¹³⁷Cs inventories along a climatic gradient in volcanic soils of Patagonia:

potential use for assessing medium term erosion processes

Ludmila La Manna^{a,b,*}, Leticia Gaspar^c, Manuela Tarabini^{a,b}, Laura Quijano^d, Ana Navas^{c**}

^a Centro de Estudios Ambientales Integrados, Facultad de Ingeniería, Universidad Nacional de la Patagonia San Juan Bosco (UNPSJB), Ruta 259 Km 16,4, 9200 Esquel, Chubut, Argentina

^b Consejo Nacional de Investigaciones Científicas y Técnicas (CONICET), Argentina

^c Departamento de Suelo y Agua, Estación Experimental de Aula Dei, EEAD-CSIC. Avda. Montañana 1005, Zaragoza, 50059, Spain

^d Georges Lemaître Centre for Earth and Climate Research - Earth and Life Institute, Université Catholique de Louvain, Louvain-la-Neuve, Belgium

*Corresponding author. E-mail address: llamanna@unpata.edu.ar (L. La Manna)

**Co-corresponding author. E-mail address: anavas@eead.csic.es (A. Navas).

HIGHLIGHTS

- ¹³⁷Cs inventory in Patagonia depends on precipitation and non-crystalline clays
- ¹³⁷Cs is mainly retained in the uppermost 10 cm of the soil profiles
- ¹³⁷Cs along soil profiles is controlled by organic matter
- High ¹³⁷Cs loss and erosion rates under different land uses in forest/steppe ecotone
- ¹³⁷Cs allows to assess erosion rates in Patagonian Andean Region

ABSTRACT

Fallout radionuclides, such as Caesium-137, were proven to be a valuable means for studying medium-term (c.a. 50 years) soil erosion processes. In order to gain knowledge on the spatial distribution of ¹³⁷Cs in volcanic soils of Andean Patagonia,

¹³⁷Cs reference inventories were established along a precipitation gradient. At the subhumid sector, patterns of ¹³⁷Cs gain/loss associated with land use were also determined considering native forests under grazing, a degraded rangeland and a pine plantation.

Along the rising rainfall gradient (750 – 1400 mm), pedogenesis of volcanic ash soils, non-crystalline mineral formation, organic matter content and soil porosity varied, increasing in moister areas. Radionuclide inventories varied along the edaphoclimatic gradient, reaching mean values of 192, 267 and 576 Bq m⁻², at study areas with 750, 950 and 1400 mm of annual precipitation, respectively. The ¹³⁷Cs inventory followed an exponential relationship with precipitation, which could be related to the presence of allophane as the colloidal material in the soils from the rainiest area. The penetration depth reached by ¹³⁷Cs varied between 15 and 25 cm, according to rain amount and soil texture. Most of the ¹³⁷Cs fallout was retained in the uppermost 10 cm of the profiles and an exponential decline of ¹³⁷Cs with depth, highly related to organic matter contents, was found.

At the subhumid study area, both ¹³⁷Cs mass activity and inventory, significantly decreased under the different land uses, with respect to reference soils. Although the pattern of ¹³⁷Cs gain/loss varied according to topography, soil properties (organic matter and porosity) and vegetation cover, showing eroding and aggrading profiles, most samples had ¹³⁷Cs values lower than the reference value, suggesting loss of soil as a consequence of erosion processes. Furthermore, as much as 45%, 58% and 70% of sample points from native forests, plantation and rangeland, respectively, had ¹³⁷Cs values below the limit of detection. In the study transects, the loss of the upper 15 cm of soil in the subhumid sector during the last 50 years, highly exceeding tolerable erosion rates, highlights the urgent need for applying effective soil conservation measures.

Reference inventories, which vary according to the edaphoclimatic gradient, and the loss of the radionuclide in sites with anthropic intervention, show the potential for using ^{137}Cs measurements for assessing erosion processes in the Patagonian Andean Region.

Key words: Fallout Radionuclide; Erosion rates; Non-crystalline aluminosilicates; Land use; Forest; Rangeland

1. INTRODUCTION

Soil erosion is a widespread problem throughout the world, frequently associated with land use changes, with negative impacts on sustainable production as well as on environmental conservation (e.g. Lal, 2017; Lizaga et al., 2018). Furthermore, land use changes have strong impacts on variation of soil properties that occur in few decades (Navas et al., 2008; Lizaga et al., 2019). Patagonia is not outside of this global trend. The Patagonian Andean Region has suffered high anthropic pressure, associated with both the late nineteenth century period of white settlement and the earlier activities of the aboriginal population, who used fire for hunting guanacos (Veblen et al., 1999). Severe overgrazing by sheep and livestock during the early decades of twentieth century was the main factor in the shift in vegetation and soil degradation (Veblen and Lorenz, 1988). During the last century and at present, land use changes have triggered several soil erosion processes, being the region subjected to deforestation, grazing, fires, agriculture, exotic plantations, and urbanization (Carabelli and Scoz, 2016; Colazo et al. 2019). Recent studies showed that the replacement of native forest modified physical, chemical and magnetic soil properties, resulting from soil desiccation and led to selective removal of fine particles by erosion (La Manna et al., 2018a). Volcanic soils,

dominant in this region, were highly erodible where the mineral soil remained exposed, without litter nor vegetation cover (La Manna et al., 2016).

In order to understand the influence of land use on soil degradation and to select effective soil conservation measures, reliable quantification of soil erosion rates is a requirement (Navas et al., 2014). Isotopic techniques based on the use of fallout radionuclides such as Caesium-137 (^{137}Cs), affords a valuable means for studying erosion and deposition processes within the landscape and have been increasingly applied for quantifying medium-term (c.a. 50 years) soil erosion rates, mainly in the northern hemisphere. ^{137}Cs was introduced into the atmosphere by the detonation of fission nuclear weapons. Its use for erosion studies is based on the strong fixation of the radionuclide by the fine components of the soil (Tamura, 1964) and its subsequent redistribution in the landscape by physical processes. Although the global pattern of ^{137}Cs inventories shows a latitudinal variation, with much higher values in the northern hemisphere, the ^{137}Cs input showed that South America would be a region appropriate for soil erosion studies using the radiometric technique (García Agudo, 1998), though it is required a ^{137}Cs deposition in the soil great enough to allow its precise quantification. About 25% of the total global fallout has been deposited in the southern hemisphere (UNSCEAR, 1982, UNEP, 1984). ^{137}Cs deposition not only depends on the latitude, but also on the amount of rainfall, since the radionuclide is removed from the atmosphere by rain, snow or dry precipitation (García Agudo, 1998). Several studies have shown that ^{137}Cs deposition is positively correlated with precipitation (Pálsson et al., 2006; Porto et al., 2011). This has been shown along a precipitation-altitudinal gradient from semiarid to alpine environments in N Spain (Navas et al., 2007). Furthermore, physiographic conditions with effect on local rainfall patterns also have an effect on the initial spatial distribution of the ^{137}Cs fallout (Porto et al., 2001).

The potential of ^{137}Cs as tracer of soil erosion has been successfully demonstrated in a wide range of environments (Ritchie et al., 1970; Walling et al., 1986; Pennock and de Jong, 1987; Zhang et al., 1990; Navas and Walling, 1992; Owens et al., 1996; Sadiki et al. 2007; Gaspar et al., 2013; Quijano et al., 2016a). The results of these studies show that the spatial distribution of ^{137}Cs reflect soil redistribution processes and sediment transport, which are controlled by topographic and land use conditions. However, this technique rarely has been used in South America. In Argentina, Buján et al. (2000) and Juri Ayub et al. (2008, 2010) applied fallout ^{137}Cs in the relatively flat central areas of Cuyo and Pampean regions, but no published information is available from the Patagonian Andean Region. The nearest studies were carried out in Chile, at the east from Andes mountain chain, with different edaphoclimatic conditions (Schuller et al., 1997, 2002).

Soils of the Patagonian Andean Region are developed mainly from volcanic ashes and, even worldwide, studies about ^{137}Cs in volcanic soils are rare (Cox and Fankhauser, 1984; Schuller et al., 2002; Sigurgeirsson et al., 2005; Muñoz-Salinas and Castillo, 2018). To fill this gap studies aimed at improving the knowledge on the ^{137}Cs behaviour in volcanic soils are needed, mainly studies in relation to erosion processes, since volcanic soils are highly erodible where the soil remains bare (La Manna et al. 2016). Studies about ^{137}Cs in volcanic soils showed that common soil conservation practice, such as ditches, trenches and afforestation, could result non effective in these soils (Muñoz-Salinas and Castillo, 2018).

Soils formed over volcanic ashes have, as dominant clays, non-crystalline minerals (e.g., allophane, imogolite) which gives distinctive properties to soil (Dahlgren et al. 2004, McDaniel et al. 2012). Caesium-137 has a very short radius of hydration, and it is strongly adsorbed to cation exchange sites of colloids (Tamura, 1964). While clays are

the most important binding agent of ^{137}Cs in mineral soils (Ritchie et al., 1970), the accumulation of organic matter, typical of volcanic soils, may improve ^{137}Cs retention (Sigurgeirsson et al., 2005) as it has been shown the key role of organic matter in fixing the radionuclide (Gaspar et al., 2013). Several physical and chemical soil properties control ^{137}Cs behaviour, such as type of clay minerals, pH, cation exchange capacity (CEC), organic carbon content and texture (Gaspar et al., 2013; Quijano et al., 2016b). Studies on volcanic soils showed that ^{137}Cs is strongly retained in the uppermost few centimetres of soil, by non-crystalline minerals, and that the vertical migration of the radionuclide is controlled by clay content and porosity (Cox and Fankhauser, 1984; Schuller et al., 2002; Sigurgeirsson et al., 2005).

The landscape and soils in the Patagonian Andean region are characterized by a great edaphoclimatic gradient in a short distance (Colmet Daage et al., 1988). The abrupt east-west pluviometric gradient, evidenced by vegetation which changes from forests at the west to steppe towards the east, also controls the pedological processes suffered by volcanic ash (Parfitt et al., 1984). Mineralogical studies support that in the west, allophane is formed, whereas to the east—subhumid sector—soils are enriched in imogolite (non-crystalline) or/and halloysite (1:1 layer silicate) (Besoain, 1985; Colmet Daage et al., 1988). Thus, physical and chemical properties greatly varied in the east-west gradient (Gaitán and López, 2007), which could affect ^{137}Cs behaviour (Tamura, 1964; Gaspar and Navas, 2013). Research done in volcanic soils of Iceland showed that ^{137}Cs was strongly retained by allophane (Sigurgeirsson et al., 2005).

Our study aims to gain knowledge on the spatial distribution of ^{137}Cs in volcanic soils of Andean Patagonia, by measuring reference inventories along an edaphoclimatic gradient and to analyse the factors controlling the variation of ^{137}Cs mass activities and inventories. We pursue to determine the vertical distribution of the radionuclide under

representative land covers and land uses in the region and compare ^{137}Cs patterns of gain and loss associated with different land uses in the subhumid sector, considering the three most representative ones in the region: native forests under grazing, degraded rangeland and pine plantation.

2. MATERIALS AND METHODS

The study was carried out along a longitudinal gradient, at ca. 43° south latitude, from the forest-steppe ecotone in the subhumid sector, up to the Patagonian Andean forests towards the west in Argentina, comprising a mean annual precipitation gradient from 750 mm to 1400 mm, and an edaphic gradient from volcanic soils without non-crystalline minerals in the subhumid sector to allophanized soils in the rainiest area.

Along this gradient, three study areas were selected: Esquel ($71^\circ 23'$ WL; $42^\circ 56'$ SL; 750 mm mean annual precipitation), Trevelin ($71^\circ 28'$ WL; $43^\circ 2'$ SL; 950 mm) and Los Alerces National Park ($71^\circ 45'$ WL; $43^\circ 10'$ SL; 1400 mm) (Figure 1). A pristine, flat and stable site, which has suffered neither erosion nor soil accumulation, was selected for establishing the ^{137}Cs reference inventory in each of the three study areas. In each reference site, sampling points, separated ca. 20-50 m, were selected. We collected, by using a 5.9 diameter cylinder, depth interval soil profiles, where samples were sectioned at 5 cm depth increments, and soil bulk samples. In Esquel, seven points were sampled; of these, six were sectioned at 5 cm depth increments. In Trevelin, four points were sampled; two of which were sectioned. In Los Alerces National Park, five points were sampled; three of which were sectioned. To retain the entire ^{137}Cs profile and taking into account the probable penetration depth of the radionuclide in the region, samples were collected until 25 cm depth in Esquel and Trevelin and until 30 cm depth in Los

Alerces National Park. A pit was dug in each study area, and soil morphological characteristics were recorded according to USDA (Schoeneberger et al., 1998).

In Esquel, the eastern part of the study area, different land uses were also evaluated. Soil samples were taken from two native forest patches, as well as from transects along a pine plantation and a rangeland. Native forests, dominated by *Maytenus boaria* Mol with a shrub stratum, are preserved as isolated patches in an anthropized area, with different degradation conditions. Forest A, currently is a forest patch with close crowns, but in aerial photographs from 1970 it appears as a bushy patch, with only a few high crowns (Figure 2a). Forest B, both at 1970 as at present, corresponds to a small patch of trees in a degraded and eroded area (Figure 2b). Both studied forests are grazed and used as shelter by the cattle. Sampling was carried out in microsites where, currently, the soil is completely covered by litter. Eleven sampling points were considered, including five sectioned at 5 cm depth increments and six bulk samples.

The afforestation (pine plantation) and the rangeland transects were located along two adjacent slopes. The rangeland, used for grazing for the last 100 years, presented grass shrub vegetation (80% covered), dominated by *Rumex acetosella* L., a perennial and invasive exotic herbaceous plant, accompanied by species indicative of degradation and overgrazing such as *Acaena splendens* Hook. & Arn and *Mulinum spinosum* (Cav.) Pers. and grasses of low palatability such as *Pappostipa speciosa* Trin. et Rupr. At the upper part of the slope transect, the rangeland presents native trees of *Maytenus boaria*.

The afforestation corresponded to a 32 years *Pinus radiata* D. Don plantation from 1985, and a *Pinus ponderosa* plantation from 1996 at the upper part of the slope. At the uppermost slope, patches of native vegetation, mainly shrubs, are intermixed with the pines (Figure 2c).

Since the rangeland and the afforestation are located on hill slopes, the sampling was conducted along a transect parallel to the main slope direction. Ten and twelve sampling points placed ca. 25 m apart from each other, were sampled along the transects of the rangeland and plantation (Figure 2c). At each sampling point, a bulk sample was collected at 0-25 cm depth, and in three points (the upper, middle and lower slope) soil samples were sectioned at 5 cm depth increments. Where deposition was supposed to happen, sampling depth was extended until 40 cm depth to ensure that the entire ^{137}Cs profile was retained.

The microtopography of the slopes was evaluated by geometric levelling, using a TOPCON AT-G6 optical level and a 1cm-graduated telescopic sight.

All the soil samples were air-dried, homogenised and passed through a 2 mm for radionuclide gamma assays and physical and chemical properties analyses. The mass activity of ^{137}Cs was measured using a Canberra high-resolution, low-background, hyperpure germanium coaxial gamma detector coupled to an amplifier and multichannel analyser. The detector had a relative efficiency of 50% and a resolution of 1.9 keV (shielded to reduce background), and was calibrated using standard certified soil samples that had the same geometry as the measured samples. Subsamples of 50 g were loaded into plastic containers. Count times over 24 h provided an analytical precision of about $\pm 2\text{-}3\%$ at the 95% confidence level. Mass activities were expressed as Bq kg^{-1} dry soil. Inventories were expressed as activity per unit area (Bq m^{-2}) which was calculated using the weight of the <2 mm fraction as the radionuclide is only fixed in the fine fraction of the soil (Soto and Navas, 2004) and the cross section of the sample (Gaspar and Navas, 2013). The gain/loss soil under different land uses was calculated in each sample point from the difference between ^{137}Cs remaining in the profile and the ^{137}Cs reference level (Navas et al. 2013).

The soil texture was analyzed by Coulter laser granulometer after destruction of organic matter with 10% H₂O₂ at 80°C. Organic matter (SOM) was analyzed by loss on ignition (IRAM-SAGPyA 2008). Bulk density was determined in undisturbed samples, and total porosity was calculated (Taboada and Álvarez, 2008). The pH in NaF 1N (1:50) was measured using pH-meter. pHNaF measured at 2 and 60 min is considered an indicator of non-crystalline aluminosilicates in soils (Fieldes and Perrot, 1966).

Differences in soil properties between the study sites along the longitudinal gradient were analyzed by two-way analysis of variance (ANOVA), considering the study site (3 levels) as factor. Tukey's HSD (Honestly Significant Difference) was applied as post hoc analysis when ANOVA test showed a p-value < 0.05. The relationship between the ¹³⁷Cs inventory and precipitation rate was analysed by linear and non-linear regression models, selecting the model with the best fit. Correlation between ¹³⁷Cs mass activity and soil properties was evaluated by Pearson test.

Vertical distribution of ¹³⁷Cs was evaluated and the correlation between ¹³⁷Cs activity and organic matter content along the soil profiles was evaluated, separately, for reference sites and for the different land uses in Esquel, considering the sectioned samples.

Differences in soil properties between the different land uses at Esquel were analyzed by two-way analysis of variance (ANOVA) and Tukey's HSD. The correlation between ¹³⁷Cs activity and soil properties was also evaluated by Pearson test, considering the bulk samples (0-25 cm) along the slope.

The statistical analyses were carried out with the Infostat software (Di Rienzo et al., 2013).

3. RESULTS

3.1. Reference ^{137}Cs inventory

Table 1 shows morphological features of representative soils in the study area, where Holocene volcanic ash was the main soil parent material. Soils resulted light loose and easily excavated and allowed roots to penetrate to great depths. Along the precipitation gradient, soils were well drained, with medium and coarse textures. In Esquel and Los Alerces National Park, soils showed scarce horizons development, whereas in Trevelin, with finer textured soils, a Bw horizon was recorded.

Along the precipitation gradient, physical and chemical soil properties greatly varied, mainly by the presence of non-crystalline aluminosilicates, which was also related to organic matter content and bulk density (Table 2). The Los Alerces National Park was the only site where soils were rich in allophane. All the soil samples from Los Alerces showed high pH NaF values (i.e., pH NaF 2'greater than 9.2), suggesting the presence of allophane (Irisarri, 2000). At Trevelin, soils were rich in imogolite (i.e., pH NaF 60'greater than 9.2) and at Esquel most samples showed absence of non-crystalline minerals. The organic matter contents increased along the edaphoclimatic gradient. The stable aggregation of volcanic soils resulted in low density and high porosity, the last one increasing in allophanized soils (Table 2).

Textural classes varied from loam to sand, being representatives of dominant soil textures from the study area (La Manna et al., 2018b). In Esquel, soil samples were classified as sandy loam, loamy sand and sand; in Trevelin textures were loam and silt loam; while in Los Alerces National Park textures varied from silt loam to sandy loam.

The reference ^{137}Cs activity and ^{137}Cs inventory both increased progressively in relation to the rising rainfall gradient, reaching in Los Alerces National Park, values that almost were three times higher than at Esquel (Table 3). The ^{137}Cs inventory was

positively correlated with the mean annual precipitation rate, following an exponential relationship, as shown in Figure 3. The allophanic soils of Los Alerces National Park, even having low clay contents (Table 2), showed ^{137}Cs values even higher than the expected by a lineal regression model between ^{137}Cs inventories and annual precipitation (Figure 3).

In the study area, ^{137}Cs was positively associated not only with precipitation but also with pH NaF ($r=0.53$; $p=0.03$), organic matter ($r=0.45$; $p=0.08$) and soil porosity ($r=0.51$; $p=0.04$), soil variables increasing in relation to precipitation rates (Table 2).

3.2. Vertical distribution of ^{137}Cs

In all the study areas, most of ^{137}Cs fallout was retained in the uppermost 10 cm of the soil. As much as 95, 90 and 87% of ^{137}Cs was found in the 0-10 cm depth interval in Esquel, Trevelin and Los Alerces National Park, respectively. An exponential decline of ^{137}Cs with depth was found (Figure 4). However, in Esquel, some profiles showed the highest ^{137}Cs values at 5-10 cm depth, instead of at 0-5 cm.

The ^{137}Cs activity along the soil profiles varied according to organic matter content. Organic matter along the profiles followed the same pattern as ^{137}Cs for the three study areas, showing significant correlation between ^{137}Cs activity and organic matter: $r=0.81$ ($p<0.001$); $r=0.86$ ($p=0.01$); $r=0.93$ ($p<0.001$) for Esquel, Trevelin and Los Alerces National Park, respectively.

The penetration depth of ^{137}Cs varied between the study sites. The greatest penetration was 25 cm in Los Alerces National Park, whereas in Esquel and Trevelin the radionuclide was detected just up to 15 cm depth, with the exception of one site from Esquel, which showed a penetration of 20 cm depth.

3.3. ¹³⁷Cs in Esquel: relationship with land use

Soil properties varied according to the land use. Native forest showed the greatest values of pH NaF, organic matter, porosity and silt granulometric fraction and the lower sand proportion (Table 4). Unlike the other study sites in Esquel, some soils under native forest presented non-crystalline minerals (imogolite), which even differed from reference soils (see Table 2).

Despite these differences in physical and chemical properties, the values of ¹³⁷Cs, both mass activity and inventory, did not differ between the different land uses (Table 5). However, ¹³⁷Cs was significantly lower ($p < 0.05$) in soils from native forest under grazing, pine plantation and rangeland, than in reference sites (see Table 2).

3.3.1. Native forests

The ¹³⁷Cs was totally lost in some samples by erosion, whereas in other points, the ¹³⁷Cs inventories increased reaching values higher than those found in references, indicating soil accumulation from elsewhere (Figure 5). The loss of ¹³⁷Cs, compared to reference inventory, varied between 27 and 100% and 89 and 100% for Forest A and B, respectively. While the gain of ¹³⁷Cs showed a maximum value of 19% in Forest A and 9% in Forest B. These results suggest a net loss of soil in these patches of native forests. Erosion process seems to be more severe in Forest B, which appears as a small patch of trees in a degraded and eroded area (Figure 2b).

The ¹³⁷Cs mass activity was significantly correlated with porosity ($r = 0.69$; $p = 0.03$) and tended to be related with organic matter content ($r = 0.45$; $p = 0.10$).

The vertical distribution of ¹³⁷Cs was similar to the found in reference sites, showing an exponential decline with depth (Figure 6). The penetration depth of ¹³⁷Cs was 20 cm

and 10 cm for Forest A and B, respectively. A significant correlation was found between ^{137}Cs activity and organic matter content (Person's $r = 0.67$; $p < 0.01$).

3.3.2. Plantation and Rangeland

In the highest part of the pine plantation hillside, the ^{137}Cs inventory was similar to those found at the Esquel reference sites (P1 in Figure 7). However, as seen in Figure 7, there were marked differences along the plantation transect, and there was erosion in most sample points. The radionuclide was not detected in many samples. On the other hand, at the bottom slope and in sectors along the hill where the slope degree decreases forming micro depressions, the percentage of ^{137}Cs loss varied between 0.2 and 54%.

The ^{137}Cs mass activity for the bulk samples (0-25 cm depth) was significantly correlated with organic matter content ($r=0.61$; $p=0.03$) and porosity ($r=0.64$; $p=0.02$).

Along the rangeland transect, most samples (7 out of 10) showed negligible values of the radionuclide, that was below the limit of detection (Figure 8). However at the bottom slope (P10 in Figure 8) and at the uppermost slope (P3), where soil is partially protected by shrubs and native trees (Figure 2c), soil loss was lower, ca. a loss of 35-41% of the ^{137}Cs reference inventory. In the highest part of the hillside (P1), radionuclide gain was recorded, with ^{137}Cs inventory around 13% higher than the reference value (Figure 8). The ^{137}Cs activity for the bulk samples (0-25 cm depth) was significantly correlated with porosity ($r=0.84$; $p=0.02$) and tended to be positively related to clay content ($r=0.53$; $p=0.10$).

Along both plantation and rangeland transects, an exponential decline of ^{137}Cs with depth was found in the profiles. The maximum depth reached by ^{137}Cs in the plantation was 10 cm in the profile from the highest part of the hillside (P1), whereas in P3 and P12, ^{137}Cs was found only in 0-5 cm depth samples (Figure 7). On the other hand, the

depth reached by ^{137}Cs in the rangeland was 15 cm in the profile from the highest part of the hillside (P1) and 10 cm in P3 and P10 (Figure 8).

The radionuclide pattern in the soil profiles was significantly related to organic matter content: Pearson's $r=0.85$ ($p<0.001$); $r=0.63$ ($p=0.04$), for plantation and rangeland, respectively.

4. DISCUSSION

4.1. Reference ^{137}Cs inventory

The edaphoclimatic gradient, characteristic of the Patagonian Andean Region (Colmet Daage et al., 1988), was covered in this study, from soils without non-crystalline aluminosilicates in Esquel, soils with imogolite in Trevelin, to allophanized soils in Los Alerces National Park (Table 2). In agreement with increasing in pH NaF (indicator of non-crystalline aluminosilicates) according to precipitation rates, organic matter and soil porosity also increased towards the west.

Allophane and imogolite, under an electron microscope, appear as nanoparticles of hollow spheres and hollow threads or tubules, respectively (Theng and Yuan, 2008). Unique physical attributes of volcanic soils are related to structural assemblages of these hollow spheres and tubules as mineral entities into resilient, progressively larger aggregated domains (McDaniel et al., 2012). This stable aggregation results in low density, high porosity and high surface area, as it was found in the study area (Table 2). According to these properties, cation exchange capacity of volcanic soils is relatively high, often higher than 50 cmolc kg^{-1} (McDaniel et al., 2012). Soil surveys developed in the study area found CEC higher than $100 \text{ cmolc kg}^{-1}$ in allophanized soils, even in coarse textured soils (Morales and La Manna, 2011).

The ^{137}Cs reference inventory varied from 192 to 576 Bq m^{-2} (Table 3), and was positively related to the mean annual precipitation rate, in accordance with several studies both in volcanic (Cox and Fankhauser, 1984; Schuller et al., 2002; Sigurgeisson et al., 2005) and non-volcanic soils (e.g. Blagoeva and Zikovsky, 1995; Pálsson et al., 2006; Navas et al., 2007; Porto et al., 2011). It is known that atmospheric washout, and therefore precipitation, is one of the dominant factors controlling the deposit of radionuclides. The highest values of ^{137}Cs inventory (576 Bq m^{-2}) were found in Andisols from Los Alerces National Park, with 1400 mm of annual precipitation rate, which resulted almost three times higher than those found in Esquel. Studies on volcanic soils from Chile, located at northern latitudes than our study area (38° - 41° South Latitude), recorded ^{137}Cs inventories from 450 to 5410 Bq m^{-2} , highly correlated with annual precipitation rates, which ranged from 750 mm to 4000 mm (Schuller et al., 2002).

The relationship between ^{137}Cs inventory and precipitation followed an exponential model (Figure 3); showing the allophanic soils from the rainiest sector, ^{137}Cs values higher than the expected by a lineal regression model. This result could be related to the presence of allophane as colloidal material in Los Alerces National Park. Studies on volcanic soils of Iceland showed that ^{137}Cs fallout is strongly retained by allophane (Sigurgeisson et al., 2005). Thus, in the studied rainfall gradient, ^{137}Cs in soils was not only related with annual precipitation, but also with key soil properties such as pH NaF, organic matter and porosity, which are soil properties that vary according to pedological processes suffered by ash in relation to the amount of precipitation (Parfitt et al., 1984).

4.2. Vertical distribution of ^{137}Cs

The penetration depth reached by ^{137}Cs varied in the study area, according to rain amount and soil characteristics, reaching maximum values of 20, 15 and 25 cm for Esquel, Trevelin and Los Alerces National Park, respectively. The maximum depth reached by ^{137}Cs along the edaphoclimatic transect could be explained by mean annual precipitation and by soil textures, since high rainfall rates (as in Los Alerces National Park) and coarse textures (as those found in Esquel) favour downward migration of the radionuclide (Schuller et al., 1997). In Trevelin, where soils showed the finest textures with clay contents significantly higher than the others study sites (Table 2), penetration depth was similar or even lower than in Esquel, in spite of increasing precipitation. Several studies show that precipitation, together with texture and coarse-pore volume control the vertical migration of ^{137}Cs (He and Walling, 1996). The penetration depth of ^{137}Cs found in Los Alerces National Park (i.e., 25 cm) agrees with results found by Schuller et al. (2002) in volcanic soils from Chile with similar precipitation, but reaching much higher values, up to 70 cm, where mean precipitation increases up to 4000 mm.

Most of the ^{137}Cs fallout was retained in the uppermost 10 cm of the soil in the three reference study areas (Figure 4). An exponential decline of ^{137}Cs with depth was found in the profiles, being typical of uncultivated areas (Walling and Quine, 1995; Navas et al., 2005; Gaspar and Navas, 2013).

The retention of ^{137}Cs at the first portion of the profile could be attributed to organic matter, since this soil property was highly correlated with ^{137}Cs mass activity in the different reference sites. The key role of organic matter in fixing the radionuclide has been shown elsewhere (e.g. Kim et al., 2006; Gaspar et al., 2013). Organic matter is typically bound by non-crystalline minerals in volcanic soils, and one of their common

characteristics is accumulation of relatively large quantities of organic matter (Warkentin and Maeda, 1980). Both non-crystalline clays and organic matter modify two key soil variables: they increase moisture retention capacity, reducing the amount of water that percolates; and they increase cation-exchange capacity, and thereby, ^{137}Cs adsorption sites. A strong retention of ^{137}Cs at the soil surface was found in volcanic soils from Iceland, which was attributed to the presence of both non-crystalline minerals and organic matter (Sigurgeirsson et al., 2005).

The depth of the maximum concentration of ^{137}Cs in Trevelin and Los Alerces National Park was 0-5 cm. At the Esquel site, instead, some profiles showed the highest values at 0-5 cm and others at 5-10 cm. This result, together with finding ^{137}Cs at greater depth in a soil profile from Esquel, suggest bioturbation processes, since burrow systems by *Ctenomys* sp. are frequent in different natural environments from ecotone of Patagonian Andean region (Tammone et al., 2012). It is known that burrowing animals can affect the ^{137}Cs distribution in soils to greater depths (Müller-Lemans and Van Dorp, 1996).

The depth of the maximum concentration found in our study area, agrees with studies done in the northern hemisphere, where ranged between 0 and 12 cm (Gaspar and Navas, 2013, Jagercikova et al., 2015, Quijano et al., 2016a, Abraham et al., 2018).

4.3. ^{137}Cs patterns of gain and loss: relationship with land use

Both ^{137}Cs mass activity and inventory, significantly differed between the reference soils from Esquel (i.e., pristine, flats and stable sites) and the different land uses (i.e., forest under grazing, pine plantation and rangeland), where marked ^{137}Cs losses were recorded (Tables 3 and 5).

Soil properties varied according to the land use. Native forest showed the greatest values of pH NaF, organic matter and silt granulometric fraction, and unlike the other land uses, some profiles presented non-crystalline minerals (imogolite) (Table 4). Despite these differences in soil properties, the values of ^{137}Cs , both mass activity and inventory, did not significantly differed between the different land uses (Table 5). Most samples showed ^{137}Cs values lower than the reference value for Esquel, suggesting loss of soil as consequence of erosion processes. As much as 72% (8 out of 11), 83% (10 out of 12) and 90% (9 out of 10) of sample points from native forests, plantation and rangeland, respectively, had ^{137}Cs values lower than the reference sites (Figures 5, 7 and 8). Besides, 45%, 58% and 70% of sample points from native forests, plantation and rangeland, respectively, showed ^{137}Cs values below the limit of detection.

The ^{137}Cs values in native forest patches under grazing suggest that erosion/sedimentation processes were active in the last 50 years, showing eroding, but also aggrading profiles. Most eroding profiles have lost between 80 and 100% of ^{137}Cs , whereas gain of ^{137}Cs just ranged between 4 and 19% (Figure 5). The loss of ^{137}Cs was related to the degradation condition of the forest patch, with more critical values in the most degraded Forest B (Figure 2). It is known that ecological and management factors, like canopy cover, litter cover, and forest vertical stratification, control both soil degradation and erosion (Pennock and Van Kessel, 1997, Morales et al. 2013, Altieri et al. 2018). The aggrading profiles, with highest ^{137}Cs mass activity, also showed the highest porosity and organic matter content. Recent studies developed on native forests under grazing in Patagonia suggest that allophanized soils, even under high livestock stocking rates, keep high physical-chemical fertility, which was explained by resilience capacity of volcanic soils (Gomez et al. 2018). Although our results at Esquel site, on

non-allophanized soils, also showed high fertility in native forest under grazing (Table 4), the high loss of ^{137}Cs evidence severe erosion processes.

Along the plantation transect, loss of ^{137}Cs was recorded, except at midslope (P7) and at the highest part of the hillside (P1), where the ^{137}Cs content was similar to the reference value. These results suggest that at P7, a balance between soil loss and gain exists, while uppermost (P1), where slope diminishes and pines coexist with native shrubs, soil appears not to have suffered neither erosion nor soil accumulation (Figure 7). The ^{137}Cs mass activity, varying along the transect, was highly correlated with organic matter content and porosity. Numerous studies have shown that both, loss/gain ^{137}Cs as well as soil properties, are closely related to topography along the hillsides (e.g. Gaspar et al. 2013, Navas et al. 2017).

Along the rangeland transect, the loss of ^{137}Cs ranged between 33 and 100%, respect to reference value (Figure 8). Only at the highest part of the hillside (P1), where slope degree diminishes and soil is protected by shrubs and native trees (Figure 2c), radionuclide gain was recorded, and the ^{137}Cs inventory resulted 13% higher than the reference value (Figure 8). This result suggest that patches of native shrubs and trees have played a role in reducing erosion in this degraded rangeland.

The vertical distribution of ^{137}Cs under the different land uses was similar as in reference sites, and exhibited the characteristic exponential decline of the radioisotope activity with depth of uncultivated areas (Walling and Quine, 1995). Non eroding profiles showed the highest ^{137}Cs penetration depth (Figures 6, 7 and 8). The depth of the maximum concentration of ^{137}Cs was 0-5 cm for all the profiles, and ^{137}Cs activity was significantly correlated with organic matter under the different land uses, agreeing with studies worldwide on different soils (Quine et al., 1994; Mabit et al., 2008; Gaspar et al., 2013; Alewell et al., 2014; Quijano et al., 2016b; Mesrar et al., 2017).

4.4. Erosion rates in Esquel: relationship with land use

Rangelands from the transition (ecotone) between the Andean forests and the Patagonian steppe, as those located in Esquel, have suffered the greatest human pressure and overgrazing by livestock and sheep for most than a century (Veblen and Lorenz, 1988). The great loss of ^{137}Cs found in the rangeland, proving soil loss, agrees with the history of use of these ecotonal areas. However, results also suggest that under both, patches of native forest and pine plantation, soil erosion during the last 50 years was high. The patches of native forests appears as isolated patches in an anthropized area under grazing, both currently and in the past, and erosion indicators can be evidenced in the surroundings, such as bare soils, superficial gravel and tree root exposure (Figure 2).

On the other hand, the pine plantation, with 32 and 21 years old, at bottom and mid-slopes, and at upper slope, respectively, replaced a degraded and shrubby native forest, which had been used for grazing and extracting wood. It is known that pine plantations do not efficiently protect the soil until crowns are quite close and a fresh litter layer is formed (Maestre and Cortina, 2004; Gu et al. 2013), which, in the study area, is reached when the plantations are around 10-13 years old (Tarabini et al. in press).

Both native forest patches and pine plantation were sampled in microsites where, currently, the soil is completely covered by litter. In fact, both vegetation types showed a continuous litter cover, which protects the soil and minimizes the erosion rates (La Manna et al. 2016). Studies on pine plantations based on ^{137}Cs , found minimum values of soil loss in stand with high canopy cover and biomass (Altieri et al., 2018). Thus, ^{137}Cs results are obviously conditioned not only by current use but also by land use in the last decades.

In each land use, ^{137}Cs greatly varied according to microtopography and soil protection, however, the different land uses agree in the existence of many sample sites where ^{137}Cs values were below the limit of detection. Although the adjustment of equations to determine soil loss rates based on radiometric techniques could require a larger sampling of the reference inventory (Sutherland, 1994; Mabit et al., 2002), and radionuclides that allow determinations in shorter terms should be considered (Gaspar et al., 2017), negligible values of ^{137}Cs suggest the loss of at least 15 cm of soil in the last 50 years in many sample points under the different land uses. This value equals $30 \text{ m}^3 \text{ ha}^{-1} \text{ year}^{-1}$. According to bulk density values found under different land uses (Table 4), erosion rates in Esquel reached values of at least $22 \text{ t ha}^{-1} \text{ year}^{-1}$ in native forests under grazing and $33 \text{ t ha}^{-1} \text{ year}^{-1}$ in plantation and rangeland slopes.

Setting tolerable erosion rates requires taking into account reference rates of soil formation, resulting from mineral weathering as well as dust deposition (Verheijen et al., 2009). Recent studies assessed the rates of soil formation in pine plantations from Esquel, by entrapping particles, finding mean rates of 0.1 mm year^{-1} or $1 \text{ m}^3 \text{ ha}^{-1} \text{ year}^{-1}$, equivalent to $0.9 \text{ t ha}^{-1} \text{ year}^{-1}$ (Tarabini et al. in press). These soil formation rates, similar to those assessed in temperate regions of Europe (Verheijen et al., 2009) highly differed from the erosion rates estimated in our study. These results highlight the requirement of reliable quantification of soil erosion rates in Patagonia and the urgent need of applying effective soil conservation measures.

5. CONCLUSIONS

This work is the first approach to the use of radioisotopes in relation to erosion in the Andean Patagonia region of Argentina. The ^{137}Cs reference inventories varied according to the edaphoclimatic gradient. Allophanized soils showed the highest ^{137}Cs

values, even higher than the expected according to the precipitation amount. This result suggests a high adsorption of ^{137}Cs by allophanized soils, related to the non-crystalline nature of the clays and to the high content of organic matter.

In the subhumid study area, high erosion rates were found under different land uses. These results emphasize the high erodability of non-allophanized volcanic soils, which seem to suffer severe loss of soil under any land use, once the original vegetation and soil cover are lost.

Reference inventory values and the loss of the radionuclide in sites with high anthropic intervention, show the potential for using ^{137}Cs measurements for assessing erosion processes in these latitudes.

6. ACKNOWLEDGEMENTS

This research was funded by PICT-FONCyT 1715/10, CGL2014-52986-R and a Fellowship from Carolina Foundation. We acknowledge M. Calderón for his help with topographic fieldwork.

7. REFERENCES

- Abraham, J., Meusburger, K., Waldis, J. K., Ketterer, M. E., and Zehringer, M., 2018. Fate of ^{137}Cs , ^{90}Sr and $^{239+240}\text{Pu}$ in soil profiles at a water recharge site in Basel, Switzerland, *J. Environ. Radioactiv.* 182, 85–94.
- Alewell, C., Meusburger, K., Juretzko, G., Mabit, L., Ketterer, M.E., 2014. Suitability of $^{239+240}\text{Pu}$ and ^{137}Cs as tracers for soil erosion assessment in mountain grassland. *Chemosphere* 103, 274 – 280.
- Altieri, V., De Franco, S., Lombardi F., Marziliano P.A., Menguzzato, G. and Porto, P., 2018. The role of silvicultural systems and forest types in preventing soil erosion

570 processes in mountain forests: a methodological approach using cesium-137
 571 measurements. *J. Soils Sediment.* 18(12), 3378–3387.

572 Besoain, E., 1985. *Mineralogía de arcillas de suelos*. Instituto Interamericano de
 573 Cooperación para la Agricultura CATIE, pp. 1205.

574 Blagoeva, R., Zikovsky, L., 1995. Geographic and vertical distribution of Cs-137 in
 575 soils in Canada. *J. Environ. Radioactiv.* 27, 269–274.

576 Buján, A., Santanatoglia, O.J., Chagas, C., Massobrio, M., Castiglioni, M., Yáñez, M.S.,
 577 Ciallella, H., Fernández, J., 2000. Preliminary study on the use of the ¹³⁷Cs method
 578 for soil erosion investigation in the pampean region of Argentina. *Acta Geol. Hisp.*
 579 35(3–4), 271–277.

580 Carabelli, F., Scoz, R., 2016. Human-induced alterations in native forests of Patagonia,
 581 Argentina, in: La fortaleza, R., Sanesi, G., Chen, J., Crow, T.R. (Eds.), *Patterns and*
 582 *Processes in Forest Landscapes*. Springer, Dordrecht, pp. 89–105.

583 Colazo, J.C., Carfagno, P., Gvozdenovich, J., Buschiazzo, D., 2019. Soil erosion, in:
 584 Rubio, G., Lavado, R., Pereyra, J. (Eds.), *The soils of Argentina*. World Soils Book
 585 Series. Elsevier. p. 239–250.

586 Colmet Daage, F., Marcolin, A., Lopez, C., Lanciotti, M.L., Ayesa, J., Bran, D.,
 587 Andenmatten, E., Broquen, P., Girardin, J.L., Cortes, G., Irisarri, J., Besoain, E.,
 588 Sadzawka, G., Sepúlveda, S., Massaro, S., Millot, P., Bouleau, P., 1988.
 589 Características de los suelos derivados de cenizas volcánicas de la cordillera y
 590 precordillera del Norte de la Patagonia. Bariloche. Convenio INTA – ORSTOM. S.
 591 C. de Bariloche, Río Negro. 167, 39 pp.

592 Cox, M. E., Fankhauser, B. L., 1984. Distribution of fallout cesium-137 in Hawaii.
 593 *Health Phys.* 46(1), 65–71.

594 Dahlgren, R.A., Saigusa, M., Ugolini, F.C., 2004. The nature, properties and
 595 management of volcanic soils. *Adv. Agron.* 82, 113–182.

596 Di Rienzo, J.A., Casanoves, F., Balzarini, M.G., González, L., Tablada, M., Robledo,
 597 C.W., 2013. InfoStat versión 2013. Grupo InfoStat, FCA, UNC, Argentina.
 598 <http://www.infostat.com.ar>.

599 Fieldes, M.P., Perrot, K.W., 1966. The nature of allophane in soils. Part 3. Rapid field
 600 and laboratory test for allophane. *New Zeal. J. Sci.* 9(3), 623–629.

601 Gaitán, J.J., López, C., 2007. Análisis del gradiente edáfico en la región
 602 Andinopatagónica. *Ciencia del Suelo* 25(1), 53–63.

603 García Agudo, E., 1998. Global distribution of ^{137}Cs inputs for soil erosion and
 604 sedimentation studies, in: International Atomic Energy Agency (Ed.), Use of ^{137}Cs
 605 in the Study of Soil Erosion and Sedimentation. IAEA-TECDOC-1028, Vienna,
 606 Austria, pp. 117–121.

607 Gaspar, L., Navas, A., 2013. Vertical and lateral distributions of ^{137}Cs in cultivated and
 608 uncultivated soils on Mediterranean hillslopes. *Geoderma*, 207–208, 131–143.

609 Gaspar, L., Navas, A., Walling, D.E., Machín, J., Gómez Arozamena, J., 2013. Using
 610 ^{137}Cs and $^{210}\text{Pb}_{\text{ex}}$ to assess soil redistribution on slopes at different temporal scales.
 611 *Catena* 102, 46–54.

612 Gomez, F., Tarabini, M., von Müller, A., La Manna, L., 2018. Propiedades edáficas en
 613 bosques de ñire bajo uso silvopastoril en la provincia del Chubut. *Proceedings IV*
 614 *Congreso Nacional de Sistemas Silvopastoriles*. Villa La Angostura, Neuquén 535–
 615 540.

616 Gu, Z. J., Wu, X. X., Zhou, F., Luo, H., Shi, X. Z., Yu, D. S., 2013. Estimating the
 617 effect of *Pinus massoniana* lamb plots on soil and water conservation during rainfall
 618 events using vegetation fractional coverage. *Catena* 109, 225–233

He, Q., Walling, D.E., 1996. Interpreting particle size effects in the adsorption of ^{137}Cs and unsupported ^{210}Pb by mineral soils and sediments. *J. Environ. Radioactiv.* 30(2), 117–137.

IRAM-SAGPyA 29571-1. 2008. Calidad Ambiental-Calidad del suelo. Determinación de materia orgánica en suelos. Parte 1, Método de pérdida de masa por calcinación.

Irisarri, J., 2000. La propuesta de reclasificación de los Andepts de Argentina, de acuerdo al Orden Andisoles. In *Workshop Soil Taxonomy*. Instituto Nacional de Tecnología Agropecuaria, AICET, Asociación Argentina de la Ciencia del Suelo 18–27.

Jagercikova, M., Cornu, S., Bourlès, D., Antoine, P., Mayor, M., Guillou, V., 2015. Understanding long-term soil processes using meteoric ^{10}Be : A first attempt on loessic deposits. *Quat. Geochronol.* 27. 11–21.

Juri Ayub, J., Velasco, R.H., Rizzotto, M., Quintana, E., Aguiar, J., 2008. ^{40}K , ^{137}Cs and ^{226}Ra soil and plant content in semi-natural grasslands of Central Argentina, in: Paschoa, A.S. (Ed.), *The Natural Radiation Environment 8th International Symposium*. American Institute of Physics.

Juri Ayub, J., Rizzotto, M., Toso, J.P., Velasco, R.H., 2010. ^{137}Cs deposition and vertical migration in soils from Argentina. *Proceeding International Conference on Environmental Radioactivity Viena, Austria*, 1–5.

Kim, Y., Cho, S., Kang, H.D., Kim, W., Lee, H.R., Doh, S.H., Kim, K., Yun, S.G., Kim, D.S., Young, G., 2006. Radiocesium reaction with illite and organic matter in marine sediment. *Mar. Pollut. Bull.* 52, 659–665.

Lal, R., 2017. Soil erosion by wind and water: Problems and prospects, in : *Soil and Water Conservation Society and St. Lucie Press (Eds.), Soil erosion Research Methods*, New York, pp. 1–352.

644 La Manna, L., Buduba, C.G., Rostagno, C.M., 2016. Soil erodibility and quality of
645 volcanic soils as affected by pine plantations in degraded rangelands of NW
646 Patagonia. *Eur. J. For. Res.* 135(4), 643–655.

647 La Manna, L., Gaspar, L., Rostagno, C. M., Quijano, L., Navas, A., 2018a. Soil changes
648 associated with land use in volcanic soils of Patagonia developed on dynamic
649 landscapes. *Catena* 166, 229–239.

650 La Manna, L., Tarabini, M., Gómez, F., Noli, P. A., 2018b. Estimación de la capacidad
651 de retención de agua de suelos volcánicos en función de variables de fácil
652 determinación a campo. *Ciencia del Suelo* 36(1), 23–29.

653 Lizaga, I., Quijano, L., Leticia Gaspar, L., Navas, A., 2018. Estimating soil
654 redistribution patterns with ^{137}Cs measurements in a Mediterranean mountain
655 catchment affected by land abandonment. *Land Degrad. Dev.* 29, 105–117.

656 Lizaga, I., Quijano, L., Gaspar, L., Ramos, M.C., Navas, A., 2019. Linking land use
657 changes to variation in soil properties in a Mediterranean mountain agroecosystem.
658 *Catena*, 172, 516–527

659 Mabit, L., Bernard, C., Laverdière, M.R., 2002. Quantification of soil redistribution and
660 sediment budget in a Canadian watershed from fallout caesium-137 (^{137}Cs) data.
661 *Can. J. Soil Sci.* 82, 423–31.

662 Mabit, L., Bernard, C., Makhoul, M., Laverdière, M.R., 2008. Spatial variability of
663 erosion and soil organic matter content estimated from ^{137}Cs measurements and
664 geostatistics. *Geoderma* 145, 245–251

665 Mesrar H, Sadiki A, Faleh A, Quijano L, Gaspar L, Navas A., 2017. Vertical and lateral
666 distribution of fallout ^{137}Cs and soil properties along representative toposequences
667 of central Rif, Morocco. *J. Environ. Radioactiv.* 169–170, 27–39.

668 Maestre, F., Cortina, J., 2004. Are *Pinus halepensis* plantations useful as a restoration
669 tool in semiarid Mediterranean areas? *For Ecol Manage* 198, 303–317

670 McDaniel, P.A., Lowe, D.J., Arnalds, O., Ping, C.L., 2012. Andisols, in: P.M. Huang,
671 Y. Li, M.E. Sumner (Eds.). *Handbook of Soil Sciences: Properties and Processes*.
672 pp. 29–48.

673 Morales, D., La Manna, L., 2011. La fertilidad química del suelo y el mal del ciprés en
674 Patagonia, Argentina. *Ciencia del Suelo* 29(1): 151-160.

675 Morales, D., Rostagno, M., La Manna L., 2013. Runoff and erosion from volcanic soils
676 affected by fire: the case of *Austrocedrus chilensis* forests in Patagonia, Argentina.
677 *Plant Soil* 370, 367–380.

678 Müller-Lemans, H., Van Dorp, F., 1996. Bioturbation as a mechanism for radionuclide
679 transport in soil, relevance of earthworms. *J. Environ. Radioactiv.* 31, 7–20

680 Muñoz-Salinas, E., Castillo, M., 2018. Assessing conservation practices in Amalacaxco
681 Gorge (Izta-Popo National Park, Central Mexico) using fallout ¹³⁷Cs and Optically
682 Stimulated Luminescence (OSL). *J. Mount. Sci.* 15(3): 447-460.

683 Navas, A., Walling, D., 1992. Using caesium-137 to assess sediment movement in a
684 semiarid upland environment in Spain, in: Walling, D.E., Davies, T.R., Hasholt, B.
685 (Eds.), *Erosion, Debris Flows and Environment in Mountain Regions*. International
686 Association Hydrological Sciences (IAHS) Publ. n° 209, 129–138.

687 Navas, A., Machín, J., Soto, J., 2005. Assessing soil erosion in a Pyrenean mountain
688 catchment using GIS and fallout ¹³⁷Cs. *Agr. Ecosyst. Environ.* 105(3), 493–506.

689 Navas, A., Walling, D.E., Quine, T., Machín, J., Soto, J., Domenech, S., López-Vicente,
690 M., 2007. Variability in ¹³⁷Cs inventories and potential climatic and lithological
691 controls in central Ebro valley, Spain. *J. Radioanal. Nucl. Ch.* 274 (2), 331–339.

692 Navas, A., Machín, J., Beguería, S., López-Vicente, M., Gaspar, L., 2008. Soil
693 properties and physiographic factors controlling the natural vegetation re-growth in
694 a disturbed catchment of the Central Spanish Pyrenees. *Agroforest. Syst.* 72(3),
695 173–185.

696 Navas, A., López-Vicente, M., Gaspar, L., Machín, J. 2013. Assessing soil
697 redistribution in a complex karst catchment using fallout ^{137}Cs and GIS.
698 *Geomorphology*. 196, 231–241.

699 Navas, A., López-Vicente, M., Gaspar, L., Palazón, L., Quijano, L., 2014. Establishing
700 a tracer-based sediment budget to preserve wetlands in Mediterranean mountain
701 agroecosystems (NE Spain). *Sci. Total Environ.* 496, 132–143.

702 Navas, A., Quine, T.A., Walling, D.E., Gaspar, L., Quijano, L., Lizaga, I., 2017.
703 Relating intensity of soil redistribution to land use changes in abandoned Pyrenean
704 fields using fallout caesium-137. *Land Degrad. Dev.* 28, 2017–2029

705 Owens, P.N., Walling, D.E., He, Q., 1996. The behaviour of bomb-derived caesium-137
706 fallout in catchment soils. *J. Environ. Radioactiv.* 32(3), 169–191.

707 Pálsson S.E., Howard, B.J., Wright, S.M., 2006. Prediction of spatial variation in global
708 fallout of ^{137}Cs using precipitation. *Sci. Total Environ.* 367, 745–756.

709 Parfitt, R.L., Saigusa, M., Cowie, J.D., 1984. Allophane and halloysite formation in a
710 volcanic ash bed under differing moisture conditions. *Soil Sci.* 138, 360–364.

711 Pennock, D.J., de Jong, E., 1987. The influence of slope curvature on soil erosion and
712 deposition in hummock terrain. *Soil Sci.* 144(3), 209–217.

713 Pennock, D.J., van Kessel, C., 1997. Clear-cut forest harvest impacts on soil quality
714 indicators in the mixedwood forest of Saskatchewan, Canada. *Geoderma* 75, 13–32.

715 Porto, P., Walling, D.E., Ferro, V., 2001. Validating the use of caesium-137
 716 measurements to estimate soil erosion rates in a small drainage basin in Calabria,
 717 Southern Italy. *J. Hydrol.* 248, 93–108.

718 Porto, P., Walling, D., Callegari, G., 2011. Using ^{137}Cs measurements to establish
 719 catchment sediment budgets and explore scale effects. *Hydrol. Process.* 25: 886–
 720 900.

721 Quijano, L., Beguería, S., Gaspar, L., Navas, A., 2016a. Estimating erosion rates using
 722 ^{137}Cs measurements and WATEM/SEDEM in a Mediterranean cultivated field.
 723 *Catena*, 138, 38–51.

724 Quijano, L., Gaspar, L., Navas, A., 2016b. Spatial patterns of SOC, SON, ^{137}Cs and soil
 725 properties as affected by redistribution processes in a Mediterranean cultivated field
 726 (Central Ebro Basin). *Soil Till. Res.* 155, 318–328.

727 Quine, T., Navas, A., Walling, D.E., Machín, J., 1994. Soil erosion and redistribution on
 728 cultivated and uncultivated land near Las Bardenas in the Central Ebro River Basin,
 729 Spain. *Land Degrad. Rehabil.* 5, 41–55.

730 Ritchie, J.C., McHenry, J.R., Gill, A.C., Hawks, P.H., 1970. Distribution of fallout
 731 cesium-137 in sediment profiles. *Health Phys.* 19, 334.

732 Sadiki, A., Faleh, A., Navas, A., Bouhlassa, S., 2007. Assessing soil erosion and control
 733 factors by the radiometric technique in the Boussouab catchment, Eastern Rif,
 734 Morocco. *Catena* 71 (1), 13–20.

735 Schoeneberger, P.J., Wysocky, D.A., Benham, E.C., Broderson, W.D., 1998. Field book
 736 for describing and sampling soils. Natural Resources Conservation Service, USDA,
 737 National Soil Survey Center, Lincoln.

738 Schuller, P., Ellies, A., Handl, J., 1997. Influence of climatic conditions and soil
 739 properties on ^{137}Cs vertical distribution in selected Chilean soils. *J. Plant Nutr. Soil*
 740 *Sci.* 160,423–426.

741 Schuller, P., Voigt, G., Handl, J., Ellies, A., Oliva, L. 2002. Global weapons' fallout
 742 ^{137}Cs in soils and transfer to vegetation in south-central Chile. *J. Environ.*
 743 *Radioactiv.* 62(2), 181–193.

744 Sigurgeirsson, M. A., Arnalds, O., Palsson, S. E., Howard, B. J., Gudnason, K., 2005.
 745 Radiocaesium fallout behaviour in volcanic soils in Iceland. *J. Environ. Radioactiv.*
 746 79(1), 39–53.

747 Soto, J., Navas, A. 2004. A model of ^{137}Cs activity profile for soil erosion studies in
 748 uncultivated soils of Mediterranean environments. *J. Arid Environ.* 59, 719–730.

749 Sutherland, R.A., 1994. Spatial variability of ^{137}Cs and the influence of sampling on
 750 estimates of sediment redistribution. *Catena* 21, 57–71

751 Taboada, M., Alvarez, C.R., 2008. Fertilidad física de los suelos, second ed. Facultad de
 752 Agronomía. Universidad de Buenos Aires, Argentina.

753 Tammone, M., Lacey, E., Relva, M. A., 2012. Habitat use by colonial tuco-tucos
 754 (*Ctenomys sociabilis*): specialization, variation, and sociality. *J. Mammal.* 93(6),
 755 1409–1419.

756 Tamura, T., 1964. Consequences of activity release: selective sorption reactions of
 757 cesium with soil minerals. *Nucl. Safety* 5, 262–268.

758 Tarabini, M., Gomez, F., La Manna L. Ceniza volcánica reciente como indicadora de
 759 retención de partículas en los suelos de la Patagonia andina. *Ciencia Suelo* (In
 760 press).

761 Theng, B. K.G., Yuan, G., 2008. Nanoparticles in the Soil Environment. *Elements* 4(6),
 762 395–399.

763 UNEP, 1984. Prospects for Global Ocean Pollution Monitoring. United Nations
 764 Environment Programme Regional Seas Reports and Studies Number 47. Geneva:
 765 Switzerland.

766 UNSCEAR, 1982. United Nations. Ionizing Radiation: Sources and Biological Effects.
 767 United Nations Scientific Committee on the Effects of Atomic Radiation. Report to
 768 the General Assembly E.82.IX.8 United Nations, New York.

769 Veblen, T.T., Lorenz, D.C., 1988. Recent vegetation changes along the forest/steppe
 770 ecotone of northern Patagonia. *Ann. Am. Assoc. Geogr.* 78, 93–111.

771 Veblen, T.T., Kitzberger, T., Villalba, R., Donnegan, J., 1999. Fire history in northern
 772 Patagonia: The roles of humans and climatic variation. *Ecol. Monogr.* 69(1), 47–67.

773 Verheijen, F.G.A., Jones, R.J.A., Rickson, R.J., Smith, C.J., 2009. Tolerable versus
 774 actual soil erosion rates in Europe. *Earth Sci. Rev.* 94, 23–28.

775 Walling, D.E., Bradley, S.B. Wilkinson, C.J., 1986. A caesium-137 budget approach to
 776 the investigation of sediment delivery from a small agricultural drainage basin in
 777 Devon, UK. International Association of Hydrological Sciences Publication, No.
 778 159, 423–435.

779 Walling, D.E., Quine, T.A., 1995. The use of fallout radionuclides in soil erosion
 780 investigations. *Nuclear Techniques in Soil – Plant Studies for Sustainable*
 781 *Agriculture and Environmental Preservation. Publication ST1/PUB/947.*
 782 International Atomic Energy Agency, Vienna, pp. 597–619.

783 Warkentin, B.P., Maeda, T., 1980. Physical and mechanical characteristics of Andisols,
 784 in B.K.G. Theng (Ed.), *Soils with variable charge*. New Zeal. Soc. Soil Sci. Lower
 785 Hutt, pp. 281–301 New Zealand.

Zhang, X., Higgitt, D.L., Walling, D.E., 1990. A preliminary assessment of the potential for using caesium-137 to estimate rates of soil erosion in the Loess Plateau of China. Hydrol. Sci. J. 35, 267–276.

FIGURES

Figure 1. Location of the study area.

Figure 2. Location of the soil profiles sampled at the study site 1 - *Esquel*: a) Forest A,
b) Forest B and C) along two transects.

Figure 3. Relationship between ^{137}Cs inventory (Bq m^{-2}) and precipitation for the study site 1 - *Esquel*, study site 2 - *Trevelin* and study site 3 - *Los Alerces National Park*.

Figure 4. Mean values of ^{137}Cs mass activity (Bq kg^{-1}) and Inventory (Bq m^{-2}) for sectioned soil profiles at the study site 1 - *Esquel*, study site 2 - *Trevelin* and study site 3 - *Los Alerces National Park*.

Figure 5. Inventories of ^{137}Cs (Bq m^{-2}) for soil profiles in forest A and forest B at the study site 1 - *Esquel*.

Figure 6. Mean values of ^{137}Cs mass activity (Bq kg^{-1}) for sectioned soil profiles at the study site 1 - *Esquel* in forest A and forest B.

Figure 7. Topography of the plantation transect, ^{137}Cs inventories (Bq m^{-2}) for the twelve soil profiles (placed ca. 25m apart) and detailed information of ^{137}Cs mass activity (Bq kg^{-1}) for sectioned soil profiles 1, 3 and 12.

Figure 8. Topography of the rangeland transect, ^{137}Cs inventories (Bq m^{-2}) for the ten soil profiles (placed ca. 25m apart) and detailed information of ^{137}Cs mass activity (Bq kg^{-1}) for sectioned soil profiles 1, 3 and 10.

TABLES

Table 1. Morphological soil properties of the reference sites along a precipitation gradient in the Patagonian Andean region.

Horizon	Depth (cm)	Color ^a	Texture ^b	Structure ^c	Roots ^d	Rock fragments (%)	NaF reaction	Boundary ^e
Sampling Site 1 - Esquel – Precipitation: 750 mm - Mollisol								
A	0 - 25	10YR 3/3	Ls	1 SBK F	2 F	15	-	C W
AC	25 - 70	10YR 3/4	Ls	1 SBK M	2 F	15	-	G W
C	70 - 90+	10YR 3/4	S	1 SBK CO	2 M	20	-	
Sampling Site 2 - Trevelin - Precipitation: 950 mm - Andisol								
A	0 - 23	10YR 3/2	Ls	2 GR F	2 F	0	+	C S
ABw	23 - 48	10YR 4/3	Ls	1 SBK M	3 M	0	+	G S
Bw	48 - 75	10YR 4/4	L	1 GR M	2 CO	10	+	C W
C	75 - 110+	10YR 5/4	scl GX	0 MA	1 CO	70	+	
Sampling Site 3 - Los Alerces National Park - Precipitation: 1400 mm - Andisol								
A	0 - 16	10YR 3/2	SI	0 SGR	3 F	0	+	C W
AC	16 - 45	10YR 3/4	Ls	0 SGR	3 CO	0	+	C S
C1	45 - 75	10YR 4/4	S	0 MA	1 M	0	+	G S
C2	75 - 125+	10YR 5/4	S	0 MA	1 M	0	+	

^a Color: Colors are under moist condition according to Munsell soil color charts (1990).

^b Texture: L = loam, S = sand, Ls = loamy sand, SI = sandy loam, scl = sandy clay loam. GX = extremely gravelly.

^c Structure: 0 = structureless, 1 = weak, 2 = moderate; MA = massive, GR = granular, SGR = single grain SBK = subangular blocky; F = fine, M = medium, CO: coarse.

^d Roots: 1 = few, 2 = common, 3 = many; F = fine, M = medium, CO = coarse.

^e Horizon boundary: A = abrupt, C = clear, G = gradual, S = smooth, W = wavy.

Table 2. Physical and chemical soil properties of the reference sites along a precipitation gradient in the Patagonian Andean region. Mean values and standard errors for 0-25 cm depth are shown. Different letters indicate significant differences between the study areas.

	Esquel	Trevelin	Los Alerces National Park
Precipitation (mm)	750	950	1400
pH NaF 2'	8.0 ± 0.1 c	8.6 ± 0.3 b	10.3 ± 0.03 a
pH NaF 60'	8.8 ± 0.2 c	9.5 ± 0.3 b	11.2 ± 0.04 a
Non crystalline minerals	----	Imogolite	Allophane
SOM (%)	8.8 ± 1.1 b	13.4 ± 1.5 a	14.0 ± 0.2 a
Bulk density (gcm ⁻³)	0.82 ± 0.11 a	0.72 ± 0.07 ab	0.63 ± 0.03 b
Total porosity (%)	69.1 ± 4.0 a	72.8 ± 2.5 ab	76.2 ± 1.0 b
Clay (%)	4.7 ± 0.8 b	8.6 ± 0.9 a	4.7 ± 0.3 b
Silt (%)	29.2 ± 4.0 b	52.2 ± 1.8 a	46.9 ± 1.3 a
Sand (%)	66.1 ± 4.6 a	39.2 ± 1.9 b	48.4 ± 1.5 b

Table 3. ^{137}Cs inventories and mass activities in soils from the Patagonia Andean region. Values for 0-25 depth are shown. Different letters indicate significant differences between the study areas.

^{137}Cs Inventory (Bq m^{-2})			
	Esquel	Trevelin	Los Alerces National Park
n	7	4	5
Mean	192.25 a	266.71 b	576.08 c
Median	189.23	268.99	579.26
Standard deviation	18.16	22.06	38.28
Standard error	6.86	11.03	17.12
Value range	174.62-219.39	240.89-287.99	523.92-620.44

^{137}Cs Activity (Bq Kg^{-1})			
	Esquel	Trevelin	Los Alerces National Park
n	7	4	5
Mean	6.22 a	9.17 b	19.72 c
Median	5.45	9.17	19.72
Standard deviation	2.17	1.13	1.19
Standard error	0.89	0.80	0.84
Value range	3.40–8.84	8.37-9.97	18.88-20.56

Table 4. Physical and chemical soil properties under different land uses at the study site 1 - Esquel. The properties are summarised for 0-25cm depth. Mean values and standard errors are shown. Different letters indicate significant differences between land uses.

	Native forest	Plantation	Rangeland
pH NaF 2'	8.5 ± 0.1 b	8.0 ± 0.1 a	7.9 ± 0.1 a
pH NaF 60'	9.0 ± 0.1 b	8.6 ± 0.1 a	8.5 ± 0.1 a
Non crystalline minerals	None - Imogolite	---	---
SOM (%)	13.5 ± 1.2 b	4.2 ± 0.3 a	4.8 ± 0.2 a
Bulk density (g cm ⁻³)	0.74 ± 0.11 a	1.10 ± 0.09 b	1.09 ± 0.05 b
Total porosity (%)	72.0 ± 4.0 b	58.4 ± 3.5 a	58.9 ± 1.8 a
Clay (%)	4.4 ± 0.5 ab	3.2 ± 0.6 a	5.9 ± 0.6 b
Silt (%)	44.9 ± 4.5 b	16.8 ± 2.8 a	27.3 ± 2.2 a
Sand (%)	50.7 ± 4.6 a	80.0 ± 3.3 c	66.8 ± 2.7 b

Table 5. ^{137}Cs inventories and mass activities in soils under different land uses at the study site 1 - Esquel. Values are summarised for 0-25 cm depth. Mean values and standard errors are shown. Same letters indicate non-significant differences between land uses.

^{137}Cs Inventory (Bq m^{-2})			
	Native forest	Plantation	Rangeland
n	11	12	10
Mean	75.95 ± 29.18 a	54.14 ± 23.22 a	46.00 ± 24.89 a
Median	21.15	0	0
Value range	0-229.03	0-193.61	0-217.98

^{137}Cs activity (Bq Kg^{-1})			
	Native forest	Plantation	Rangeland
n	11	12	10
Mean	2.21 ± 0.96 a	0.86 ± 0.37 a	1.06 ± 0.57 a
Median	0.41	0.19	0
Value range	0–8.36	0-3.40	0-4.83