1 Secondary dispersal of seagrass seeds in complex micro-topographies.

- 2 A. Alvarez,
- 3 Department of Marine Ecology (MARE), Instituto Mediterraneo de Estudios Avanzados-
- 4 IMEDEA (CSIC-UIB), C/ Miquel Marques 21, 07190 Esporlas,
- 5 Spain, e-mail: alberto.alvarez@imedea.uib-csic.es
- 6
- 7

8 Abstract

Motivated by observational and experimental evidence, a theoretical model is proposed 9 to relate the secondary dispersal of seagrass seeds with the complexity of micro-10 topography in natural environments. Complexity is encoded in terms of the Hurst 11 12 exponent of a fractal description of the micro-topographical geometry. The percentage of 13 a seafloor transect where secondary dispersal of seagrass seeds occurs, is quantified in terms of the mainstream velocity, bottom complexity and properties of the seeds. 14 Theoretical expressions are validated considering the cases of Zostera marina and 15 Posidonia oceanica seeds and using computational fluid dynamics (CFD). A total of 200 16 17 CFD simulations with different bottom complexities and flow conditions, were done for each seagrass genus to validate the theoretical model. Numerical results agree with 18 theoretical predictions. This finding provides a guideline to artificially adapt seafloor 19 roughness in seed-based restoration areas to control secondary dispersal of seeds. 20

21

22

23

24 Keywords: seagrass meadows; seed dispersal; micro-topography; fractals;

25 computational fluid dynamics.

27 **1. Introduction**

Seagrasses are marine aquatic angiosperm widely distributed through the coastal regions 28 across the globe (Orth et al., 2006a). Seagrass meadows are ecologically important 29 because they constitute the habitat of diverse species (Duffy, 2006; Heck et al., 2008) and 30 improve the water quality conditions in coastal regions (Denninson et al., 1993; 31 McGlathery et al., 2007). Moreover, it has been hypothesized that they can represent an 32 33 important carbon sink (Duarte et al., 2005; Fourqurean et al., 2012) and nitrogen removal (Zarnoch et al., 2017) in marine areas. Unfortunately, anthropogenic disturbances in the 34 coastal environments have resulted in a reduction of seagrass populations (Duarte, 2002; 35 36 Orth et al., 2006a; Waycott et al., 2009). Understanding the different aspects of the ecological dynamics of the seagrass meadows, is required for their present and future 37 preservation. 38

Dispersal is a determinant aspect of the ecological and evolutionary processes in seagrass 39 40 meadows (Orth et al., 2006b). In particular, seed dispersal shapes their distribution, 41 structure and resilience (Les et al., 2003). Two phases characterize the dispersal of seagrass seeds: the primary dispersal refers to the movement of the seed from the parent 42 plant to the sediment surface. Transport by water currents seems to be the dominant 43 44 primary dispersal mechanism of seagrass seeds (MacMahon et al., 2014). A long-distance dispersal results in this phase when seeds are positive buoyant or the movement is 45 46 mediated through rafting on the water surface of spathes, flowering branches (rhipidia) and reproductive shoots (Harwell and Orth, 2002; Kallstrom et al., 2008; Kendrick et al., 47 2012; McMahon et al., 2014; Hosokawa et al., 2015). Observational evidence suggests 48 49 that the primary dispersal distances of seagrass seeds of as much as 150 km, are possible (Harwell and Orth, 2002; Olsen et al., 2004; Kallstrom et al., 2008). Instead, a short 50

distance primary dispersal around the parent population occurs in seagrass genera
producing negative buoyant seeds (Orth et al., 1994).

53 The movement from the initial point of settlement to other location is known as secondary 54 dispersal. This mobilization of negatively buoyant seeds from the sediment, occurs through abiotic processes (currents, waves, sediment movement) (Orth et al., 1994) 55 56 and/or biotic vectors (Orth et al., 2000). However, hydrodynamically-mediated transport 57 seems to be the dominant mobility mechanism at the seafloor (Marion and Orth, 2010). To this regard, the relationship between secondary dispersal and hydrodynamic 58 conditions at the seafloor needs further research (Koch et al., 2010; Ruiz-Montoya et al., 59 60 2012).

The transport of negatively buoyant seeds as bedload, occurs when the shear stress at the 61 62 seabed exceed some critical threshold (named as critical erosion shear stress or Shields parameter when it is dimensionless (Ternat et al., 2008)). The bottom shear stress is a 63 64 function of the currents, the grain size and the micro-topographic features such as sand 65 ripples or bioturbation structures (Ackerman and Hoover, 2001; Jumars and Nowell, 1984). Regarding the latter, observational and flume studies found that small bottom 66 surface irregularities appear to be accumulation sites for the seagrass seeds. Orth et al. 67 68 (1994) consistently observed patterns of Zostera marina seeds trapping within a region of deposition, despite the existence of hydrodynamic conditions favorable for their 69 70 transport. These authors suggested that micro-topography reliefs on the seafloor armored seeds from ambient flow. Luckenbach and Orth (1999) observed in flume experiments, 71 72 the retention of Z. marina seeds by micro-topographic reliefs generated by benthic 73 invertebrates. Seeds which should be widely disperse as bedload were trap even by small non-uniformities at the seafloor. Inglis (2000) also evidenced that deposition of seeds of 74 75 Halodule uninervis occurred in greatest abundance in the troughs of small sand waves and

in the trails excavated by dugongs. Balestri et al. (2017) analysed published and 76 77 unpublished data of recruitment of Posidonia oceanica to examine, among other factors, the habitat characteristics at recruitment sites. The authors discovered that recruitment on 78 79 rocky habitats was twofold more than on sandy substrates. Pereda-Briones et al. (2018) found in flume experiments that complex substrata, are more capable of retaining 80 81 different seagrass seeds. These observations and flume experiments suggest an important 82 correlation between the complexity of the micro-topography and the dispersal and colonization of habitats by seagrasses. Mathematical formalization of this relationship 83 would contribute to further understand the ecological implications of secondary dispersal, 84 85 as well as it could have practical applications in seed-based restoration where secondary dispersal is a limiting factor (Marion and Orth, 2010). 86

Micro-topography refers to the scale of roughness concerned with small-scale 87 sedimentary structures such as pebble clusters, transverse ribs, pits, mounds, burrows etc. 88 (Robert, 1991). In terms of flow resistance processes, the distribution of sizes and shapes 89 90 of these micro-topographic features increases the energy loss of the flow. Characterizing the hierarchy of roughness elements of the micro-topography, remains an important 91 92 difficulty. Experimental quantification of small-scale bottom features has been attempted 93 with high-resolution sonar mapping systems, stereo photogrammetry and microtopographic laser scanners (Briggs, 1989; Lyons and Pouliquen, 2004; Du Preez and 94 95 Tunnicliffe, 2012). Still, the lack of a robust approach to characterize in situ the seafloor micro-topography is repeatedly cited as a source of uncertainty in derived analysis (Lyons 96 97 and Pouliquen, 2004).

Micro-topography often demonstrates self-similar fractal-like properties (Fox and Hayes,
1985). For this reason, fractal analysis has provided a particularly fruitful methodology
to characterize the complexity of roughness (Xu et al., 1993; Smith, 2014). A scaling

parameter in fractals known as the Hurst exponent (H), is of special interest for the present 101 study. An increase of a factor r in the horizontal scales, corresponds to an increment of 102 the vertical scales by a factor r^{H} . Since horizontal scales at the seafloor increases more 103 rapidly than vertical lengths, $0 \le H \le 1$. The fractal dimension of a profile (D) is related 104 to the Hurst exponent, by the relationship D=2-H (Smith, 2014). Thus, decreasing H 105 increases the degree of roughness of a fractal profile. Moreover, the power spectrum of a 106 fractal profile, β , is also related to H by β =2H+1. The use of a power law form to 107 characterize bottom roughness over a wide range of spatial frequencies, is a useful and 108 109 popular method (Fox and Hayes, 1985; Lyons et al., 2002). Even though the seafloor is continually changing, the previous statistical description is assumed to be relatively stable 110 (Lyons and Pouliquen, 2004). 111

Motivated by the experimental findings previously described and by its potential practical 112 113 applications, this article explores the relationship between the secondary dispersal of 114 seagrass seeds by currents and the complexity of micro-topographic profiles. Specifically, 115 the present objective is to quantify in terms of the Hurst exponent and flow conditions, the cumulative length in a fractal micro-topographic profile where secondary dispersal of 116 seagrass seeds occurs. A simple theoretical framework is first developed to derive general 117 scaling dependences for practical use. Theoretical expressions are then validated using 118 119 numerical techniques from computational fluid dynamics (CFD). Specifically, the conditions for secondary dispersal of seeds of Z. marina and P. oceanica on micro-120 121 topographic profiles are simulated for different fractal dimensions of the bottom profile 122 and flow conditions. The paper is organized as follows: Section 2 present the theoretical 123 development while details about the numerical fluid model are described in Section 3. Results are reported and discussed in Sections 4 and 5, respectively. 124

126 **2. Methods**

127 **2.1 Theoretical method**

128 An analytic treatment is proposed here to provide simple predictive expressions to 129 estimate the portion of a complex micro-topographic transect, where secondary dispersal 130 of seeds occurs. The analytical approach starts considering an isolated roughness element 131 placed in a flow channel with a smooth bottom. The bottom shear stress, τ_B , at the element 132 is partitioned into two components (Einstein and Banks, 1950):

133

 $134 \quad \tau_B = \tau_F + \tau_S \tag{1}$

135

136 Where τ_F and τ_S are the form and the skin friction shear stresses, respectively. The former 137 results from pressure differences due to local flow separation and re-circulation in the lee 138 region of the roughness element. These pressure gradients are normal to the surface and 139 thus the form drag shear stress has no impact on the bedload transport. The latter, results 140 from the skin friction of the water and the bed floor and it is the responsible of the 141 transport of the bed sediments.

142 An effective reduction of the skin friction shear stress, is expected at the lee of the 143 roughness element as a result of its wake. Its average value over the length of the 144 roughness element, $\bar{\tau}_{S}$, is expected to be attenuated according:

145

146 $\bar{\tau}_S = \tau_o \theta$ (2)

148	Where τ_o is the unperturbed skin friction shear stress in similar flow conditions. τ_o is		
149	assumed to exceed the critical value τ_{c} to initiate the motion of the seeds and thus,		
150	secondary dispersal of seeds would occur in the entire bottom transect if unperturbed. θ		
151	is a dimensionless parameter smaller than one, that is related to an effective unsheltered		
152	length of the roughness element. For a rough bottom constituted by a streamwise array of		
153	roughness elements with similar morphology, the resulting skin friction shear stresses at		
154	the n -th element can be approximated by superimposing n individual bottom stresses		
155	(Raupach, 1992; Raupach et al., 1993):		
156			
157	$\bar{\tau}_S = \tau_o \theta^n \tag{3}$		
158			
159	According to Eq. (3), $\bar{\tau}_{S}$ equals the critical value τ_{c} after n_{c} roughness elements	, with n _c	
160	verifying:		
161			
162	$\bar{\tau}_S = \tau_c = \tau_o \theta^{n_c} \tag{4}$		
163			

or,

 $n_c = \frac{1}{\gamma} \ln \left(\frac{\tau_o}{\tau_c} \right)$ (5)

168 Where γ =-In(θ) is a positive constant (θ <1). No movement of seeds would be expected 169 after a streamwise distance L_c=n_cl^{*} over the rough bottom, where l^{*} is the characteristic 170 length of the roughness element. The length of the transect in which the skin friction shear 171 stress exceeds the critical value can then be estimated as L_s= θ L_c. Using Eq.(5):

172

173
$$\frac{L_s}{L} = \frac{n_c \, l^* \theta}{L} = \frac{\theta}{\gamma} \frac{l^*}{L} \ln\left(\frac{\tau_o}{\tau_c}\right) = \frac{2e^{-\gamma}}{\gamma} \frac{l^*}{L} \ln\left(\frac{U_o}{U_c}\right) \tag{6}$$

174

Where U₀ is the actual free stream velocity and U_c corresponds to the free stream velocity 175 that induces a bottom stress τ_c in the unperturbed transect. The last term in Eq. (6) is 176 derived from the relationship $\tau_{o,c} = \frac{1}{2}\rho C_f U_{o,c}^2$, where ρ and C_f are the density of the fluid 177 and the skin friction coefficient of the smooth bottom, respectively. In a fractal bottom 178 179 transect, a characteristic length of the roughness relief (1^*) is provided by the correlation length, which is estimated by the ratio between the length of the transect (L) and number 180 (N) of random reliefs (Rees and Arnold, 2006). Statistically, N corresponds to the number 181 182 of positive zero crossings in the transect (Munro, 1989). In terms of the Hurst exponent, the parameters above follow the relationships (Church, 1988): 183

184

185
$$L = N l^*; N = \frac{(4H+1)}{2H^2}; \frac{l^*}{L} = \frac{2H^2}{(4H+1)}$$
 (7)

186

187 In this analytical framework, Eqs.(6) and (7) provide an estimate of the cumulative length 188 of the transect with secondary dispersion, L_s/L , in terms of the free stream velocity (U_o), 189 the characteristics of the roughness elements (H) and the conditions for secondary 190 dispersal of the seeds (U_c) .

191

192 **2.2 Numerical method**

193 A numerical model has been developed to test the theoretical predictive expressions obtained in the previous Section. To do that, the flow over a rough bottom is simulated in 194 195 a flume two dimensional section of 0.5 m height and 1.2 m long. Assuming a mainstream symmetry is common in flume studies, because cross flows are negligible with respect to 196 197 the mainstream. While this assumption captures the major features of the physical processes involved, it significantly reduces the computational efforts. With suitable 198 199 boundary conditions (detailed later), the domain represents a portion of an open flume 200 arbitrarily long and deep enough to disregard the effects of a free surface. The rough 201 bottom is 1 m long and it is located at the center of the flume, 0.1 m apart from the inflow 202 and outflow boundaries. Realizations of fractal micro-topographies, were generated as 203 one dimensional fractional Brownian motions with Hurst parameters of 0.2, 0.4, 0.5, 0.6 204 and 0.8 (Figure 1). Despite to the scarce data, diverse observational and experimental 205 studies have characterized the root mean square (rms) height roughness of the microtopography with values as high as 0.03 m (Briggs, 1989; Wiberg and Harris, 1994; Inglis, 206 207 2000; Lyons et al., 2002; Guillen et al., 2008). Accordingly, the rms height (h) of each 208 profile was fixed to 0.03 m.

Numerical simulations are particularized for the seeds of *Z. marina* and *P. oceanica*, which are marine aquatic angiosperm widely distributed through the coastal regions of the Northern Hemisphere (Short and Moore, 2006). The former represents a case where seeds are easily transported as bedload (Luckenbach and Orth, 1999), while the latter

exemplifies a case where the secondary dispersal of seeds by usual ambient flows is rather 213 limited (Pereda-Briones et al., 2018). Specifically, the critical shear velocity, $u_c^* = u_c^*$ 214 $\frac{\tau_c}{\rho}$ where τ_c y ρ are the critical bottom shear stress and the fluid density, respectively), 215 to initiate the motion of Z. marina seeds on a smooth sand bed is $u_c^*=0.007$ m/s, which 216 resulted from a critical free stream velocity of U_c=0.1 m/s (Orth et al., 1994). Free stream 217 velocities Uo, of 0.15 m/s, 0.2 m/s, 0.25 m/s and 0.3 m/s were considered in the 218 simulations. This range of speeds are typical for the locations where Pereda-Briones et al. 219 (2018) collected Z. marina seeds. The transport of P. oceanica seeds as bedload is 220 221 significatively harder than for the Z. marina seeds. A critical free stream velocity of U_c=0.2 m/s is required to initiate the motion of *P. oceanica* seeds on a smooth sand 222 (Pereda-Briones et al., 2018). For this reason, free stream velocities U_o, of 0.25 m/s, 0.3 223 m/s, 0.35 m/s and 0.4 m/s were considered in this case. 224

A Reynolds Average Navier–Stokes (RANS) approach along with a turbulence model
have been considered for the numerical simulations. Mathematically, the steady RANS
model is described by the equations:

228

229
$$\frac{\partial(U_i U_j)}{\partial x_j} = -\frac{1}{\rho} \frac{\partial P}{\partial x_i} + (v + v_T) \frac{\partial^2 U_i}{\partial x_j \partial x_j}$$
(8)

$$230 \quad \frac{\partial U_i}{\partial x_i} = 0 \tag{9}$$

231

where U_i is the *i-th* component of the time averaged velocity, ρ is the density, *P* is the pressure and v is the kinematic viscosity (10⁻⁶ m²/s). The Einstein summation convention is assumed for repeated indices. The eddy viscosity v_T which parametrizes the dispersal effect of the small scales onto the large scales, is computed with the Baldwin-Lomax eddy viscosity model (Baldwin and Lomax, 1978; Wilcox, 1993; Appendix 1). This model is
widely employed for its ease of implementation and robustness. It faithfully reproduces
friction velocity and velocity profiles for incompressible turbulent boundary layers,
providing reasonable accuracy for steady flows with no or mild separation (Willcox,
1993). The latter is the situation expected in the present study.

241 The system of Equations (8) - (9) is solved with the following boundary conditions:

242

	$\mathbf{U} = \mathbf{U}(\mathbf{y}), \mathbf{V} = 0, \ \mathbf{n} \cdot \nabla \mathbf{P} = 0$	Inflow boundary
	$\mathbf{n} \cdot \nabla \mathbf{U} = \mathbf{n} \cdot \nabla \mathbf{V} = 0, \mathbf{P} = 0$	Outflow boundary
243	$\mathbf{n} \cdot \nabla \mathbf{U} = \mathbf{n} \cdot \nabla \mathbf{V} = \mathbf{n} \cdot \nabla \mathbf{P} = 0$	Top boundary
	$\mathbf{U} = \mathbf{V} = \mathbf{n} \cdot \nabla \mathbf{P} = 0$	Bottom boundary

244

245 where \mathbf{n} is the vector normal to the boundary. A logarithmic inlet velocity profile with components U(y) = u_e^* In (9.793 y u_e^*/v)/K (where u_e^* is a theoretical estimate of the 246 247 friction velocity u^{*} based on the freestream velocity and K=0.4 is the von Karman's constant) and V=0 is prescribed at the left boundary. This inflow condition facilitates the 248 convergence of the numerical algorithm. The outflow boundary condition is of the do-249 nothing type (Gresho, 1991), allowing the fluid to cross the boundary in either direction 250 to preserve continuity. Slip and non-slip boundary conditions are prescribed for the 251 252 velocity field at the top and bottom boundaries, respectively. This type of boundary conditions allows us to consider the computational domain as part of a larger open flume. 253 254 A standard Galerkin finite element method based on the triangular Taylor-Hood element (Taylor and Hood, 1973), has been employed for the spatial discretization of Equations 255 (8)-(9). In the Taylor-Hood element, pressure and velocity nodes are located at the 256

vertices and at the vertices and middle-edge points of the triangular element, respectively. 257 Thus, the velocity and pressure discrete approximations are both piecewise continuous, 258 259 the former being quadratic and the latter linear (Donea and Huerta, 2003). The Taylor-Hood elements are stable in that they fulfill the inf-sup compatibility condition (a 260 261 condition that discrete spaces must satisfy to guarantee the stability of the results, Brezzi and Bathe, 1990). They also exhibit optimal quadratic convergence (Donea and Huerta, 262 263 2003). For these reasons, the Taylor-Hood finite element is the standard finite element 264 for simulating incompressible fluid flow (Larson and Bengzon, 2013).

A Picard iteration technique was used to solve the nonlinear equations resulting after the spatial discretization. Starting with an initial guess for the velocity field, Picard's method constructs a sequence of approximate solutions (\vec{U}^{k+1}, P^{k+1}) by solving at each iteration the linear Oseen problem (Rhebergen et al., 2013):

269

270
$$\frac{\partial \left(U_i^{k+1}U_j^k\right)}{\partial x_j} + \frac{1}{\rho} \frac{\partial P^{k+1}}{\partial x_i} - \left(v + v_T^k\right) \frac{\partial U_i^{k+1}}{\partial x_j \partial x_j} = g_i$$
(10)

$$271 \quad \frac{\partial U_i^{k+1}}{\partial x_i} = 0 \tag{11}$$

272

273 Where superscript k=1,2... in the model variables refers to the solution at the *k-th* 274 iteration. g_i collects the boundary forcing terms prescribed by the inflow and boundary 275 conditions. The convergence criterion used in this work to stop the iteration process is:

277
$$\frac{\|\vec{v}^{k+1} - \vec{v}^k\|}{\|\vec{v}^k\|} < 10^{-3}$$
(12)

This convergence criterion resulted on a good compromise between accuracy andcomputational cost.

An adaptive mesh which is refined in the vicinity of the bottom boundary, was constructed in order to capture fine features of the flow. The domain geometry was tessellated into 26331 triangular elements with characteristic sizes ranging from 1.5×10^{-2} ⁴ m at the seafloor up to 2×10^{-2} m at the upper boundary (Figure 2). Regarding the resolution of the roughness elements, more than 10 grid points resolve the roughness in wall-normal and streamwise directions. It has also been verified that the mesh is dense enough to resolve the structure of the boundary layer (see below).

The computational methodology to generate fractal profiles of the seafloor has been detailed in previous articles (Kroese and Botev, 2015), to which interested readers are referred for further information.

291

292 **3 Results**

293 A numerical simulation was initially done considering a smooth bottom, to validate the numerical implementation as well as the suitability of the mesh resolution. The structure 294 295 of the turbulent boundary layer of a viscous flow close to a smooth wall is well known 296 from scaling considerations (Wilcox, 1993). Briefly, three distinct regions are observed 297 namely viscous sublayer, log layer and defect layer. Only the first two regions are relevant for this study. The viscous sublayer is the closest to the wall ($y^+ < 11$ where $y^+ = y u^*/v$ is 298 299 the dimensionless distance from the wall in wall coordinates) and the flow dynamics is mainly controlled by viscous forces. The dimensionless velocity $U^+=U/u^*$ and distance to 300 the wall y^+ follows the relation $U^+=y^+$ in this sublayer. The log layer (25 $< y^+ < 1000$) is 301

found after the viscous sublayer. The dynamics is mainly dictated by nonlinear interactions and the average turbulent velocity at a particular point is proportional to the logarithm of the distance of the point under consideration from the wall, U⁺=In(9.793 y^+)/K.

306 Figure 3 displays the structure of the boundary layer in the numerical channel, obtained from an inflow with a free stream velocity of Uo=0.2 m/s. Results evidence that the 307 308 numerical model nicely reproduces the expected structure of the turbulent boundary layer. Further comparison between the numerical results and the theory is done in Figure 4. 309 Specifically, the Figure compares the theoretical vertical profile of U⁺ with the numerical 310 311 one in a station at the middle of the channel. Again, a good agreement is found with the 312 theoretical profile at the viscous sublayer and log layer. Notice the existence of a transition region between both layers, $11 < y^+ < 20$, where theoretical scaling arguments are 313 314 not applicable. The Figure also compares the inflow profile introduced as boundary 315 condition, with the theoretical profile corresponding to the structure of the turbulent 316 boundary layer. This comparison reveals that the inflow boundary condition is well adjusted to the correct scaling in the external log layer. Conversely, the inflow profile 317 significantly differs from the correct one in the viscous sublayer. 318

319 A total of 200 numerical simulations with the CFD model considering different fractal bottom profiles and free stream velocities, were done for each seagrass genus to validate 320 321 the theoretical model. Specifically, an ensemble of 10 random realizations of microtopographic profiles were considered for each Hurst exponent (H) and free stream 322 velocity (U_0). For each realization, the bottom friction velocity, $u^*(x)$, was numerically 323 324 computed along the bottom profile and the cumulative length (L_S) of the regions where 325 $u^{*}(x) > u_{c}$, calculated. Finally, the cumulative length for secondary dispersal of Z. marina seeds for a given H and U_0 , was obtained as the ensemble average $\langle L_S \rangle$ of the cumulative 326

lengths of the individual realizations. Figure 5 displays the relationship between $\langle L_S \rangle / L$ 327 and $l^{*}In(U_{o}/U_{c})/L$ obtained for the considered values of H and U_o and each seagrass genus. 328 A straight line with a slope of 2.67 and a regression coefficient of 0.983, results from the 329 linear fit of the data points for the Z. marina. A value of θ =0.754 is derived from this 330 linear fit. For the case of the *P. oceanica*, the slope and regression coefficients are 2.77 331 and 0.986, respectively, with a value of θ =0.760. Finally, a slope of 2.69, a correlation 332 333 coefficient of 0.984 and a θ value of 0.755 results from the linear fitting of the data regardless of the seagrass genus. The latter fit is displayed in Figure 5. Numerical results 334 335 confirm the validity of the relationship Eq. (6) obtained from the analytical study.

336

337 **4 Discussion**

This study provides a quantification of the role of micro-topography in the secondary 338 dispersal of seagrass seeds. An estimate of the portion of a rough bottom where secondary 339 340 dispersal occurs was obtained based on fundamental arguments. Specifically, two physical assumptions are considered when developing the theory. The first hypothesis 341 considers the partition of the total shear stress between that arising from pressure 342 343 differences over and around roughness elements (form drag) and the frictional drag on the intervening surface. While a wide consensus exists on the arithmetic partition of the 344 345 total shear stress, different methodologies have been suggested to estimate its components in terms of the flow conditions and bottom properties (interested readers are referred to 346 Le Bouteiller and Venditti, (2015) for a concise review). To this regard, a second 347 hypothesis considers that the average frictional stress over a roughness element, is 348 expected to be smaller than the unobstructed stress τ_0 . This results from the negligible 349 contribution to the skin friction of the front and rear separation zones or the standing 350

351 eddies, located on the roughness element. Moreover, the attenuation coefficient (θ) is assumed independent of the characteristic length of the roughness element (1^{*}). This 352 353 situation is expected if roughness elements are significantly embedded in the wakes of its 354 neighbors. Experiments and numerical simulations evidence that at high Reynolds numbers, the turbulent boundary layer reattaches at a distance of about 6h after bedforms 355 356 of characteristic height h (Nelson and Smith, 1989). Thus, the independence of θ with l^* would suggest that l^* is less or of the same order than 6h. Briggs (1989) provides some 357 observational evidence that this could be the case for relatively rough micro-topography. 358 This author measured a correlation length of 0.085 m in a rough bottom with rms height 359 360 (h) of 0.025 m. This condition also holds in most of the fractal descriptions of the microtopography used in the simulations and derived from seafloor parameters reported in 361 362 observations. Nevertheless, no observational nor experimental study has yet addressed 363 how general this condition is in natural environments.

364 Theoretical results have been compared against numerical simulations of different flow regimes over fractal micro-topographies. Numerical results agree with theoretical 365 expectations. Specifically, the ensemble average length where secondary dispersal occurs 366 367 in a fractal micro-topography, was found linearly related with the product of the 368 characteristic length of the roughness elements times the logarithm of the ratio between 369 the free stream and critical velocities. As expected from the theory, a similar value of the 370 attenuation coefficient θ of about 0.75 has been numerically obtained for both cases of Z. marina and P. oceanica. This value can be intuitively understood considering that 371 372 conditions for secondary dispersal statistically occurs at the exposed front part of the 373 roughness element. Moreover, the result also confirms the independence of θ with 1^{*}. To 374 this regard, CFD simulations have been revealed as a useful tool to test theoretical 375 arguments in virtual physical environments complex enough to represent the physical

processes of interest, but simple enough to facilitate statistical inferences. Certainly,
numerical modelling technologies cannot fully represent reality and thus, present results
encourage their future experimental and observational confirmation.

379 Theoretical and numerical results contribute to clarify and quantify the role of micro-380 topography in the entrapment of seagrass seeds in natural environments. Observational 381 evidence supports the hypothesis that the secondary dispersal of these seeds is limited to 382 the immediate vicinity of seed release. This limited secondary dispersal may be interpreted in terms of the micro-topography complexity. Vegetation significantly 383 384 attenuate the capability of the ambient flow to disperse seeds that fall within the seagrass 385 meadow (Orth et al., 1994). Moreover, a high micro-topography complexity (low Hurst exponent) is expected inside the seagrass meadows, due to secondary flows resulting from 386 the interaction of the mainstream with the body of the plant as well as due to plant litter. 387 The micro-topography is expected to play a more significant role in the secondary 388 dispersal of seeds that land within open vegetation or exposed soils. Significant seed 389 390 accumulations can be found in these regions even with supercritical flows if small-scale disturbances are present (Orth et al., 1994; Inglis, 2000). 391

392 Seeds can be an important mechanism to successfully restore large areas of seagrass 393 coverage (Pickerell et al., 2005; Orth et al., 2006c). For seagrass species producing abundant and easy-to-harvest seeds, seed-based restoration introduces significant 394 395 technological and economic advantages versus restoration techniques relying on adult plants, (Marion and Orth, 2010). However, low initial seedling establishment rates have 396 397 been recognized as a major limitation for seed-based restoration projects (Marion and 398 Orth, 2010). Seed retention and seedling establishment can be facilitated by artificially 399 adapting seafloor roughness. Natural or artificial roughness elements could be used, to 400 engineer appropriate mulches to trap naturally dispersed seeds and to enhance seed 401 retention. Present results provide a guideline to determine, a priori, the required 402 complexity of the micro-topographical mulch to satisfy, given seed properties and 403 expected environmental conditions, a desired degree of seed retention in the habitat to be 404 restored.

405 The findings of the present work could also be applied, with certain caution, to land systems. This will work when wind is the dominant dispersal agent. Observations also 406 407 evidence that wind-blown seeds in land systems, are more easily trapped on rough soils than on smooth ones (Johnson and Fryer, 1992; Chambers and MacMahon, 1994). This 408 study provides a physical foundation for these observations. Notice, however, that 409 410 secondary dispersal of seeds in land systems often results from overland flows generated by intense precipitations (Chambers and MacMahon, 1994). Although being a 411 hydrodynamically-mediated transport of seeds, present findings are not directly 412 applicable to this case. The proximity of the air-water interface to the soil and the usual 413 high speeds of overland flows, may invalidate some of the assumptions considered in the 414 415 theory.

416 To conclude, this study has proposed and numerically validated a link between the smallscale properties of micro-topography (H or l^*) and the global bed characteristics (<Ls>) 417 418 for secondary dispersal of seagrass. Although there are some observational and/or experimental evidences that would support the results, their confirmation by dedicated 419 420 observations and/or experiments still remains open. The proposed relationship constitutes a valuable tool to understand the dispersal and colonization of habitats by seagrasses. To 421 422 this regard, it provides the physical explanation to observational evidence. It could also 423 support seed-based restoration techniques.

424 **References**

- 425 Ackerman, J. D., Hoover, T. 2001. Measurement of local bed shear stress in streams using
- 426 a Preston-static tube. *Limmol. Oceanogr.*, 46, 2080-2087.
- 427 Baldwin, B. S., Lomax, H. 1978. Thin Layer Approximation and Algebraic Model for
- 428 Separated Turbulent Flows, *AIAA*, 78-257.
- 429 Balestri, E., Vallerini, F., Lardicci, C. 2017. Recruitment and patch establishment by seed
- 430 in the seagrass Posidonia oceanica: Importance and conservation implications. *Front*.
- 431 *Plant Sci.*, 8, 1067.
- Brezzi, F., Bathe, K. J. 1990. A discourse on the stability conditions for mixed finite
 element formulations. *Comp. Methods Appl. Mech. Eng.*, 82, 27-57.
- Briggs, K. B. 1989. Microtopographical roughness of shallow water continental shelves. *IEEE J. Ocean. Eng.*, 14, 360-367.
- 436 Chambers, J. C., MacMahon, J. A. 1994. A day in the life of a seed: movements and fates
- 437 of seeds and their implications for natural and managed systems. *Ann. Rev. Ecol. Sys.*, 25,
 438 263-292.
- 439 Church, L. E. 1988. Fractal surface finish. Appl. Opt., 27, 1518-1526.
- 440 Denninson, W. C., Orth, R. J., Moore, K. A., Stevenson, J. C., Carter, V., Kollar, S. et al.
- 441 1993. Assessing water quality with submersed aquatic vegetation: Habitat requirements
- 442 as barometers of Chesapeake Bay health. *Bioscience*, 43, 86-94.
- 443 Donea J., Huerta A. 2003. Finite element methods for flow problems. Wiley and Sons
- 444 Ltd., Chichester, West Sussex, UK, pp. 350.
- 445 Duarte, C. M. 2002. The future of seagrass meadows. *Environ Conserv.*, 29, 192–206.

- Duarte, C. M., Middelburg, J., Caraco, N. 2005. Major role of marine vegetation on the
 oceanic carbon cycle. *Biogeoscience*, 2,1–8.
- 449 Duffy, J. E. 2006. Biodiversity and the functioning of seagrass ecosystems. *Mar. Ecol.*
- 450 *Prog. Ser.*, 311, 233–250.
- 451 Du Preez, C., Tunnicliffe, V. 2012. A new video survey method of microtopographic
 452 laser scanning (MiLS) to measure small-scale seafloor bottom roughness. *Limnol.*453 *Oceanogr.: Methods*, 10, 899–909.
- Einstein, H., A., Banks, R. B., 1950. Fluid resistence of composited roughness. *Trans. AGU*, 31, 603-610.
- 456 Fourqurean, J. W., Duarte, C. M., Kennedy, H., Marba, N., Holmer, M., Mateo, M. A.,
- 457 Apostolaki, E. T., Kendrick, G. A., Krause-Jensen, D., McGlathery, K. J., 2012. Seagrass
- 458 ecosystems as a globally significant carbon stock. *Nat. Geosci.*, 5, 505-509.
- 459 Fox, C. G., Hayes, D. E., 1985. Quantitative methods for analyzing the roughness of the
- 460 seafloor. *Rev. Geophys.*, 23, 1-48.
- Gresho, P. M., 1991. Some current CFD issues relevant to the incompressible NavierStokes equations. *Comput. Methods Appl. Mech. Eng.*, 87, 201–252.
- Guillén, J., Soriano, S., Demestre, M., Falqués, A., Palanques, A., Puig, P., 2008.
- 464 Alteration of bottom roughness by benthic organisms in a sandy coastal environment.
- 465 *Cont. Shelf Res.*,28, 2382-2392,
- 466 Harwell, M. C., Orth, R. J., 2002. Long-distance dispersal potential in a marine
 467 macrophyte. *Ecol.*, 83, 3319-3330.

- 468 Heck Jr., K. L., Carruthers, T. J. B., Duarte, C. M., Hughes, A. R., Kendrick, G., Orth, R.
- 469 J., Williams, S. W., 2008. Trophic transfers from seagrass meadows subsidize diverse
- 470 marine and terrestrial consumers. *Ecosystems*, 11, 1198-1210.
- 471 Hosokawa, S., Nakaoka, M., Miyoshi, E., Kuwae, T., 2015. Seed dispersal in the seagrass
- 472 Zostera marina is mostly within the parent bed in a protected bay. Mar. Ecol. Prog. Ser.,
- **473 523**, **41-56**.
- 474 Inglis, G. J.,2000. Disturbance-related heterogeneity in the seed banks of a marine
 475 angiosperm. *J. Ecol.*, 88, 88–99.
- 476 Johnson, E. A., Fryer, G. I., 1992. Physical characterization of seed microsites-movement
- 477 on the ground. J. Ecol., 80, 823-836.
- Jumars, P. A., Nowell, A. R., 1984. Effects of benthos on sediment transport: difficulties
 with functional grouping. *Cont. Shelf Res.*, 3, 115-130.
- 480 Kalstrom B., Nyqvist, A., Aberg, P. Bodin, M, André, C., 2008. Seed rafting as a dispersal
- 481 strategy for eelgrass (*Zostera marina*). *Aquat. Bot.*, 88, 148-153.
- 482 Kendrick, G. A., Waycott, M., Carruthers, T. J. B., Cambridge, M. L. and others, 2012.
- The central role of dispersal in the maintenance and persistence of seagrass populations. *Bioscience*, 62, 56-65.
- Koch, E. W., Ailstock, M. S., Booth, D. M., Shafer, D. J., Magoun, A. D., 2010. The role
- of currents and waves in the dispersal of submersed angiosperm and seedlings. *Restor*. *Ecol.*, 18, 584-595.
- 488 Kroese, D. P., Botev, Z. I., 2015. Spatial Process Simulation. In: Stochastic Geometry,
- 489 *Spatial Statistics and Random Fields*, ed. Schmidt, V. Springer, Cham, pp. 369-404.

- 490 Larson, M. G., Bengzon, F., 2013. The finite element method: Theory, implementations
- 491 *and applications*. Springer, Heidelberg, pp. 385.
- 492 Le Boutieller, G., Venditti, J. G., 2015. Sediment transport and shear stress partitioning
- 493 in a vegetated flow. *Water Resour. Res.*, 51, 2901-2922.
- 494 Les, D. H., Crawford, D. J., Kimball, R. T., Moody, M. L., Landolt, E. 2003.
- Biogeography of discontinuously distributed hydrophytes: A molecular appraisal of
 intercontinental disjunctions. *Int. J. Plant Sci.*, 164, 917-932.
- 497 Luckenbach, M. W., Orth R. J., 1999. Interactions between benthic infauna and the
- transport and burial of Zostera marina seeds. *Aquatic Bot.*, 62, 235-247.
- 499 Lyons, A., P, Fox, W. L. J., Hasiotis, T. Pouliquen, E. 2002. Characterization of the two-
- dimensional roughness of wave-rippled sea floors using digital photogrammetry. *IEEE J*. *Ocean. Eng.*, 27, 515-524.
- 502 Lyons, A. P., Pouliquen, E., 2004. Advances in high resolution seafloor characterization
- 503 in support of high-frequency underwater acoustics studies: techniques and examples.
- 504 Meas. Sci. Technol., 15, 59-72.
- 505 Marion, S. R., Orth, R. J., 2010. Innovative techniques for large-scale seagrass restoration
- using Zostera marina (eelgrass) seeds. Rest. Ecol., 18, 514-526.
- McGlathery, K. J., Sundback, K., Anderson, I. C., 2007. Eutrophication in shallow coastal
 bays and lagoons: The role of plants in the coastal filter. *Mar. Ecol. Prog. Ser.*, 348, 118.
- 510 McMahon, K., van Dijk, K., Ruiz-Montoya, L., Kendrick, G. A., Krauss, S. L., Waycott,
- 511 M., Verduin, J., Lowe, R., Statton, J., Brown, E., Duarte, C., 2014. The movement
- 512 ecology of seagrasses. *Proc. R. Soc. B*, 281, 20140878.

- Nelson, J. M., Smith, J. D., 1989. Mechanics of flow over ripples and dunes. *J. Geophys. Res.*, 94, 8146-8162.
- 515 Munro, D. S., 1989. Surface roughness and bulk heat transfer on a glacier; comparison

516 with eddy correlation. J. Glaciol., 35, 343-348.

- 517 Olsen et al. 2004. North Atlantic phylogeography and large-scale population
- differentiation of the seagrass *Zostera marina L. Mol. Ecol.*, 13, 1923-1941.
- 519 Orth, R. J., Luckenbach, M., Moore, K. A., 1994. Seed dispersal in a marine macrophyte:

520 implications for colonization and restoration. *Ecology*, 75, 1927-1939.

- 521 Orth, R. J., Harwell, M. C., Bailey, E. M., Barthlomew, A., Jawad, J. T., Lombana, A. V.,
- 522 Moore, K. A., Rhode, J. M., Woods, H. E., 2000. A review of issues in seagrass seed
- dormancy and germination: implications for conservation and restoration. *Mar. Ecol. Prog. Ser.*, 200, 277-288.
- 525 Orth, R. J., Carruthers, T. J., B., Denninson, W. C., Duarte, C. M., Fourqurean, J. W.,
- Heck Jr., K. L., *et al.* 2006a. A global crisis for seagrass ecosystems. *Bioscience*, 56, 986987.
- 528 Orth, J. R., Harwell, M. C., Inglis, G. J., 2006b. Ecology of seagrass seeds and seagrass
- 529 dispersal processes. In: *Seagrasses: Biology, ecology and conservation*, eds. Larkum, A.
- 530 W. D., Orth, R. J., Duarte, C. M. Springer, The Netherlands, pp. 111-133.
- 531 Orth, R. J., Luckenbach, M. L., Marion, S. R., Moore, K. A., Wilcox, D. J., 2006c.
- 532 Seagrass recovery in the Delmarva Coastal Bays, USA. *Aquat. Bot.*, 90, 204-208.
- 533 Pereda-Briones, L., Infantes, E., Orfila, A., Tomas, F., Terrados, J., 2018. Dispersal of
- seagrass propagules: interaction between hydrodynamics and substratum type. *Mar. Ecol.*
- 535 Prog. Ser., 593, 47-59.

- Pickerell, C. H., Schott, S., Wyllie-Echevarria, S., 2005. Buoy deployed seeding:
 demonstration of a new eelgrass (*Zostera marina L.*) planting method. *Ecol. Eng.*, 25,
 127-136.
- Raupach, M. R., 1992. Drag and drag partition on rough surfaces. *Bound.-Lay. Meteorol.*,
 5, 285-308.
- Raupach, M.R., Gillette, D.A., Leys, J. F., 1993. The effect of roughness elements on
 wind erosion threshold. *J. Geophys. Res.*, 98, 3023-3029.
- 543 Rees, W. G., Arnold, N. S., 2006. Scale-dependent roughness of a glacier surface:
- implications for radar backscatter and aerodynamic roughness modelling. *J. Glaciol.*, 46,
 445-452.
- 546 Rhebergen, R., Cockburn, B., Jaap, J. W. van der Vegt, 2013. A space-time discontinuous
- 547 Galerkin method for incompressible Navier-Stokes equations, *J. Comp. Phys.*, 233, 339–
 548 358.
- Robert A., 1991. Fractal properties of simulated bed profiles in coarse-grained channels. *Math. Geol.*, 23, 367-382.
- 551 Ruiz-Montoya, L., Lowe, R. J., Van Niel, K. P., Kendrick, G. A., 2012. The role of
- hydrodynamics on seed dispersal in seagrasses. *Limmol. Oceanogr.*, 57, 1257-1265.
- 553 Short, F. T., Moore, K. A., 2006. Zostera: Biology, ecology and management In:
- 554 Seagrasses: Biology, ecology and conservation, eds. Larkum, A. W. D., Orth, R. J.,
- 555 Duarte, C. M. Springer, The Netherlands, pp. 361-386.
- 556 Smith, M. W., 2014. Roughness in earth science. Earth-Sciences Rev., 136, 202-225.
- 557 Taylor, C., Hood, P., 1973. A numerical solution of the Navier-Stokes equations using
- the finite element technique. *Comput. Fluids*, 1, 73-100.

- 559 Ternat, F., Boyer, P., Anselmet, F., Amielh, M., 2008. Erosion threshold of saturated
- natural cohesive sediments: Modeling and experiments. *Water Resour. Res.*, 44, W11434.
- 561 Waycott et al. 1999. Accelerating loss of seagrasses across the globe threatens coastal
- 562 ecosystems. Proc. Natl. Acad. Sci. USA, 106, 12377-12381.
- 563 Wiberg, P. L., Harris, C. K., 1994. Ripple geometry in wave-dominated environments. J.
- 564 *Geophys. Res.*, 99, 775-789.
- Wilcox, D.C. (1993), *Turbulence modeling for CFD*. DCW Industries Inc., California,
 460 pp.
- Xu, T., Moore, I. D., Gallant, J. C., 1993. Fractals, fractal dimension and landscapes: a
 review. *Geomorphology*, 8, 245-262.
- Zarnoch, C. B., Hoellein, T. J., Bradley, T. F., Bradley, J. P. 2017. Eelgrass meadows,
- 570 Zostera marina (L.), facilitate the ecosystem service of nitrogen removal during
- simulated nutrient pulses in Shinnecock Bay, New York, USA. Mar. Poll. Bull., 124, 376-
- 572 387.

573 **Figure Captions**

574 Figure 1. Random realizations of micro-topographic transects for different Hurst575 exponents.

Figure 2. Finite element grid for a given micro-topography. The scope of the figure isintended to provide an understanding of the resolution of the mesh.

Figure 3. Scatter plot (grey dots) of the relationship between U^+ and y^+ obtained in the numerical simulation of a flow with free stream velocity $U_0=0.2$ m/s over a smooth surface. Black solid lines represent the theoretical relationship for different values of y^+ .

Figure 4. Comparison between the theoretical and numerical vertical profiles of U^+ in the station at the middle of the test section (circles) and at the inflow boundary (triangles).

Figure 5. Scatter plot of the numerical relationship found between $\langle L_S \rangle$ and $l^*In(U_o/U_c)$. Each circle (triangle) represents the ensemble average of L_S over 10 random realizations with a given H and U_o for the case of *Z. marina* (*P. oceanica*). Error bars correspond to the standard deviation in the ensemble. The grey straight line shows the best linear fit of the all data points.

588



590 Figure 1



Figure 2



592 Figure 3



593 Figure 4



594 Figure 5

595 Appendix 1: The Baldwin-Lomax eddy viscosity model

The Baldwin-Lomax eddy viscosity model, the most widely used algebraic model, assumes a two-layer concept wherein the boundary layer is split into inner and outer regions. Each layer is characterized by different turbulent length and velocity scales. The eddy viscosity in the outer region is formulated by the relation (Baldwin and Lomax, 1978; Wilcox, 1993):

601

$$602 \quad [v_T]_{outer} = \alpha C_{cp} F_{wake} F_{kleb} \tag{A.1}$$

603

604 With the closure coefficients α =0.0168 and C_{cp}=1.6. The outer function F_{wake} is: 605

606
$$F_{wake} = min\left(Y_{max}F_{max}, C_{wk}Y_{max}\frac{U_{dif}^2}{F_{max}}\right)$$
(A.2)

607

608 Where $C_{wk} = 0.25$ and U_{dif} is the difference between the maximum and minimum 609 velocities in the profile. $F_{max}=max(y\Omega D)$ with Ω being the magnitude of the local time 610 averaged vorticity and $D = \left(1 - e^{\frac{y^+}{A^+}}\right)$ is the Van-Driest damping factor (A⁺=26 and y⁺ = 611 y u^{*}/v, is the dimensionless distance from the wall in wall coordinates). Y_{max} is the value 612 of y where F_{max} occurs. The Klebanoff intermittency factor F_{kleb} is given by:

613

614
$$F_{kleb} = \frac{1}{1+5.5(C_{kleb}\frac{y}{Y_{max}})^6}$$
 (A.3)

615

616 With $C_{kleb}=0.3$. For the inner region, the eddy viscosity is given by the Prandtl-Van Driest 617 formulation:

$$[v_T]_{inner} = (KyD)^2 |\omega|$$
(A.4)

620

621 Where K=0.4 is the von Karman's constant and $|\omega|$ is the magnitude of the local vorticity.

- 622 The boundary of the outer and inner layers is established at each streamwise station by
- 623 the smallest value of y where $[v_T]_{outer}$ and $[v_T]_{inner}$ are equal.