid phosphatase</b> ( $\mu\text{mol PNP g}^{-1}\text{h}^{-1}$ )	0	209.1 $\pm$ 2.1 a	210.4 $\pm$ 1.3 a	209.0 $\pm$ 2.9 a	203.0 $\pm$ 1.4 b	200.3 $\pm$ 1.9 b	199.5 $\pm$ 0.9 b	202.8 $\pm$ 2.2 b	212.0 $\pm$ 0.4 a	191.0 $\pm$ 1.5 c
	1	206.0 $\pm$ 1.3 b	211.2 $\pm$ 0.9 a	169.4 $\pm$ 1.1 g	200.5 $\pm$ 0.8 d	190.3 $\pm$ 1.2 f	201.9 $\pm$ 1.2 cd	204.3 $\pm$ 1.5 bc	194.0 $\pm$ 0.7 e	202.1 $\pm$ 1.0 cd
	5	130.4 $\pm$ 1.5 f	144.0 $\pm$ 0.9 e	144.2 $\pm$ 2.3 e	174.2 $\pm$ 2.0 d	202.8 $\pm$ 1.4 b	190.4 $\pm$ 2.4 c	205.9 $\pm$ 1.6 b	235.0 $\pm$ 0.7 a	189.6 $\pm$ 0.9 c
	15	121.0 $\pm$ 3.0 g	158.6 $\pm$ 0.7 e	144.9 $\pm$ 1.5 f	181.8 $\pm$ 2.0 c	195.8 $\pm$ 1.5 b	196.9 $\pm$ 1.1 b	174.5 $\pm$ 2.9 d	195.3 $\pm$ 1.8 b	211.2 $\pm$ 1.5 a
	30	125.4 $\pm$ 0.6 f	154.2 $\pm$ 1.4 c	132.7 $\pm$ 1.7 e	84.5 $\pm$ 2.2 g	163.9 $\pm$ 1.4 b	164.4 $\pm$ 1.5 b	144.0 $\pm$ 2.3 d	165.6 $\pm$ 0.6 b	196.8 $\pm$ 0.4 a
<b><math>\beta</math>-glucosidase</b> ( $\mu\text{mol PNG g}^{-1}\text{h}^{-1}$ )	0	37.7 $\pm$ 0.5 d	39.1 $\pm$ 0.8 d	38.6 $\pm$ 0.8 d	56.1 $\pm$ 1.9 bc	54.7 $\pm$ 0.6 c	56.9 $\pm$ 1.1 bc	58.4 $\pm$ 1.4 b	54.8 $\pm$ 1.1 c	61.8 $\pm$ 1.1 a
	1	35.0 $\pm$ 0.3 f	57.3 $\pm$ 0.6 cd	54.9 $\pm$ 0.6 de	51.5 $\pm$ 1.4 e	52.7 $\pm$ 1.8 e	65.6 $\pm$ 0.5 a	56.6 $\pm$ 3.0 cd	58.8 $\pm$ 0.6 bc	61.9 $\pm$ 0.9 b
	5	31.1 $\pm$ 0.6 f	47.2 $\pm$ 1.2 c	59.1 $\pm$ 1.6 b	42.6 $\pm$ 0.4 d	37.8 $\pm$ 0.3 e	69.8 $\pm$ 1.3 a	45.0 $\pm$ 1.1 cd	42.2 $\pm$ 0.7 d	60.5 $\pm$ 1.5 b
	15	30.2 $\pm$ 0.3 d	33.2 $\pm$ 0.7 c	35.0 $\pm$ 0.6 c	29.3 $\pm$ 0.5 d	38.6 $\pm$ 0.3 b	44.3 $\pm$ 0.7 a	33.9 $\pm$ 1.0 c	25.4 $\pm$ 0.9 e	30.0 $\pm$ 0.9 d
	30	28.8 $\pm$ 0.8 ab	23.7 $\pm$ 0.8 de	21.5 $\pm$ 0.7 e	30.0 $\pm$ 1.0 ab	25.8 $\pm$ 0.5 cd	24.6 $\pm$ 1.0 d	31.2 $\pm$ 1.1 a	28.9 $\pm$ 1.2 ab	27.9 $\pm$ 1.1 bc
<b>Manganese peroxidase</b> ( $\mu\text{mol Mn}^{+3} \text{g}^{-1} \text{min}^{-1}$ )	0	262.5 $\pm$ 0.3 c	263.8 $\pm$ 0.5 bc	261.4 $\pm$ 1.0 c	249.5 $\pm$ 0.3 d	251.8 $\pm$ 2.0 d	251.2 $\pm$ 0.3 d	265.6 $\pm$ 1.6 b	299.6 $\pm$ 0.5 a	302.1 $\pm$ 0.3 a
	1	263.9 $\pm$ 0.3 d	251.0 $\pm$ 0.1 g	218.5 $\pm$ 0.3 h	252.9 $\pm$ 0.4 f	252.8 $\pm$ 1.2f	258.8 $\pm$ 0.5 e	272.0 $\pm$ 0.5 c	306.7 $\pm$ 0.5 b	360.2 $\pm$ 0.5 a
	5	320.3 $\pm$ 0.1 f	331.6 $\pm$ 0.3 e	356.1 $\pm$ 0.3 d	302.9 $\pm$ 0.4 g	508.7 $\pm$ 1.4 a	412.5 $\pm$ 0.2 c	351.4 $\pm$ 10.2 d	437.9 $\pm$ 0.2 b	419.1 $\pm$ 0.7 c
	15	293.7 $\pm$ 1.1 f	318.1 $\pm$ 0.6 d	325.2 $\pm$ 0.6 c	262.9 $\pm$ 0.3 g	347.2 $\pm$ 1.1 b	399.2 $\pm$ 0.3 a	329.5 $\pm$ 5.0 c	309.3 $\pm$ 0.1 e	309.5 $\pm$ 0.3 e
	30	237.6 $\pm$ 2.5 b	228.2 $\pm$ 0.8 d	209.4 $\pm$ 0.1 e	204.9 $\pm$ 0.3 f	168.4 $\pm$ 1.5 g	209.1 $\pm$ 0.2 e	246.9 $\pm$ 1.4 a	209.3 $\pm$ 0.2 e	232.4 $\pm$ 1.2 c
<b>Fluorescein diacetate</b> ( $\mu\text{g FDA g}^{-1}$ )	0	57.9 $\pm$ 0.1 c	56.9 $\pm$ 0.2 c	56.6 $\pm$ 0.2 c	63.7 $\pm$ 0.1 a	63.1 $\pm$ 0.1 ab	62.3 $\pm$ 0.2 b	62.6 $\pm$ 1.4 ab	63.0 $\pm$ 0.3 ab	63.8 $\pm$ 0.2 a
	1	56.5 $\pm$ 0.2 d	50.6 $\pm$ 0.2 f	51.3 $\pm$ 0.3 f	63.3 $\pm$ 0.2 b	51.3 $\pm$ 0.2 f	52.0 $\pm$ 0.2 ef	66.8 $\pm$ 1.4 a	53.3 $\pm$ 0.3 e	59.4 $\pm$ 0.2 c
	5	56.3 $\pm$ 0.1 g	81.7 $\pm$ 0.1 b	64.6 $\pm$ 0.1 ef	65.4 $\pm$ 0.2 e	62.8 $\pm$ 0.2 f	76.4 $\pm$ 0.3 c	72.3 $\pm$ 1.8 d	88.6 $\pm$ 0.3 a	87.9 $\pm$ 0.3 a
	15	57.9 $\pm$ 0.1 e	65.2 $\pm$ 0.2 d	66.8 $\pm$ 0.1 dc	65.3 $\pm$ 0.2 d	69.9 $\pm$ 0.1 b	77.4 $\pm$ 0.2 a	67.5 $\pm$ 2.1 c	70.6 $\pm$ 0.1 b	67.7 $\pm$ 0.1 c
	30	45.7 $\pm$ 0.3 f	42.0 $\pm$ 0.1 h	43.8 $\pm$ 0.1 g	50.5 $\pm$ 0.3 b	44.8 $\pm$ 0.2 fg	57.8 $\pm$ 0.3 c	52.1 $\pm$ 0.8 e	56.0 $\pm$ 0.4 d	69.2 $\pm$ 0.3 a

S=Soil; S2= Soil + simazine at 2 mg Kg<sup>-1</sup>; S20= Soil + simazine at 20 mg Kg<sup>-1</sup>; SR= Soil + organic residue; SR2= Soil + organic residue + simazine at 2 mg Kg<sup>-1</sup>; SR20= Soil + organic residue + simazine at 20 mg Kg<sup>-1</sup>; SRT= Soil + organic residue + *Trametes versicolor*; SRT2= Soil + organic residue + *T. versicolor* + simazine at 2 mg Kg<sup>-1</sup>; SRT20= Soil + organic residue + *T. versicolor* + simazine at 20 mg Kg<sup>-1</sup>

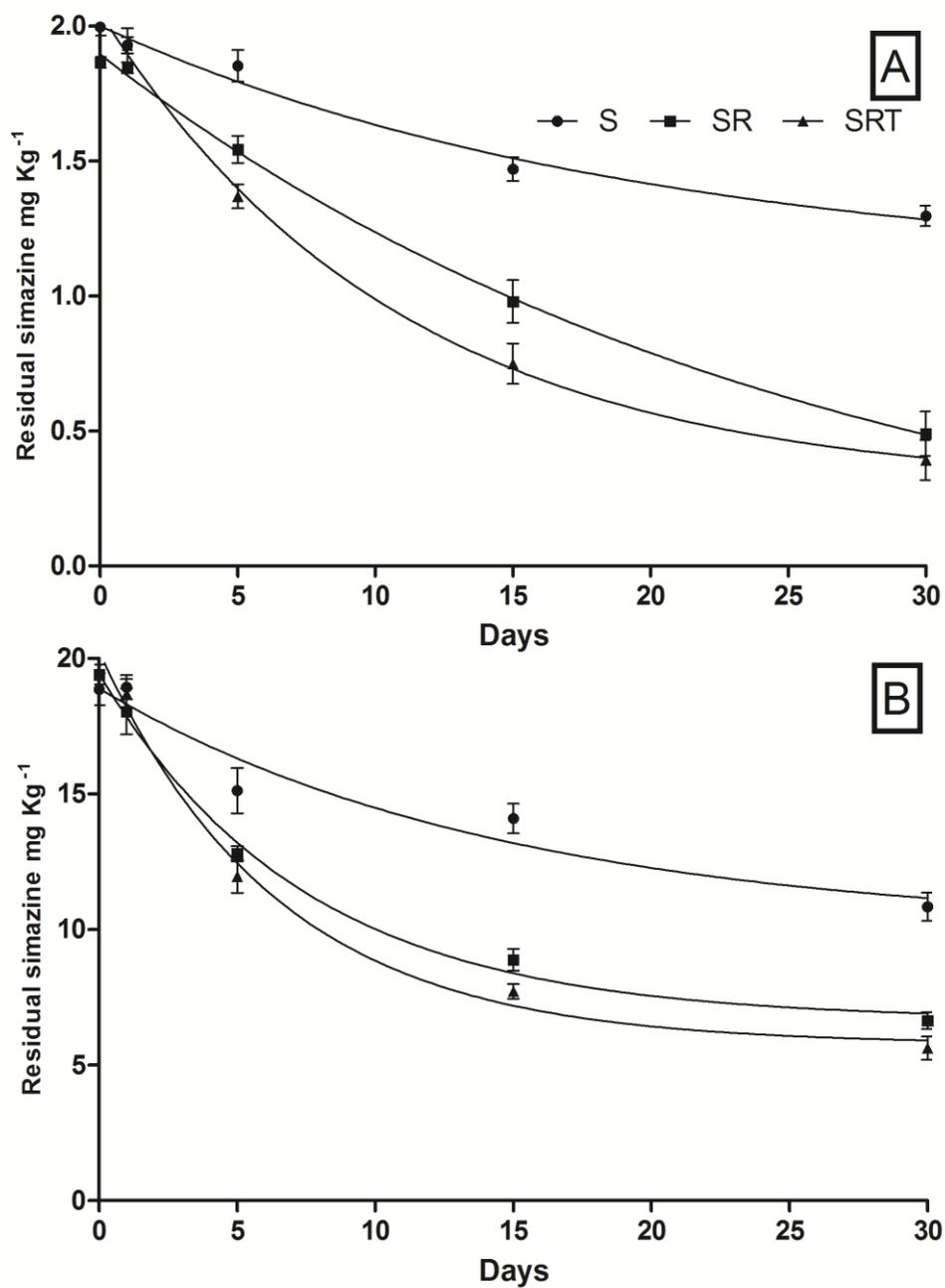
**Table 3:** Degradation rate ( $k$ ) and half-life ( $t_{1/2}$ ) values for simazine in the different microcosms.

Treatments	R <sup>a</sup>	$k^a$ (days <sup>-1</sup> )	$t_{1/2}$ (days) <sup>a</sup>
S2	0.981 ± 0.011	0.051 ± 0.010	13.6 ± 0.4
SR2	0.979 ± 0.008	0.058 ± 0.004	12.0 ± 0.9
SRT2	0.983 ± 0.006	0.071 ± 0.007	9.8 ± 0.9
S20	0.944 ± 0.042	0.019 ± 0.002	35.9 ± 4.0
SR20	0.945 ± 0.055	0.033 ± 0.005	21.3 ± 3.2
SRT20	0.984 ± 0.003	0.044 ± 0.004	15.8 ± 1.5

445 <sup>a</sup> For every treatment, the average was calculated from the DT50 values from three soil  
446 replicates (each in triplicate). (**S2**) soil plus simazine at 2 mg kg<sup>-1</sup>; (**S20**) soil plus simazine  
447 at 20 mg kg<sup>-1</sup>; (**SR**) soil plus organic residues; (**SR2**) soil plus organic residues and simazine  
448 at 2 mg kg<sup>-1</sup>; (**SR20**) soil plus organic residues, and simazine at 20 mg kg<sup>-1</sup>; (**SRT**) soil plus  
449 organic residues + *T. versicolor*; (**SRT2**) soil plus organic residues + *T. versicolor* and  
450 simazine at 2 mg kg<sup>-1</sup>; (**SRT20**) soil plus organic residues + *T. versicolor* and simazine at 20  
451 mg kg<sup>-1</sup>. A first order kinetic model was used.

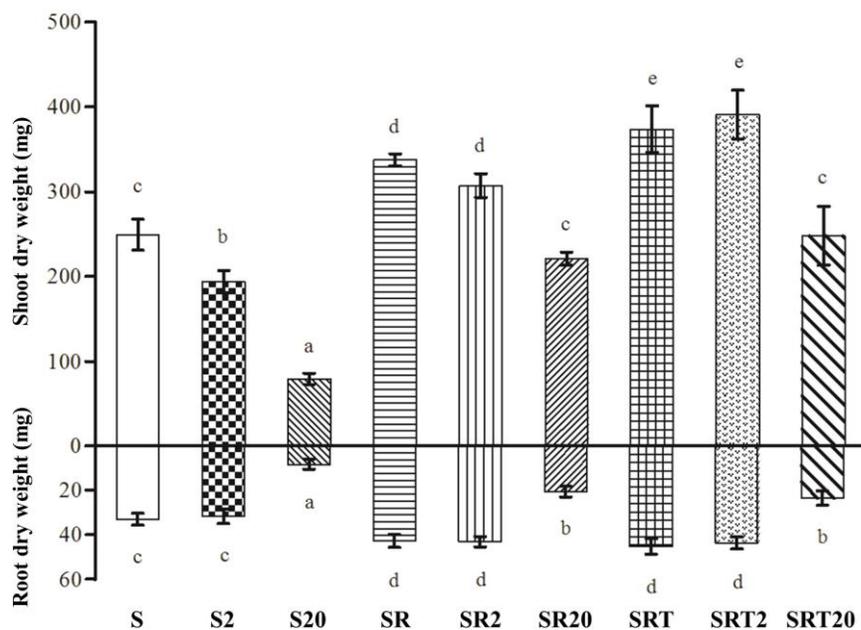
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456 **Figure 1.** Simazine degradation in soils with an initial contamination of 2 mg kg<sup>-1</sup> (A) and  
 457 20 mg kg<sup>-1</sup> (B); (S) control soils, (SR) soil plus organic residues and (SRT) soil plus organic  
 458 residues + *T. versicolor*.



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460 **Figure 2.** Shoot and root dry weight of *Solanum lycopersicum* plants established in the  
 461 treated soil. The plants were harvested 45 days after sowing. (S) soil control; (S2) soil plus  
 462 simazine at 2 mg kg<sup>-1</sup>; (S20) soil plus simazine at 20 mg kg<sup>-1</sup>; (SR) soil plus organic residues;  
 463 (SR2) soil plus organic residues and simazine at 2 mg kg<sup>-1</sup>; (SR20) soil plus organic residues  
 464 and simazine at 20 mg kg<sup>-1</sup>; (SRT) soil plus organic residues + *T. versicolor*; (SRT2) soil  
 465 plus organic residues + *T. versicolor* and simazine at 2 mg kg<sup>-1</sup>; (SRT20) soil plus organic  
 466 residues + *T. versicolor* and simazine at 20 mg kg<sup>-1</sup>. The same letter is not significantly  
 467 different according to Tukey's multiple range test ( $p < 0.05$ ).

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