

Greening and Browning in a Climate Change Hotspot: the Mediterranean Basin

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Abstract

To improve predictions of the future of ecosystems in a changing world it is necessary to consider fine-scale processes. We propose that for the Mediterranean region (a hotspot of climate change and biodiversity) there are three local processes that have often been overlooked in predictive models and are key to understanding vegetation changes: rural abandonment that increases wildlands; population changes that boost fire ignitions; and coastal degradation that enhances drought. These processes are not directly driven by global warming, and act in different directions (greening and browning). The current balance is still towards greening, as land abandonment is buffering the browning drivers; however, it is likely to switch with increasing warming. The challenge is to mitigate the browning processes given that climatic warming is not directly driving these processes, and local management can make a difference in reducing the overall impact on the landscape and society.

Keywords: global change, climate change, water cycling, Mediterranean Basin, Mediterranean ecosystems, wildfires, land abandonment, local scale processes

Introduction

Our planet is going through an unprecedented change that affects all ecosystems and organisms. The magnitude of these effects and the mechanisms involved vary among ecosystems because of differing environmental conditions, varying past and present land-uses, and differing historical and evolutionary legacies. Only by acquiring a detailed understanding of the diversity of processes operating at different scales and in different regions can we improve predictions for the future of our planet (Taylor et al. 2017, Tierney et al. 2017). Here we focus on the paleartic Mediterranean region (Table 1): which is a biogeographical, environmental, and historical unit – as well as a biodiversity hotspot (Myers 2000, Thompson 2005, Keeley et al. 2012). By the Mediterranean region we refer to the land with a mediterranean climate that is around the Mediterranean Sea (including all the islands; Fig. 1). It is a large and highly populated area and a top tourist and retirement destination. The climate dynamics of this region are strongly determined by the enclosed sea (the world's largest inland sea) and the mountains that surround it; these characteristics generate meso-scale atmospheric processes with consequences far beyond the region. Environmental disruptions in the Mediterranean region have consequences in this large densely populated region, as well as in the whole catchment area (which exceeds the Mediterranean climate zone, see Fig. 1), and elsewhere through atmospheric teleconnections (Gangoiti et al. 2006, Park et al. 2016, Ciarlo & Aquilina 2016).

Anthropogenic increases in greenhouse gases and the associated increase in temperature are affecting all ecosystems including those in the Mediterranean region (IPCC 2014). The region is considered a primary climate change hotspot (Giorgi 2006, Giorgi and Lionello 2008) as the warming of this region is above the global average (~ 1.3°C versus 0.85°C worldwide; Guiot & Cramer 2016). There is abundant literature documenting recent climatic changes (e.g., Bolle 2003, Pausas 2004, Mariotti et al. 2008, Lelieveld et al. 2012, Skliris et al. 2018); however, this paper focuses on some local drivers that are more specific (although not exclusive) to the region, and that have often been overlooked in global and regional vegetation models. For instance, Guiot and Cramer (2016) suggest that a warming of 2°C above pre-industrial levels in the Mediterranean region (i.e. the threshold proposed by the 2015 Paris Agreement) would generate vegetation changes of a magnitude unprecedented over the past 10,000 years. This is important because the region is currently near this threshold (Guiot and Cramer 2016). The results were obtained by comparing past (pollen-reconstructed), present, and future biome distribution using a simulation model. Predictive models were based on the main drivers of global change, i.e., temperature (monthly) and CO₂ (annual), simulated at a coarse scale (grid of 2° latitude x 4° longitude; ~ 222 x 346 km).

Our hypothesis is that to increase the accuracy of vegetation change predictions we need to integrate processes operating at a smaller scale. Specifically, we propose that the main current dynamic drivers in the Mediterranean region are: rural (land) abandonment; wildfires driven by

human ignitions; and coastal degradation. These often-overlooked drivers are mostly caused by local socio-economic and human changes not directly linked to global warming; and they are key to understanding the disruption of the natural fire and drought regimes in Mediterranean ecosystems (Fig. 2; mainly, increasing frequency and intensity). Although other drivers of change remain very important, we propose that these three drivers have important consequences for understanding and predicting current and future vegetation changes.

The greening: rural abandonment provides biomass and fuel

Many ecoregions worldwide are suffering from an increase in fragmentation accompanied by a reduction in the area covered by natural ecosystems. Others show the opposite pattern, as is the case of ecosystems in the Mediterranean region. This region has been extensively and intensely used for millennia; it is the cradle of Western civilization, and since then many cultures have inhabited the region, some with very high population densities. The consequence is that Mediterranean landscapes have been extensively modified by deforestation and overgrazing, and much of the land is farmed. However, recent (late 20th century) socio-economic changes in the region (due to industrialization and tourism) have led many people to abandon rural areas. Thus, these areas are emptying, and urban and peri-urban areas are increasingly crowded. These processes imply in many Mediterranean landscapes an abrupt reduction of land for agriculture, grazing, and wood gathering, and poor management of forest plantations. Some imprints of the past land overuse are easily observable (e.g. Fig. 3).

The first and direct consequence of this process is a sudden increase of wildlands and plant biomass. Oldfields are now a primary feature in many Mediterranean landscapes, which increases the amount and connectivity of vegetation in many regions. Many European Mediterranean landscapes are defaunated of large herbivores, and resemble grazing exclusion experiments because the loss of livestock has not been accompanied by an increase in wild herbivores that once occupied the land. In addition, forest plantations are poorly maintained – further increasing the biomass and connectivity of the landscapes. Comparison of current landscapes with old photographs (Debussche et al. 1999), old maps (Falcucci et al. 2007), or comparisons among consecutive forest inventories (Vayreda et al. 2012, De Cáceres et al. 2015) show a clear increase in wildlands and forest tree density. Even remote-sensing data gathered over the last 15 years suggest a generalized greening of the region compared with global trends (Fig. 4a). The atmospheric increase in CO₂ may have contributed to biomass growth but is unlikely to be the main driver as the magnitude of the greening is greater than global estimates (Fig. 4a). The Mediterranean region is a greening hotspot, despite also being a climate change hotspot (Giorgi 2006).

A consequence of this vegetation increase in a warming climate is increased drought stress and mortality in plants. That is, in addition to direct climate-induced drought mortality that may be occurring independently of any recent vegetation changes (Barbeta et al. 2015, Lloret et al. 2016), there is drought-stress and mortality attributed to the increased vegetation density, plant competition, and natural self-thinning processes (Vayreda et al. 2012, De Cáceres et al. 2015). These processes are natural in plant communities, but they now occur in a low grazing and high CO₂ environment that increases the magnitude of the plant stress and mortality.

Browning: semi-urban population provides ignitions

Land abandonment and the associated increase in vegetation imply an increase in the amount and connectivity of fuels, and thus in fire hazard. This is occurring in an area where human population is increasing – and especially the urban population (i.e., people not living from the land) in the wildland-urban interface (Modugno et al. 2016). The consequence of such human changes is an increase in fire ignitions (by negligence and arson), and, given the large amount of fuel available in the landscape, an increase in the fire size and annual area burned (Pausas 2004, Moreira et al. 2011, Pausas & Fernández-Muñoz 2012). This leads to an increase in fire recurrence at the local scale. Simultaneously with this increase in fuel and ignitions, there has been a strong policy of fire suppression in most countries, which further promotes fuel accumulation and large fires (Piñol et al. 2007). The increased frequency of conditions conducive to fire (hotter and longer fire seasons due to global warming; Moriondo et al. 2006), have further contributed to this recent fire regime shift (Fig. 2). However, the major change in fuel amount and distribution is probably the most important driver of abrupt fire regime changes (Pausas & Fernández-Muñoz 2012; Pausas & Keeley 2014). In Mediterranean Europe, fires used to be fuel-limited when the rural population was large, and fires are now driven by drought events as fuel is no longer a limiting factor (Pausas & Fernández-Muñoz 2012; Gouveia et al. 2016). In northern Africa, where the rural population is still significant, fires are smaller and fuel-limited (Chergui et al. 2018). Failing to consider land abandonment precludes predicting fire regime shifts and vegetation changes in the Mediterranean region.

Vegetation changes can be depicted by discontinuities in the normalized difference vegetation index (NDVI) from remote sensing (Verbesselt et al. 2010). Abrupt discontinuities in the NDVI are common in the Mediterranean region (Fig. 4b) and can be produced by major droughts, logging activities, or fires. Fire regime shifts that can generate abrupt vegetation changes are especially important. Mediterranean ecosystems have an evolutionary history linked to fire; however, disruptions in sustainable fire regimes are common (Keeley et al. 2012). For instance, non-serotinous coniferous forests were not historically subject to fire – or were subject to low intensity understory fires that left the forest structure unaffected (Fulé et al. 2008, Slimani et al. 2014). Yet some of these forests are now suffering high-intensity crown-fires due to increased crowding (related to reduced grazing and lack of low intensity fires) and warming. A single

crown-fire in a non-serotinous pine forest abruptly switches the community to a new vegetation state, typically shrubland or oak forest (Retana et al. 2002). In contrast, serotinous pines are well equipped to survive crown-fires (Pausas 2015). However, two consecutive fires with an interval shorter than the maturity age of the dominant pine, switches the pine forest to shrubland. These vegetation-type switches do not revert easily, especially in human-fragmented landscapes, and are common across the Mediterranean basin (Moreira et al. 2011).

Predictions for the Mediterranean region suggest an increased fire risk in the decades ahead (Bedia et al. 2014, Sounsa et al. 2015, Turco et al. 2018). This prediction is often based on climate, and thus, given the observed and predicted increased temperature and reduced summer rainfall, the conditions for fire spread are increasing. However, for predicting fire regimes (and not just the meteorological conditions conducive to fire – the fire weather), it is necessary to account for other interacting factors, especially those related to fuel amount and contiguity (fire hazard). Recent simulation studies suggest that including CO₂ and human population dynamics (e.g., urbanization) in climate-based global models, switches the predictions from an increase to a decrease in fire activity at a global scale (Knorr et al. 2016). In the Mediterranean region, such a decrease of fire activity in the (near) future is unlikely in much of the region, because fuel increase from land abandonment (the greening above) together with increased climatic conditions conducive to fire (drought and longer fire seasons) and the continuous increase in human ignitions, provide conditions for larger and more frequent fires (Fig. 2). Only in very dry regions can more severe drought reduce fuels and so limit fires (Pausas & Paula 2012).

Browning: coastal degradation enhances drought

The Mediterranean coastline was once fully vegetated and rich in marshes and lagoons that contributed to the local water cycle. The sea breeze enriched with the water that evaporated and transpired from these ecosystems increased in altitude when moving inland over the coastal mountains, and typically condensed before surpassing the mountains. The rich volatile organic compounds emitted by Mediterranean plants contributed to the nucleation of clouds (Kyrkby et al. 2016). This water vapor condensation was responsible for convective showers over the mountains around the Mediterranean Sea (Bolle 2003). The desiccation of the marshes (started by the Romans to prevent malaria), deforestation for agriculture, and more recently, explosive coastal urbanization and tourism have drastically reduced the original ecosystems and thus the water available for the sea breeze. These land use changes have also increased soil heating (the urban heat island effect). The consequences are the destruction of several ecosystems, changes in species composition, a lowering of the water table, changes in the albedo, but also, with a warmer and drier air breeze, the condensation point has increased in altitude and thus storms develop less often (Millán 2014; Supporting Information S2). There is evidence in recent decades of reductions in summer (convective-orographic) rainfall and increases in autumnal (often torrential) rain (Pausas 2004, Millán et al. 2005a, Pastor et al. 2015). These summer ‘storm

losses' affect downstream vegetation and marshes, and thus generate a negative feedback process that can lead to aridification and water shortage for humans. This is not driven by global warming, but by local changes in land use; and global warming exacerbates this process by further heating and reducing evapotranspiration (Fig. 2). This disruption of the water cycle also increases dry storms, which are a cause of wildfires. The water vapor that does not precipitate, together with photo-oxidants (ozone) and aerosols, increase sea temperature; and this facilitates the water vapor to feed torrential autumnal rains in the region (Bozkurt and Sen 2009, Pastor et al. 2015) and leads to flooding in other parts of the catchment and neighboring areas (Fig. 1; e.g., central-eastern Europe: Gangoi et al. 2011; North Africa: Park et al. 2016). Note that this accumulation of water vapor and ozone is very relevant as both are greenhouse gases (GHG) with high global warming potential (Held & Soden 2000); and there is evidence of heat intensification by GHG along the coast (Diffenbaugh et al. 2007). In other words, the Mediterranean basin can behave like a large cauldron recirculating air masses in different directions; it is also an example of how human activities may affect the dynamics of water vapor (an important greenhouse gas).

Land degradation on the coast is not compensated by the greening from the land abandonment outlined above for two reasons. Firstly, most degradation occurs very close to the coastline and replaces natural systems with urban structures, while land abandonment and recolonization by wild vegetation is a general process occurring in the slopes, mainly in low-value land (i.e, not near the coast). The second reason is that degradation of the coastal area implies a reduction in moisture for the water cycle, as well as an increase in air temperature (urban heat effect and soil sealing). As temperatures rise, the capacity for the air to retain vapor (saturated vapor density) increases exponentially, and thus the condensation level moves up in altitude (Millán et al. 2005b). Therefore, a small change in air temperature at the coast implies a large increase in the altitude at which the water vapor driven by sea breeze precipitates, and thus fewer storms occur (Millán 2014). All the feedback processes related to coastal degradation are driven by fine-scale patterns (orography, size of minor catchments) and these patterns are not well captured by most simulation models (Millán 2014, Tierney et al. 2017).

The Mediterranean conundrum: policy implications

Greening driven by rural abandonment is generally masking the browning effects of increased droughts, unnatural fire regimes, and the disruption of the hydrological cycle that is occurring simultaneously. Both greening and browning processes outlined above are especially occurring in Mediterranean Europe; and will be exacerbated as countries from the southern rim (North Africa) abandon their rural life style as a consequence of increasing industrialization and tourism (Chergui et al. 2018).

Land abandonment is, to some extent, buffering the consequences of global warming. Thus, it may also reduce the social perception of the effects of warming in the region. Drought-induced defoliation and mortality events occur (Barbeta et al. 2015, Lloret et al. 2016), phenological changes are noticeable (in crops and wildlife; e.g., Gordo and Sanz, 2005), and soil erosion is observed. Yet for the moment, greening is still the dominant trend (Fig. 4), and this contributes to reduced soil losses (Keesstra 2007). As droughts increase in frequency the balance could switch easily with the browning offsetting the greening. How far are we from the turning point when water availability will limit further greening? Can land management delay reaching this threshold?

The urgent need for global policies to reduce emissions is well-known; however, given the importance of fine-scale processes (Fig. 2), some local actions are also important, and probably underestimated by local agencies. Here we outline some local policies and research needed to disrupt the two feedback processes sketched in Fig. 2.

Fire. Fire requires three components (fuel, ignitions, and dry conditions) and all three are increasing in Mediterranean ecosystems (Fig. 2), thus fires are likely to increase in the coming decades. Fire policies in most Mediterranean countries are based on the zero-tolerance principle, that is, preventing and extinguishing all fires; this is often very effective (at short time scales) but leads to larger high intensity fires (megafires) in the long run (Piñol et al. 2007; Curt & Frejaville 2018). A policy towards small and/or low-intensity fires would be desirable for preventing large catastrophic fires and thus generating a more sustainable fire regime. This can be achieved by the introduction of prescribed burns, currently forbidden in many countries, and by letting some wildfires burn (those that burn in relatively mild conditions; Boisramé et al. 2017). Low-intensity fires can ameliorate water stress in Mediterranean trees (Alfaro-Sánchez et al. 2016, Young et al. 2017) and prevent large high-intensity fires. In addition, the introduction of grazing animals (e.g., enhancing pastoralism, prescribed grazing, or rewilding natural herbivores) would be helpful as long as they are introduced with care (e.g., avoiding overgrazing, especially in postfire environments). Other management options such as forest thinning can also be helpful at strategic points to reduce the chance of fire ignitions and spread. Appropriate urban planning is also key because semi-urban areas provide the main fire ignition points. In many Mediterranean ecosystems, fire is a disaster only where human infrastructures are found because many plants are well adapted to fires, including to high-intensity fires (Keeley et al. 2012). That is, the wildland-urban interface (WUI) can convert ecologically sustainable fire regimes into social catastrophes. Thus, reducing the WUI or at least limiting its expansion (e.g., using taxation as a deterrent) is highly desirable.

An important topic for research in relation to fire management is the carbon balance with recurrent Mediterranean fires, and the role of fire management in modulating this balance. In the short term, fire releases large amounts of CO₂. However, in the longer term, regeneration (and especially resprouting) quickly fix large amounts of C and thus fires may switch senescent communities to C-sinks. In addition, charcoal is a recalcitrant C contribution to the soil and thus fires increase the C-sink (Santín et al. 2015). To what extent prescribed fires can be used to optimize the source-sink balance remains to be studied in detail. The dominance of species with high postfire resprouting capacity (and with large underground organs, Pausas et al. 2018) makes studying this balance in Mediterranean ecosystems especially appropriate.

Drought. An important direct policy needed for the Mediterranean regions is to reactivate the water cycle that has been disrupted by coastal degradation. Research in eastern Spain suggests that to generate water condensation on the coastal mountains (e.g., 2000 m above sea level), the average amount of water needed in the air is greater (at least 21 g/Kg) than the average observed (Millán 2014, Supplementary Material S2). This emphasizes the need to pump water into the cycling system in order to change a negative into a positive feedback and thus prevent a reduction in the number of storms. This could be undertaken by massive revegetation of the lowlands around the coast, including degraded wildlands (revegetation with native plants), and also by greening semi-urban and urban areas. This would include encouraging the planting of trees on urban streets, the creation of large urban and semi-urban gardens, and conserving and enlarging all patches of native vegetation along the coast. The evapotranspiration of all this vegetation (native, urban, and semi-urban) would benefit the water cycle upstream (Ellison et al. 2012). In fact, these actions have social benefits far beyond activating the water cycle (Willis and Petrokofsky 2017, Endreny et al. 2017). Maximum priority should be given to preserving, restoring, and enlarging coastal marshes and wetlands as they provide very important water inputs into the system, in addition to being hotspots of biodiversity and carbon sinks (Howard et al. 2017). Given that a small change in air temperature on the coast has implications for the upslope rainfall in the mountains, it is important to reduce urban heat as much as possible. Thus, in urban areas, building using cool materials (i.e., pavement and rooftops with high reflectance and high emissivity factor) could be considered (Santamouris et al. 2012). Such building techniques would also reduce cooling needs inside buildings and have a positive environmental impact (Santamouris et al. 2008). Installing fog collectors in the coastal mountains would facilitate capturing water from sea breeze (non-rainfall water; Estrela et al. 2008, Kim et al. 2017, Kaseke et al. 2017) that could be used for agricultural irrigation while saving water and reactivating the water cycle.

Mechanisms acting at a fine scale, together with global drivers (CO₂ enrichment and climatic warming) interact and drive current vegetation changes in Mediterranean landscapes. Any model aiming to predict the future of our vegetation and climate must consider these local mechanisms; and failing to consider them at an appropriate scale is likely to produce inconclusive predictions (Diffenbaugh et al. 2007, Millán 2014, Tierney et al. 2017). Current global and most regional

models for the Mediterranean area (e.g., Milano et al. 2012, Guiot and Cramer 2016) still suffer some of these deficiencies. The advantage of the importance of small-scale drivers is that local policies and management actions can make a difference in reducing global change impacts.

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References

- Alfaro-Sánchez R, Camarero JJ, Sánchez-Salguero R, Sangüesa-Barreda G, De Las Heras J. 2016. Post-fire Aleppo pine growth, C and N isotope composition depend on site dryness. *Trees* 30:581-595.
- Barbeta A, Mejía-Chang M, Ogaya R, Voltas J, Dawson TE, Peñuelas J. 2015. The combined effects of a long-term experimental drought and an extreme drought on the use of plant-water sources in a Mediterranean forest. *Global Change Biology* 21:1213–1225.
- Bedia J, Herrera S, Camia A, Moreno JM, Gutiérrez JM. 2014. Forest fire danger projections in the Mediterranean using ENSEMBLES regional climate change scenarios. *Climatic Change* 122:185-199.
- Boisramé G, Thompson S, Collins B, Stephens S. 2017. Managed wildfire effects on forest resilience and water in the Sierra Nevada. *Ecosystems* 20:717-732.
- Bolle H-J. 2003. Climate, climate variability, and impacts in the Mediterranean area: an overview. Pages 5-86 in Bolle H-J, ed. *Mediterranean Climate: Variability and Trends*. Berlin, Heidelberg: Springer Berlin Heidelberg.
- Bozkurt D, Sen OL. 2011. Precipitation in the Anatolian Peninsula: sensitivity to increased SSTs in the surrounding seas. *Climate Dynamics* 36:711-726.
- Chergui B, Fahd S, Santos X, Pausas JG. 2018. Socioeconomics drive fire regime variability in the Mediterranean Basin. *Ecosystems* 21: 619–628.
- Cortambert E, ed. 1870. *Nouvel atlas de géographie*. Librairie Hachette. Paris.
- Ciarlo JM, Aquilina NJ. 2016. An analysis of teleconnections in the Mediterranean region using RegCM4. *International Journal of Climatology* 36:797-808.
- Curt T, Frejaville T. 2018. Wildfire policy in Mediterranean France: How far is it efficient and sustainable? *Risk Analysis* 38: 472-488.
- De Cáceres M, Martínez-Vilalta J, Coll L, Llorens P, Casals P, Poyatos R, Pausas JG, Brotons L. 2015. Coupling a water balance model with forest inventory data to predict drought stress: the role of forest structural changes vs. climate changes. *Agricultural and Forest Meteorology* 213:77-90.
- Debussche M, Lepart J, Dervieux A. 1999. Mediterranean landscape changes: evidence from old postcards. *Global Ecology and Biogeography Letters* 8:3-15.
- Diffenbaugh NS, Pal JS, Giorgi F, Gao X. 2007. Heat stress intensification in the Mediterranean climate change hotspot. *Geophysical Research Letters* 34:L11706.

- Ellison D, N. Futter M, Bishop K. 2012. On the forest cover–water yield debate: from demand- to supply-side thinking. *Global Change Biology* 18:806-820.
- Estrela MJ, Valiente JA, Corell D, Millán MM. 2008. Fog collection in the western Mediterranean basin (Valencia region, Spain). *Atmospheric Research* 87:324-337.
- Falcucci A, Maiorano L, Boitani L. 2007. Changes in land-use/land-cover patterns in Italy and their implications for biodiversity conservation. *Landscape Ecology* 22:617-631.
- FAO. 2013. State of Mediterranean forests 2013. FAO.
- Fulé PZ, Ribas M, Gutiérrez E, Vallejo R, Kaye MW. 2008. Forest structure and fire history in an old *Pinus nigra* forest, eastern Spain. *Forest Ecology and Management* 255:1234-1242.
- Gangoiti G, Alonso L, Navazo M, García JA, Millán MM. 2006. North African soil dust and European pollution transport to America during the warm season: Hidden links shown by a passive tracer simulation. *Journal of Geophysical Research: Atmospheres* 111:D10109.
- Gangoiti G, Gómez-Domenech I, Sáez de Cámara E, Alonso L, Navazo M, Iza J, García JA, Ilardia JL, Millán MM. 2011. Origin of the water vapor responsible for the European extreme rainfalls of August 2002: 2. A new methodology to evaluate evaporative moisture sources, applied to the August 11–13 central European rainfall episode. *Journal of Geophysical Research: Atmospheres* 116:16.11.2011.
- Giorgi F. 2006. Climate change hot-spots. *Geophysical Research Letters* 33:L08707.
- Giorgi F, Lionello P. 2008. Climate change projections for the Mediterranean region. *Global and Planetary Change* 63:90-104.
- Gordo O, Sanz JJ. 2005. Phenology and climate change: a long-term study in a Mediterranean locality. *Oecologia* 146:484-495.
- Gouveia CM, Bistinas I, Liberato MLR, Bastos A, Koutsias N, Trigo R. 2016. The outstanding synergy between drought, heatwaves and fuel on the 2007 Southern Greece exceptional fire season. *Agricultural and Forest Meteorology* 218–219:135-145.
- Guiot J, Cramer W. 2016. Climate change: The 2015 Paris Agreement thresholds and Mediterranean basin ecosystems. *Science* 354:465-468.
- Held IM, Soden BJ. 2000. Water vapor feedback and global warming. *Annual Review of Energy and the Environment* 25:441-475.
- Howard J, Sutton-Grier A, Herr D, Kleypas J, Landis E, McLeod E, Pidgeon E, Simpson S. 2017. Clarifying the role of coastal and marine systems in climate mitigation. *Frontiers in Ecology and the Environment* 15:42-50.
- IPCC. 2014. Climate Change 2014: Synthesis Report. Contribution of Working Groups I, II and III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change [Core Writing Team, R.K. Pachauri and L.A. Meyer (eds.)]. Geneva, Switzerland: IPCC. Report no.
- Jenness J, Dooley J, Aguilar-Manjarrez J, Riva C. 2007. African Water Resource Database. GIS-based tools for inland aquatic resource management. 2. Technical manual and workbook. CIFA Technical Paper. No. 33, Part 2. Rome, FAO. Report no.
- Kaseke KF, Wang L, Seely MK. 2017. Nonrainfall water origins and formation mechanisms. *Science Advances* 3:e1603131.
- Keeley JE, Bond WJ, Bradstock RA, Pausas JG, Rundel PW. 2012. Fire in Mediterranean ecosystems: ecology, evolution and management. Cambridge University Press.
- Keesstra SD. 2007. Impact of natural reforestation on floodplain sedimentation in the Dragonja basin, SW Slovenia. *Earth Surface Processes and Landforms* 32:49-65.

- Kemp M. 2005. Inventing an icon. *Nature* 437:1238-1238.
- Kim H, Yang S, Rao SR, Narayanan S, Kapustin EA, Furukawa H, Umans AS, Yaghi OM, Wang EN. 2017. Water harvesting from air with metal-organic frameworks powered by natural sunlight. *Science* 356:430-434.
- Kirkby J, et al. 2016. Ion-induced nucleation of pure biogenic particles. *Nature* 533:521-526.
- Knorr W, Arneth A, Jiang L. 2016. Demographic controls of future global fire risk. *Nature Clim. Change* 6:781-785.
- Lelieveld J, et al. 2012. Climate change and impacts in the Eastern Mediterranean and the Middle East. *Climatic Change* 114:667-687.
- Lloret F, Riva EG, Pérez-Ramos IM, Marañón T, Saura-Mas S, Díaz-Delgado R, Villar R. 2016. Climatic events inducing die-off in Mediterranean shrublands: are species' responses related to their functional traits? *Oecologia* 180, 961–973.
- Mariotti A, Zeng N, Yoon J-H, Artale V, Navarra V, Alpert P, Li LZ. 2008. Mediterranean water cycle changes: transition to drier 21st century conditions in observations and CMIP3 simulations. *Environmental Research Letters* 3:044001.
- Milano M, Ruelland D, Fernandez S, Dezetter A, Fabre J, Servat E. 2012. Facing climatic and anthropogenic changes in the Mediterranean basin: What will be the medium-term impact on water stress? *Comptes Rendus Geoscience* 344:432-440.
- Millán MM. 2014. Extreme hydrometeorological events and climate change predictions in Europe. *Journal of Hydrology* 518, Part B:206-224.
- Millán MM, Estrela MJ, Miró J. 2005. Rainfall Components: Variability and Spatial Distribution in a Mediterranean Area (Valencia Region). *Journal of Climate* 18:2682-2705.
- Millán MM, Estrela MJ, Sanz MJ, Mantilla E, Martín M, Pastor F, Salvador R, Vallejo R, Alonso L, Gangoiti G. 2005. Climatic feedbacks and desertification: the Mediterranean model. *Journal of Climate* 18:684-701.
- Modugno S, Balzter H, Cole B, Borrelli P. 2016. Mapping regional patterns of large forest fires in Wildland–Urban Interface areas in Europe. *Journal of Environmental Management* 172:112-126.
- Moreira F, et al. 2011. Landscape – wildfire interactions in southern Europe: Implications for landscape management. *Journal of Environmental Management* 92:2389-2402.
- Moriondo M, Good P, Durao R, Bindi M, Giannakopoulos C, Corte-Real J. 2006. Potential impact of climate change on fire risk in the Mediterranean area. *Climate Research* 31:85-95.
- Myers N, Mittermeier RA, Mittermeier CG, da Fonseca GAB, Kent J. 2000. Biodiversity hotspots for conservation priorities. *Nature* 403:853-858.
- Park J-Y, Bader J, Matei D. 2016. Anthropogenic Mediterranean warming essential driver for present and future Sahel rainfall. *Nature Clim. Change* 6:941-945.
- Pastor F, Valiente JA, Estrela MJ. 2015. Sea surface temperature and torrential rains in the Valencia region: modelling the role of recharge areas. *Nat. Hazards Earth Syst. Sci.* 15:1677-1693.
- Pausas JG. 2004. Changes in fire and climate in the eastern Iberian Peninsula (Mediterranean basin). *Climatic Change* 63:337-350.
- Pausas JG, Fernández-Muñoz S. 2012. Fire regime changes in the Western Mediterranean Basin: from fuel-limited to drought-driven fire regime. *Climatic Change* 110:215-226.

- Pausas JG, Keeley JE. 2014. Abrupt climate-independent fire regime changes. *Ecosystems* 17:1109-1120.
- Pausas JG, Ribeiro E. 2017. Fire and plant diversity at the global scale. *Global Ecology and Biogeography* 26:889–897.
- Pausas JG, Lamont BB, Paula S, Appezzato-da-Glória B, Fidelis A. 2018. Unearthing belowground bud banks in fire-prone ecosystems. *New Phytologist* 217:1435–1448.
- Piñol J, Castellnou M, Beven KJ. 2007. Conditioning uncertainty in ecological models: Assessing the impact of fire management strategies. *Ecological Modelling* 207:34-44.
- Raicich F, Pinardi N, Navarra A. 2003. Teleconnections between Indian monsoon and Sahel rainfall and the Mediterranean. *International Journal of Climatology* 23:173-186.
- Retana J, Espelta JM, Habrouk A, Ordoñez JL, de Solà-Morales F. 2002. Regeneration patterns of three Mediterranean pines and forest changes after a large wildfire in northeastern Spain. *Ecoscience* 9:89-97.
- Santamouris M, Gaitani N, Spanou A, Saliari M, Giannopoulou K, Vasilakopoulou K, Kardomateas T. 2012. Using cool paving materials to improve microclimate of urban areas – Design realization and results of the flisvos project. *Building and Environment* 53:128-136.
- Santín C, Doerr SH, Preston CM, González-Rodríguez G. 2015. Pyrogenic organic matter production from wildfires: a missing sink in the global carbon cycle. *Global Change Biology* 21:1621-1633.
- Skliris N, Zika JD, Herold L, Josey SA, Marsh R. 2018. Mediterranean sea water budget long-term trend inferred from salinity observations. *Climate Dynamics* doi:10.1007/s00382-00017-04053-00387.
- Slimani S, Touchan R, Derridj A, Kherchouche D, Gutiérrez E. 2014. Fire history of Atlas cedar (*Cedrus atlantica* Manetti) in Mount Chélia, northern Algeria. *Journal of Arid Environments* 104:116-123.
- Sousa PM, Trigo RM, Pereira MG, Bedia J, Gutiérrez JM. 2015. Different approaches to model future burnt area in the Iberian Peninsula. *Agricultural and Forest Meteorology* 202:11-25.
- Taylor CM, Belušić D, Guichard F, Parker DJ, Vischel T, Bock O, Harris PP, Janicot S, Klein C, Panthou G. 2017. Frequency of extreme Sahelian storms tripled since 1982 in satellite observations. *Nature* 544:475.
- Thompson JD. 2005. *Plant evolution in the Mediterranean*. Oxford University Press.
- Tierney JE, Pausata FSR, deMenocal PB. 2017. Rainfall regimes of the Green Sahara. *Science Advances* 3:e1601503.
- Turco M, Rosa-Cánovas JJ, Bedia J, Jerez S, Montávez JP, Llasat MC, Provenzale A. 2018. Exacerbated fires in Mediterranean Europe due to anthropogenic warming projected with non-stationary climate-fire models. *Nature Communications* 9:3821.
- Vayreda J, Martínez-Vilalta J, Gracia M, Retana J. 2012. Recent climate changes interact with stand structure and management to determine changes in tree carbon stocks in Spanish forests. *Global Change Biology* 18:1028-1041.
- Verbesselt J, Hyndman R, Newnham G, Culvenor D. 2010. Detecting trend and seasonal changes in satellite image time series. *Remote Sensing of Environment* 114:106-115.
- Willis KJ, Petrokofsky G. 2017. The natural capital of city trees. *Science* 356:374-376.

Young DJN, Stevens JT, Earles JM, Moore J, Ellis A, Jirka AL, Latimer AM. 2017. Long-term climate and competition explain forest mortality patterns under extreme drought. *Ecology Letters* 20:78-86.

Table 1. Mediterranean basin facts. Extent, diversity and characteristics of the Mediterranean climate region (Fig. 1). Values are approximate as different sources use slightly different limits of the Mediterranean climate area.

Parameter	Value
Extension of the land (Km ²)	2 million
Sea area (Km ²)	2.5 million
Drainage region (Km ² ; Fig. 1)	11 million
Population (total)	~200 million
Population density (inhab./Km ²)	~ 100
East-west distance (Km)	3,800
North-south distance (Km)	>1,000
Coast length (Km)	> 40,000
Number of languages	> 20
Number of countries	18
Number of islands	> 4,000
Number of islands > 5 Km ²	189
Highest altitude (m)	4167
Number of WWF ecoregions	21
Number of plant species	25,000
Number of mammal species	305
Number of reptile species	351
Number of bird species	601

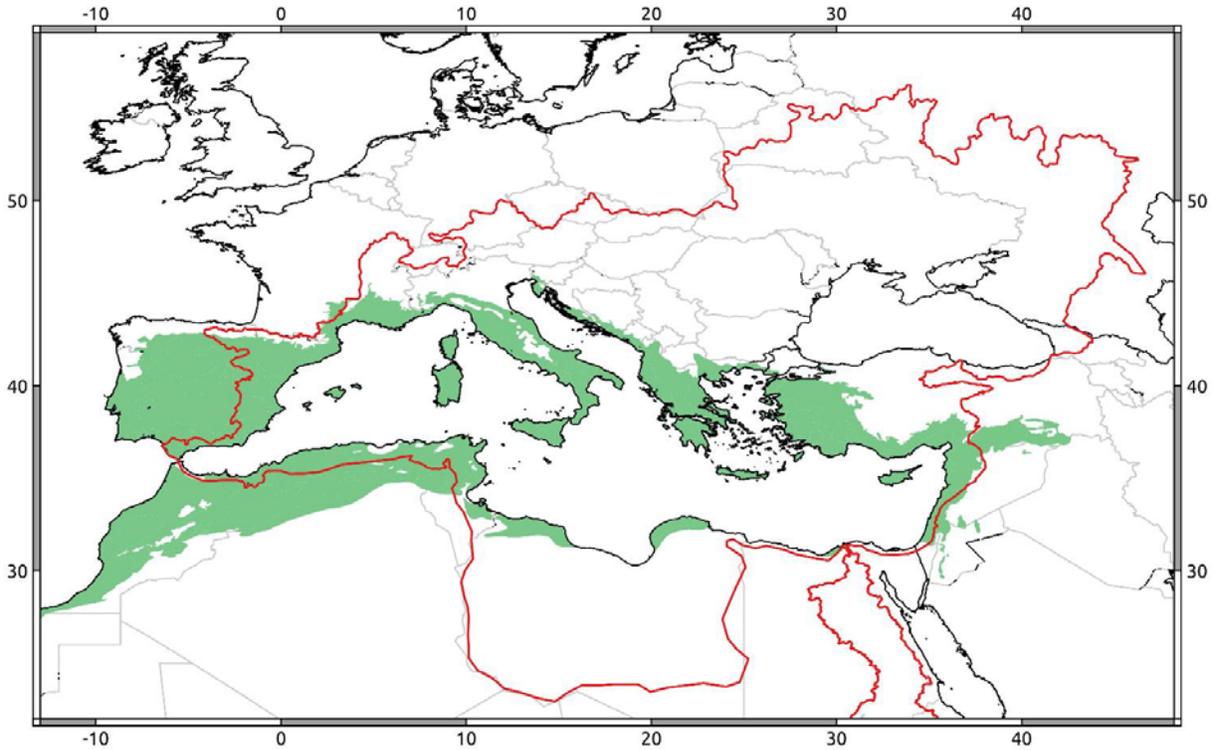
Figures

Fig. 1. The region with a mediterranean climate (filled region) surrounds the Mediterranean Sea, but the whole Mediterranean catchment is far larger (thick red line). Thus atmospheric processes occurring in the Mediterranean Sea can have hydrological implications in the central and eastern part of the European continent. Sources: Cortambert (1870), WWF, Millán (2014), Jenness et al. (2007).

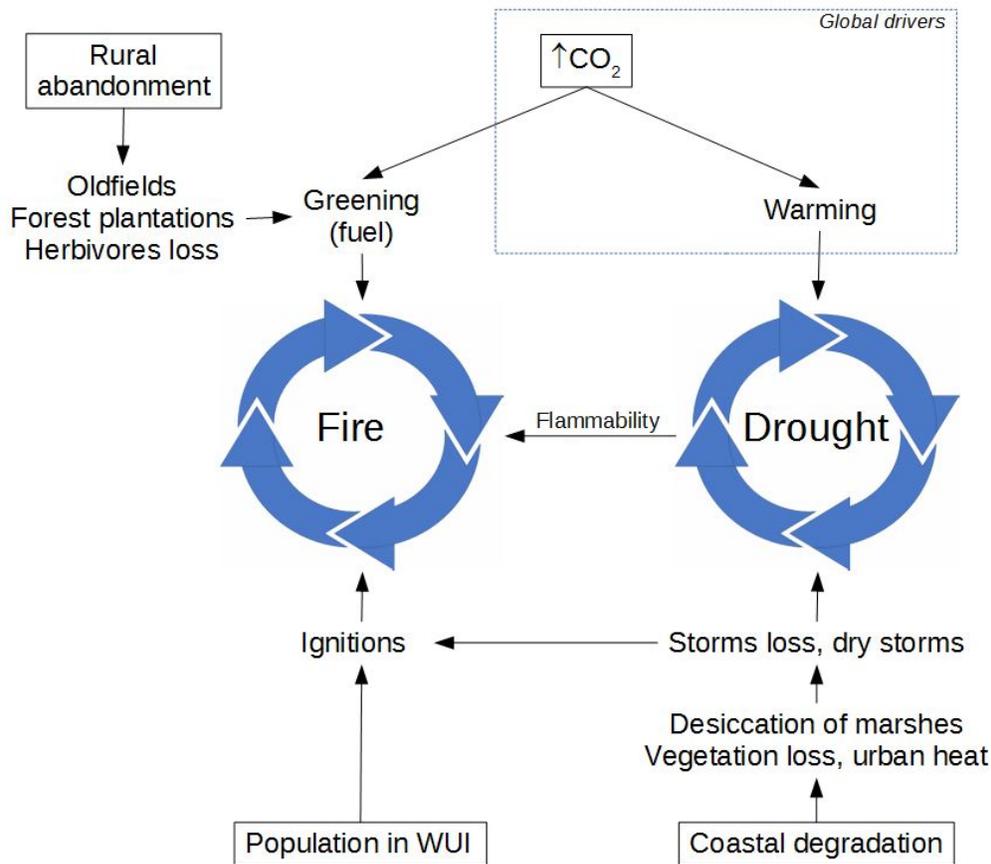


Fig. 2. The disruption of the natural fire and drought regimes in the Mediterranean landscapes is driven by global and local drivers. Increased fire activity is a response to the fuel amount and landscape homogeneity generated by rural abandonment (fire hazard) in an environment depauperate of herbivores and with increasing human ignitions (fire risk) and droughts (fire weather). The increased dry conditions are the consequence of global warming – but also of storm losses caused by the disruption of the water cycle generated by the coastal degradation (see main text and Supplementary Material S2). WUI: wildland-urban interface.



Fig. 3. Terraced slopes are a common feature of the Mediterranean landscapes and illustrate the strong human pressure in the region. Note that most of the terraces are abandoned and covered by early successional (and flammable) plants, except those at the bottom of the mountains that are still in use. The shrubland and dead trees in the foreground are the product of a wildfire. Location: Benicadell (Valencia, Spain); photo: JG Pausas.

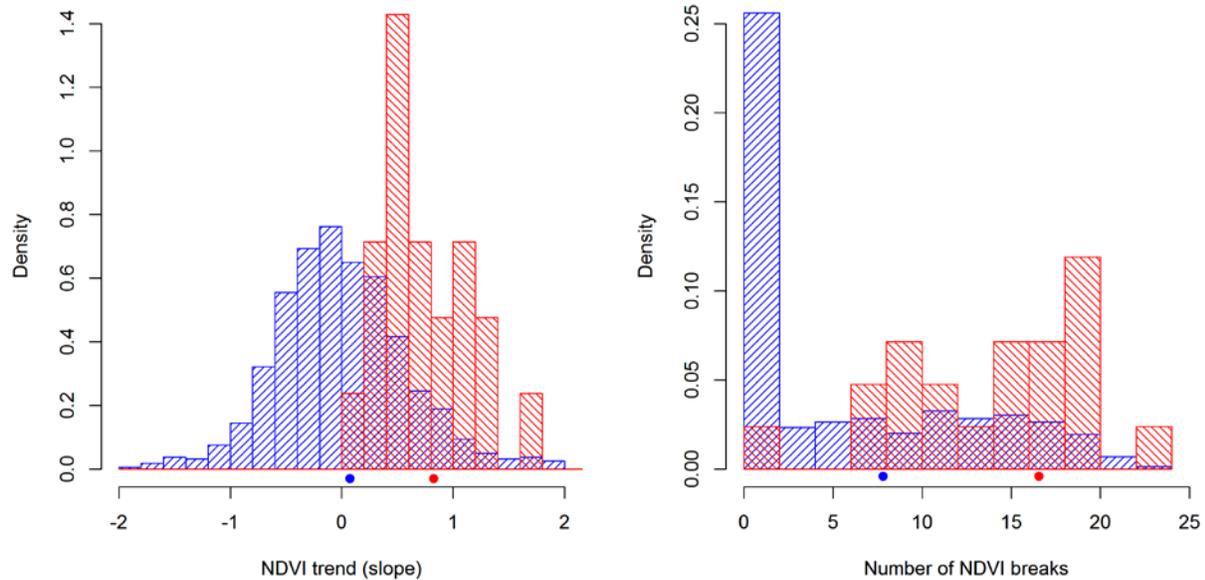


Fig. 4. Histogram (probability density) of the slopes (A, left) and the number of breaks (B, right) in the NDVI trend for the period 2000-2015 for all world ecoregions (blue right stripes, $n=792$) and for the ecoregions in the Mediterranean basin (red left stripes, $n=21$). Dots on the x -axes indicate the mean value (weighted by the size of the ecoregion). The figures show that the NDVI dynamics in the Mediterranean basin tend to be stronger (higher slopes and more breaks) than most of the other biomes. NDVI data by ecoregion from Pausas & Ribeiro (2017; see Supporting Information S1 for details).

Supporting Information

Pausas & Millán - Greening and browning in a climate change hotspot: the Mediterranean Basin. *BioScience*.

The following Supporting Information is available for this article:

S1. Methods for Figure 3.

S2. Summer storms loss on Mediterranean coast

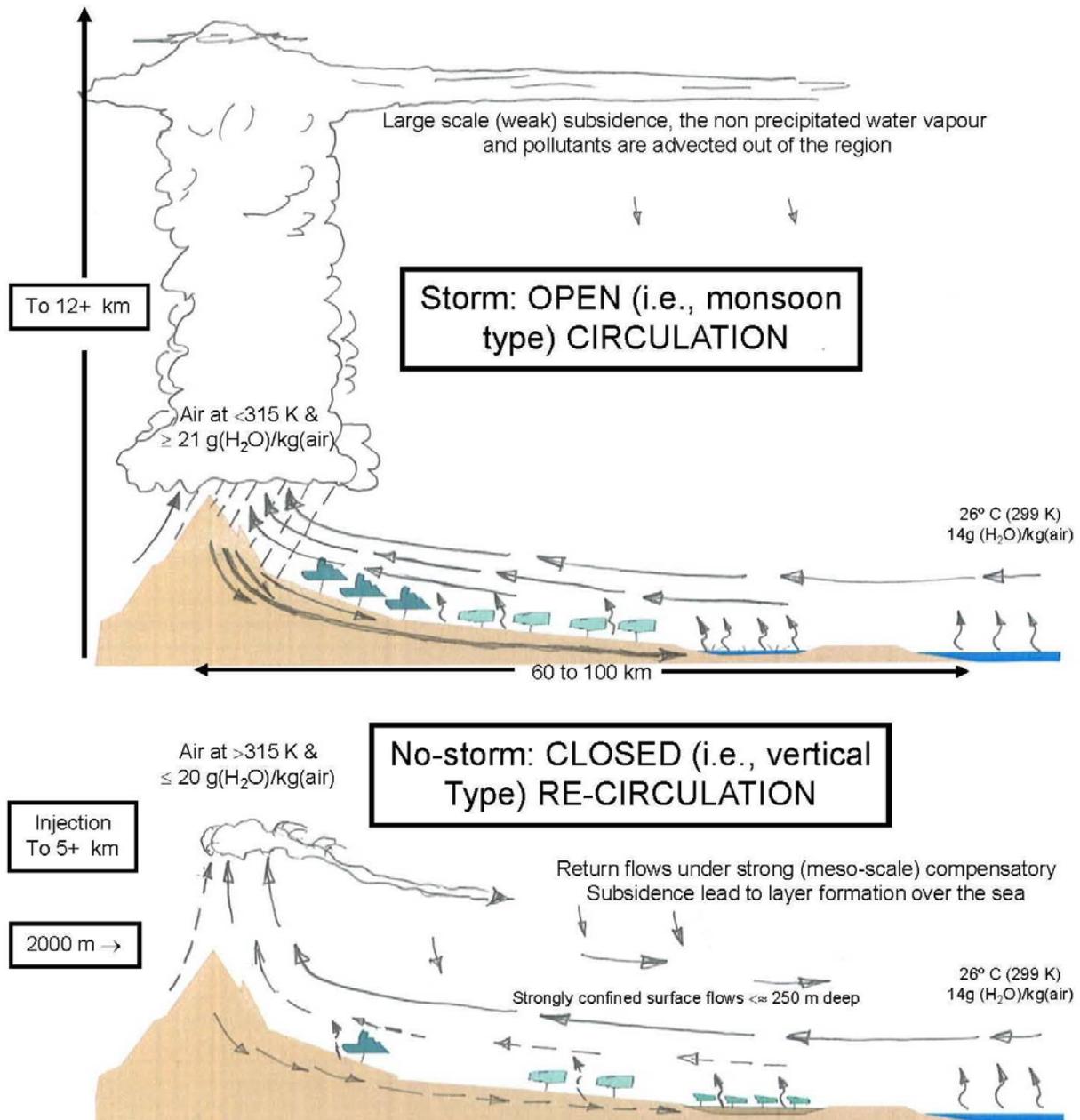
S1. Methods for Figure 3.

The NDVI data used for Figure 1 comes from Pausas & Ribeiro (2017). It is based on 16-day global NDVI images (resolution= 0.05 degrees) from 2/2000 to 7/2015 (355 global images, MODIS MOD13C1 data set obtained from the Land Processes Distributed Active Archive Center, USGS). Each image was crossed with the ecoregion map, and for each ecoregion, we extracted the mean of the NDVI. The 355 NDVI mean values for each ecoregion were treated as a time-series and decomposed into seasonality and trend using the BFAST approach (Verbesselt et al. 2010). The coefficient of the trend is used in Figure 3a, and the number of abrupt changes in the trend is used in Figure 3b.

Pausas, JG, Ribeiro, E (2017) Fire and plant diversity at the global scale. *Global Ecology and Biogeography*, 26: 889–897.

Verbesselt, J, Hyndman, R, Zeileis, A, Culvenor, D (2010) Phenological change detection while accounting for abrupt and gradual trends in satellite image time series. *Remote Sensing of Environment* 114, 2970-2980.

S2. Summer storms loss on Mediterranean coast



Schematic representation of the water cycle on the Mediterranean coasts (from Millán et al. 2005, Millán 2014). **The upper panel** represents the original (natural) conditions with little anthropogenic disturbances (open circulation), where sea breezes and water evaporated from natural ecosystems (marshes, lagoons, shrublands and forests) feed the storms in the mountains. The climatic values of the water vapour mixing ratio and the temperature at the coast of the incoming airmass in Jul-Aug are 14 g/kg and 26 C (299 K), respectively. However, by the time the same airmass reaches the interior, its temperature gain is some 16°C. Thus the airmass in

the combined breeze requires to gain an average of 7 g/kg along its land path to reach a water vapour mixing ratio of ≥ 21 g/kg to keep its Convective (orographic) Condensation Level (CCL) below the approximate height of the coastal mountain ranges (ca. 2000 m altitude). This additional water vapour comes from the evapotranspiration in natural ecosystems. **The lower panel** shows the conditions where reduced vegetation and desiccated coastal marshes do not generate enough water vapor to feed a storm (closed recirculation). Current observations suggest that the atmospheric circulations over the coasts of the Western Mediterranean Basin in summer had gone from being mostly open in the past (as they still are over central Italy) to being closed on the majority of the coasts nowadays. For more details, see Millán (2014).

Millán, M. M., M. J. Estrela, M. J. Sanz, E. Mantilla, M. Martín, F. Pastor, R. Salvador, R. Vallejo, L. Alonso, and G. Gangoiti. (2005) Climatic feedbacks and desertification: the Mediterranean model. *Journal of Climate* 18:684-701.

Millán, M.M. (2014) Extreme hydrometeorological events and climate change predictions in Europe. *Journal of Hydrology* 518, Part B, 206-224.