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Valuable compounds extraction, anaerobic digestion and composting:

A leading biorefinery approach for agricultural wastes

Fernando G. Fermoso^{1*}, Antonio Serrano^{1,2}, Bernabé Alonso-Fariñas³, Juan Fernández-Bolaños¹, Rafael Borja¹, Guillermo Rodríguez-Gutiérrez¹

*¹Instituto de Grasa, Spanish National Research Council (CSIC), Campus Universitario Pablo de Olavide
– Ed. 46, Ctra. de Utrera, km. 1, Seville, Spain*

²University of Queensland, School of Civil Engineering, Campus St Lucia, Ed. 49, 4072, QLD, Australia

*³University of Seville, Higher Technical School of Engineering, Department of Chemical and
Environmental Engineering, Camino de los Descubrimientos, s/n, Seville, Spain*

1 ABSTRACT

2 In a society where the environmental conscience is gaining attention, it is necessary to
3 evaluate the potential valorisation options for agricultural biomass to create a change in the
4 perception of the waste agricultural biomass from waste to resource. In that sense, the
5 biorefinery approach has been proposed as the roadway to increase profit of the agricultural
6 sector and, at the same time, ensure the environmental sustainability. The biorefinery approach
7 integrates biomass conversion processes to produce fuels, power, and chemicals from biomass.
8 The present review is focused on the extraction of added-value compounds, anaerobic digestion
9 and composting of agricultural waste as biorefinery approach. This biorefinery approach is,
10 nevertheless, seen as a less innovative configuration compared to other biorefinery
11 configurations as bioethanol production or white biotechnology. However, any of these

12 processes has been widely proposed as a single operation unit for agricultural waste valorization
13 and a thoughtful review on possible single or joint application has not been available on
14 literature up to know. The aim of this manuscript is to review the previous and current literature
15 about the potential valorisation of agricultural waste biomass, focusing on valuable compounds
16 extraction, anaerobic digestion and composting of agricultural waste, whether they are none,
17 partially or fully integrated.

18

19 **1. INTRODUCTION**

20 Agricultural sector and agri-food industry are high volume generating sectors of organic
21 waste, reaching up to 90 million tons per year in 2014 in the EU28 countries (Eurostat, 2017).
22 Among these waste, 46 million tons correspond to lignocellulosic waste (Gil et al., 2015).
23 Moreover, food losses and waste from agri-food sector are estimated at around 30% of global
24 food production, therefore an adequate waste management is crucial in the context of global
25 sustainability (European-Commission, 2015). The uncontrolled decomposition of organic waste
26 can result in large-scale contamination of soil, water, and air. In fact, decomposition of one
27 metric ton of organic solid waste can result in the release into the atmosphere of 50–110 m³ of
28 carbon dioxide and 90–140 m³ of methane (Macias-Corral et al., 2008).

29 For a sustainable agricultural sector, it is necessary moving from the current linear, 'take,
30 make, dispose (waste-creation)' model for resource-consumption, to the systemic, circular
31 alternative of 'reduce, reuse, recycle, regenerate' (Rhodes, 2017). In that sense, the agricultural
32 waste biomass should be considered as a sustainable resource, instead of a waste. To be
33 considered a sustainable resource, it is necessary to develop management methods able to
34 provide economic benefits and, at the same time, to ensure the environmental sustainability of
35 the agricultural sector. In order to integrate the agricultural waste management in a circular
36 system, extraction of high-value compounds, anaerobic digestion and composting is been
37 proposed for a profitable and promising management of agricultural waste. Any of these
38 processes has been widely proposed as single operation unit for agricultural waste valorization

39 and a thorough review on possible joint application has not been available on literature up to
40 know. Extraction of high-value compounds is very interesting due to the recovery of
41 compounds with high economic interest, as well as the partial detoxification of the waste due to
42 the removal of some compounds, which could be undesirable for subsequent biological post-
43 treatments (Martín et al., 2010; Negro et al., 2017; Serrano et al., 2017a). However, recovery by
44 extraction of the added value compounds from agricultural waste usually requires a previous
45 separation of a liquid phase through carrying out a pretreatment process, with the consequent
46 energy consumption (Rubio-Senent et al., 2013b). Anaerobic digestion presents the main
47 advantage of generating methane from the organic waste, which could be used as an energy
48 source due to its high calorific power ($35,793 \text{ kJ/m}^3$, at 1 atm, 0 °C) (Wheatley, 1990).
49 Moreover, anaerobic digestion also allows the partial stabilization of the treated waste, although
50 a post-treatment of the final effluent could be required for its use as an organic amendment
51 (Bustamante et al., 2013). However, different compounds present into the organic waste could
52 inhibit the process. The necessity of high investment costs has also been defined as an important
53 limiting factor for the full-scale implementation of anaerobic digestion processes (Swindal et
54 al., 2010). Composting has been proposed for a long time as a cheap option for agricultural
55 waste management (Ernst, 1990; Gutiérrez et al., 2017). During composting, the organic matter
56 is turned into stabilized humic substances through mineralization and humification. Moreover,
57 pathogens are removed by the heat generated in the thermophilic phase (Dadhich et al., 2012).
58 Unfortunately, although it is a sustainable way to returning the nutrients to the agricultural
59 sector, the economic benefit derived from the compost is relatively low, i.e. 0-9 € per t (Evans
60 and Wilkie, 2010).

61 Each of the described processes presents high advantages, but also some disadvantages which
62 could limit its implementation. A biorefinery approach combining-value compounds extraction,
63 anaerobic digestion and composting could be an attractive biorefinery approach for agricultural
64 waste. In the extraction step, a high economic benefit could be obtained through the selling of
65 the recovered compounds. The energy required for the whole biorefinery would be obtained

66 from the anaerobic digestion in the second step. Finally, the composting process could allow the
67 total stabilization of the organic matter and the nutrient recovery for the agricultural sector.

68 The aim of this manuscript is to review the previous and current literature about the potential
69 of different agricultural waste biomasses in a sustainable system based on added-value
70 compounds extraction, anaerobic digestion and composting of agricultural by-products, whether
71 they are none, partially or fully integrated.

72

73 **2. ADDED-VALUE COMPOUNDS FROM AGRICULTURE BY-PRODUCTS**

74 Recovery of added-value compounds from agricultural waste biomass represents a market
75 opportunity, with multiple applications in pharmacy, cosmetics and food industry. The interest
76 of the different added-value compounds derives from their multiple health benefits, including
77 antioxidant, cardiovascular, antihypertensive, and antiproliferative effects. The economic
78 interest of recovering added-value compounds from agricultural biomass will vary in
79 accordance with the kind of agricultural waste, generation volume, the concentration of
80 desirable compounds, etc. An overview of the potential opportunities for each kind of
81 agricultural waste is reviewed in the present manuscript. The different agriculture wastes or by-
82 products have been clustered in three main groups with similar characteristics, i.e.
83 lignocellulosic agricultural by-products, vegetable by-products, and fruit by-products.
84 Most of reviewed by-products are coming from some of the main crops produced worldwide.
85 The present review is also giving a general idea of possible biorefinery application of any other
86 by-product not reviewed but included in the any of the three clusters.

87 **2.1. Lignocellulosic agricultural by-products**

88 Examples of lignocellulosic agricultural by-products are wheat straw, rice straw, maize
89 stalks, corn straw, ensiled sorghum forage, barley, sunflower or cotton gin waste. These
90 lignocellulosic agricultural by-products contain variable amounts of cellulose, hemicellulose,
91 lignin and small amounts of protein, pectin, wax, and inorganic compounds. The use of
92 lignocellulosic agricultural by-products has been gaining interest. From lignocellulosic

93 agricultural by-products, several products as biofuels, chemicals and other biomass-derived with
94 high added value product can be obtained through the integration of clean processes (Abraham
95 et al., 2016).

96 Depending on the refinery system, these feedstocks could undergo various chemical or
97 mechanical pretreatment in order to facilitate valuable products extraction. Among the different
98 pretreatments, autohydrolysis or hydrothermal process is an interesting and well-established
99 eco-friendly process in which lignocellulosic material is pretreated with high-pressure steam.
100 After the hydrothermal pretreatment, a liquid phase and a solid phase are generated. Through
101 the hydrothermal process, hemicellulose is extracted into the liquid phase. The acidic groups
102 bonded to the hemicellulose are released at high temperature. These acids, mainly acetic acid
103 and hydronium ions coming from water auto-ionization, enhance the hydrolysis of the solid
104 lignocellulosic material that leads to the further solubilization of hemicellulose. This process
105 also facilitates hydrolysis of cellulose. The obtained liquid phase is rich in products of
106 degradation of cellulose, hemicellulose, and lignin. The introduction of a detoxification step or
107 removal of these dissolved compounds is highly recommended if a subsequent bioprocess wants
108 to be implemented (Fernández-Bolaños et al., 1998; Yu and Christopher, 2017). The liquid
109 phase (hydrolysate) contains monomeric and oligomeric hemicellulose sugars, sugar
110 degradation products, acetic acid, extractives and phenolics compounds derived from the acid-
111 soluble lignin. These last compounds can be used for the production of compounds of interest
112 for the health, cosmetic and food industries (Moure et al., 2006). The antioxidant and
113 antimicrobial potential of depolymerized lignin fraction produced by mild acid hydrolysis of
114 lignocellulosic material has been reported (Cruz et al., 2007). This type of compounds that
115 inhibit growth and metabolism of microorganism can be selectively removed by solvent
116 extraction (Fernández-Bolaños et al., 1998). The utilization of crude extracts as antioxidants,
117 instead of a pure compound or purified fraction, is a frequent approach. It is an alternative more
118 favorable from an economic point of view and, even in some case, the crude extract has

119 presented higher antioxidant capacity than the mixture of main component (Rubio-Senent et al.,
120 2014).

121 The hemicellulose fraction can be used as hydrogel being considered as an alternative for the
122 development of new polymeric blends for food packaging (Ruiz et al., 2013). The oligomeric
123 form can be used as functional food ingredients (Mäkeläinen et al., 2010) and the monomeric
124 form can be fermented to ethanol or xylitol (Avanthi et al., 2017). Xylooligomers with prebiotic
125 activity in addition of ingredients for food by stimulating the growth of bacteria in the colon and
126 improving the host's health, are also currently considered interesting in cosmetic for skin care,
127 in pharmaceutical industries by its multiple biological properties (anticarcinogenic,
128 immunomodulatory, antimicrobial activity or by their ability to decrease glucose and
129 cholesterol) and in agriculture as ripening agents (Álvarez et al., 2017).

130 The remaining solid phase is enriched in both cellulose and lignin (Rodríguez-Gutiérrez et
131 al., 2008; Rodríguez-Gutiérrez et al., 2014; Ruiz et al., 2013). Lignin is an amorphous
132 polyphenol with high molecular weight, which can vary in structure according to the extraction
133 method and plant source (Fernández-Bolaños et al., 1999; Rodríguez-Gutiérrez et al., 2014).
134 Nowadays, the major lignin use is as fuel, and only a few products like vanillin are produced
135 (Singh and Ghatak, 2017). However, lignin is becoming a high interesting product. For example
136 given to its radical scavenging properties for the formulation of plastic and cosmetic (Guilhen et
137 al., 2017; Morganti, 2016). Production of low-molecular weight compounds with potential uses
138 as surfactants or additive for liquid fuels can be obtained after lignin fractionation with alcohols
139 (Cabrera et al., 2016). Furthermore, lignin represents a removable low-cost alternative to natural
140 antioxidants and could be used as a component of polymer composites or of the polysaccharides
141 hydrogel films as antioxidant carriers or active packaging (Aguíé-Béghin et al., 2015). Lignin is
142 an important component of dietary fiber, being nontoxic and biocompatible, which undergoes
143 minimal changes in the body (non-fermentable). Lignin has been proven to bind various bile
144 acids and detoxify harmful metabolites inhibiting colonic carcinogenesis (Camire and
145 Dougherty, 2003). Concretely, a fraction containing lignin from olive stones bound significantly

146 more bile acid than any other fraction and an amount similar to that bound by cholestyramine (a
147 cholesterol-lowering, bile acid-binding drug). Therefore, this lignin fraction from olive stone
148 could contribute to the reduction of serum cholesterol levels and a decreased risk of bowel
149 cancer (Rodríguez-Gutiérrez et al., 2014).

150 Although all agricultural by-products have approximately the same composition, the yield in
151 hemicellulosic sugar or the development of valuable-added products differs, when applying the
152 same conditions or type of pretreatment. It is thus important to study each raw material to
153 determine the appropriate production conditions. For example, in the case of maize bran, which
154 is a rich source of dietary fiber and phenolic antioxidants, the hemicellulose fraction is a
155 complex heteroxylan consisting mainly of xylan backbone with arabinosyl side. The production
156 of feruloylated arabinoxylan-oligosaccharides from maize bran by autohydrolysis may provide
157 health benefits, including prebiotic effect and prevention of detrimental oxidation reaction (Rose
158 and Inglett, 2010).

159 **2.2. Vegetable by-products**

160 In general, by-products from handily and commercialization of vegetables such as tomato,
161 onion, potato or carrot, have been traditionally used as animal feedstuffs, for dietary fiber
162 production and fuel production. Most of the vegetable by-products contain the same bioactive
163 compounds than the vegetable itself (Table 1). Due to this, most of the vegetables by-products
164 have been shown to present similar health benefits than the vegetable (Table 2). Therefore, an
165 interesting approach is their use as source of phytochemicals and bioactive compounds.

166 2.2.1. Tomato by-products

167 The by-products resulting from tomato processing are mainly peel and seeds (Gharbi et al.,
168 2017). Tomato seed oil has attracted interest by its high content of unsaturated fatty acid with
169 over 50% linoleic acid (da Silva and Jorge, 2017). Tomatoes have been associated with reduced
170 risk of some types of cancer and other diseases. These beneficial effects have been linked to the
171 content of lycopene and other carotenoids as β -carotene and lutein (Fattore et al., 2016). Other
172 bioactive compounds, particularly polyphenols, has been shown to contribute to the antioxidant

173 effect. The major phenol compounds of tomatoes are the flavanones, naringenin glycosylated
174 derivatives and flavanols, quercetin, rutin and kaempferol glycoside derivatives (Kelebek et al.,
175 2017). Gharbi et al. (2017) showed that the antioxidant composition of tomato peels and tomato
176 seeds Lycopene is mostly associated with the water insoluble fraction of the peel. Furthermore,
177 tomato seed has been shown to be a source of pectin (Morales-Contreras et al., 2017) and
178 protein of high-quality (Moayedi et al., 2016). At the industrial scale, the tomato waste is been
179 using for animal feed or fertilizer, however, new techniques are been studied in order to extract
180 pectin and/or bioactive components. These techniques include solvent extraction by stirring and
181 heating, heat refluxing extraction, microwave or ultrasonic in combination with subcritical
182 water (Grassino et al., 2016).

183 2.2.2. Onion by-products

184 The major by-product resulting from industrial peeling of onion bulbs is the brown skin, the
185 outer two fleshy leaves and the top and bottom bulbs. Onions, as well as its by-products, are an
186 important source of several phytonutrients as flavonoids, fructooligosaccharides (FOS) and
187 thiosulfonates and other sulfur compounds (Liguori et al., 2017). Quercetin and kaempferol
188 glycosides are the predominant polyphenols found in onions. They are present in higher
189 concentration (280-400 mg/kg) than in other vegetables (Sharma et al., 2015). Anthocyanins are
190 also present in red onions (Wiltshire et al., 2017). Onion by-products are also a source of inulin,
191 a polysaccharide known as fructan, which is used as a dietary fiber with prebiotic effect (Smith
192 et al., 2015) or as an energy source (Hughes et al., 2017). The onion waste is not suitable for
193 food animal or for organic fertilizer, the only option is landfill, with high economic costs, and
194 adverse environmental impact. The pretreatment methods studied for its valorization are based
195 on technologies like organic extraction, supercritical carbon dioxide, supercritical water
196 treatment, microwave, microwave assisted, hydro diffusion and gravity or high-pressure
197 processing (Sharma et al., 2016).

198 2.2.3. Carrot by-products

199 Carrot is a rather inexpensive and highly nutrition vegetable. It contains natural antioxidants,
200 including phenolic compounds and carotenoids. Carrot pomace, a major by-product of carrot
201 juice processing, represents a rich source of bioactive compounds with antioxidants activities
202 (Jabbar et al., 2015). Carrot pomace represents a valuable natural source of α - and β -carotene,
203 with a total carotene content up to 2 g per kg dry matter. The pomace has been used for the
204 production of antioxidant dietary fiber powder and for the extraction of pectin (Jabbar et al.,
205 2015; Jafari et al., 2017).

206 The phytochemical profile of carrot by-product is composed of hydroxycinnamic acid
207 derivatives, particularly four chlorogenic acids (5-O-caffeoylquinic) and six derivatives
208 dicaffeoylquinic acids (Sánchez-Rangel et al., 2016) with many biological functions including
209 antioxidant, antiviral, antimutagenic, anti-inflammatory, cardioprotective, antiobesity and
210 therapy on wound healing (Akhtar et al., 2017; Bagdas et al., 2014).

211 After the carrot juice processing, around the 50 % of the raw material is discarded as a waste
212 or use for feed animal purpose. Several pretreatments have been studied to extract the bioactive
213 compounds like refluxing, boiling and heating, but some of the more promising technique is
214 based on the ultrasound-assisted extraction up to 60 °C (Jabbar et al., 2015), or the aqueous two-
215 phase system extraction (Sánchez-Rangel et al., 2016).

216 2.2.4. Potato by-products

217 Potato peels, a by-product of potato processing, are available in large amounts, and, since
218 peels have much more phenolic compounds compared to tubers, these phenolic compounds
219 have a potential application in food and non-food applications. Aqueous peel extracts were
220 shown to be used as an antioxidant on different oils (Amado et al., 2014), in minced horse
221 mackerel (Sabeena Farvin et al., 2012) or in processed lamb meat (Kanatt et al., 2005). Also,
222 this extract has been used as a source of phenolic compounds. Potato peels have relatively high
223 content in phenolic acids, especially chlorogenic acid. In addition, the potato peels are also a
224 source of water-soluble polysaccharides (Jeddou et al., 2016). Beneficial anticarcinogenic
225 properties of potato glycoalkaloids have been recently reported (Friedman et al., 2017).

226 Nevertheless, potato glycoalkaloids at high concentration might be potentially harmful to
227 human. Recently, the potato peel by-products have also been used as an effective biosorbent for
228 removal of toxic metal agents from water (Azmat et al., 2016). Commonly used techniques to
229 obtain antioxidant extracts from potato by-products are organic solvents, like ethanol, or
230 pressurized liquid extraction, microwave-assisted extraction, and subcritical water extraction or
231 even thermal treatment at 121 °C which allows the application of the further biorefinery
232 approach (Pathak et al., 2017).

233 **2.3. Fruit by-products**

234 Similar to vegetable by-products, most of the fruit by-products contain the same bioactive
235 compounds than the fruit itself (Table 1). Due to this, most of the fruit by-products have been
236 shown to present similar health benefits than the fruit (Table 2).

237 2.3.1. Grape by-products

238 The main by-product of the wine industry is known as grape pomace and consists mainly of
239 skin, seeds, stems and remaining pulp. Its composition varies considerably depending on grape
240 variety and technology of wine making. The moisture percentage varies from 50-72%. The
241 insoluble residues from this material have a lignin content ranging from 17-24%, cellulose
242 varying from 27-37% and protein content is lower than 4% (Teixeira et al., 2014). Grape
243 pomace has been used as soil conditioner, as a source of fibers and energy by methanization
244 (Barba et al., 2016). Also it has been used for seed oil extraction, as a source of tannins (Bindon
245 et al., 2017) and as a source of protein for animal feed (Brenes et al., 2016). During the
246 winemaking process part of the phenolic compounds in grape are transferred to the wine, but a
247 high proportion still remains in the waste, especially in the grape pomace (Ribeiro et al., 2015).
248 Numerous studies have demonstrated that these phenolic compounds exhibit health-promoting
249 effect ascribed with cardioprotective, neuroprotective, anti-inflammatory, anticarcinogenic and
250 other health benefits (Panzella and Napolitano, 2017) and are active against pathogenic bacteria,
251 virus, and fungi (Friedman, 2014).

252 Grape contains a large amount of different phenolic compounds distributed in pulp (10%), in
253 seeds (60-70%) and in the skin (28-35%). The most predominant polyphenols found in grape
254 pomace are a) phenolic acids- caffeic, gallic, protocatechuic, 4-hydroxybenzoic and syringic
255 acid, b) phenolic alcohols- hydroxytyrosol, c) flavonoids- (+)-catechin, catechin dimer, (-)-
256 epicatechin, epicatechin gallate trimer, procyanidin B1 and B2, quercetin-3-O-rhamnoside,
257 luteolin, d) stilbenes, with the presence of trans-resveratrol e) proanthocyanidins, also known as
258 condensed tannins, complex phenols of high molecular weight and anthocyanins (Teixeira et
259 al., 2014). These grape pomaces are currently used as a source of resveratrol and flavonoids,
260 which are used as food supplements and for the isolation of anthocyanins for use as a substitute
261 of synthetic colorants and for preparation of dietary fiber and polyphenol-rich extracts (Zhang et
262 al., 2017). Moreover, grape seeds are a rich source of other antioxidant compounds such as
263 vitamin E, which constitutes a family of lipid-soluble antioxidant compounds, containing a
264 saturated (tocopherol) or unsaturated (tocotrienols) isoprenoid side chain (Barba et al., 2016).
265 Grape seeds are also a good source of polyunsaturated fatty acids (PUFA), mainly linoleic acid,
266 followed of α -linolenic and oleic acid (Ribeiro et al., 2015).

267 Besides the conventional methods based on the heating process and/or solid liquid extraction
268 industrially used for antioxidant extracts, new techniques are been studied to improve the
269 extraction and to diminish the thermal degradation of phenols, like pulsed electric fields, high
270 voltage electrical discharges, pulsed ohmic heating, ultrasounds, microwave assisted, sub- and
271 supercritical fluid extraction, high pressure, accelerated solvent or extraction assisted by
272 hydrotropic solvents (Barba et al., 2016). The extraction method commonly used to produce
273 commercial extracts is done by hydroalcoholic solvents, mainly ethanol (Teixeira et al., 2014).

274 2.3.2. Orange and lemon by-products

275 The residue from orange and lemon juice extractions industries are a good source of bioactive
276 ingredients such as essential oil, which consist mainly in monoterpene (limonene) and
277 triterpenoids as limonoids. Other bioactive compounds are dietary fiber (DF); pectin; ascorbic
278 acid; phenols (coumaric, caffeic and ferulic acids); and flavonoids, mainly flavanones

279 glycosides (hesperidin, naringin and narirestin), flavones (hesperetin, naringenin), flavones
280 aglycon (luteolin) and polymethoxylated flavones (tangeretin) ((M'hiri et al., 2017). Flavonoids
281 concentration in citrus peels is higher than in juice and seeds (Tao et al., 2014). These
282 compounds inhibit the cell growth of a large group of microorganisms and may be useful as
283 antiviral, antifungal and antibacterial agent (Damian-Reyna et al., 2016). In recent years has
284 been accumulating evidence of the cancer-preventive effect of limonene (Mitropoulou et al.,
285 2017). Also, D-limonene has been clinically used to dissolve cholesterol-containing gallstones.
286 Because of its gastric acid neutralizing effect and its support of normal peristalsis, it has also
287 been used for relief of heartburn (Sun, 2007). Another important compound recovered from the
288 peel is pectin, which is usually extracted with hot dilute acid and used as thickening agent,
289 gelling agent, and stabilizer in the food industry (John et al., 2017). Citrus limonoids (CLs) are a
290 group of highly oxygenated terpenoid secondary metabolites found mostly in the seeds, fruits
291 and peel tissues of citrus fruits. Represented by limonin, the aglycones and glycosides of CLs
292 have shown to display numerous pharmacological activities including anticancer, antimicrobial,
293 antioxidant, antidiabetic and insecticidal among others (Gualdani et al., 2016).

294 2.3.3 Apple by-products

295 The potential use of apple by-products to isolate specific phytochemicals for application in
296 food or dietary supplements contributes to the recovery of these by-products. The by-products
297 resulting from processing of apple fruit represent approximately 25-30 % of fruit weight
298 (Schieber et al., 2003) and are made of peels, seeds and flesh. Although, normally is used as an
299 animal feed or wasted, the apple pomace is utilized for the recovery of valuable compounds,
300 such as dietary fiber and polyphenols (Sudhaa et al., 2007; Parra et al., 2015). The production of
301 pectin from apple, 10-15% of apple pomace, on a dry weight basis, have been established even
302 an industrial scale. Apple pomace is also consider a good source of natural antioxidants with
303 important properties that include antimicrobial, anticancer and cardiovascular-protective
304 activities (Eberhardt et al., 2000) (Sesso et al., 2003). The major antioxidants isolated and
305 identified include the flavanols quercetin glycosides, kaempherol, catechin, and procyanidins

306 (Kammerer et al., 2011). A method for the combined recovery of pectin and polyphenol was
307 established (Schieber et al., 2003). Apple seeds has been used for the recovery of oil and
308 bioactive compounds. The oil extractable from seed of apple stood out for its high content in α -
309 and γ -tocopherol and β -sitosterol (Da Silva & Jorge, 2017) and is interesting for their use in
310 cosmetic, food and pharmaceutical application (Walia et al., 2014).

311 2.3.4. Mango by-products

312 Mango processing by-products (peel, kernel, and seed) comprise 35-60% of the total fruit
313 weight thus representing a potentially high volume resource of exploitable biobased chemicals
314 and material within the context of biorefinery (Matharu et al., 2016).

315 The analysis of the oil extracted from fruit seeds verifies the presence of bioactive
316 compounds, such as phytosterols, as well as phenolic compounds (gallic acid, salicylic acid,
317 epicatechin, and quercetin) and tocopherols which present antioxidant capacity (da Silva and
318 Jorge, 2017). The kernel contains high amount of fiber, starch and hydrolysable tannins, which
319 possess antimicrobial activity (Jahurul et al., 2015). The mango peels have high amount of
320 extractable polyphenols, hydrolyzable tannins, flavonoids, xanthonenes and anthocyanins (Dorta
321 et al., 2014) and a high antioxidant activity (Sabino et al., 2015). Mango by-product is also a
322 rich source of highly esterified pectin (Matharu et al., 2016). For the industrial valorization of
323 mango by-products it is necessary to use heating but in combination with additional techniques
324 to reduce the volume of effluent generated, like microwaves, pulsed electric energy or
325 ultrasound (Matharu et al., 2016).

326 2.3.5. Papaya by-products

327 The papaya peel flour has a high amount of ascorbic acid and lycopene (Sabino et al., 2015).
328 The carotenoid content, including lutein, cryptoxanthin, and β -carotene, are also significant in
329 papaya by-product (de Moraes Crizel et al., 2016). It was recently found the presence of
330 terpenoids in papaya extracts with important antibacterial activity (Lawrence et al., 2015). The
331 seed oil is rich in oleic acid, α -tocopherol, carotenoids and a significant source of phytochemical
332 (da Silva and Jorge, 2017). Oil is obtained from the papaya seed by ultrasound-assisted

333 extraction, extrusion-expelling processes and solvent and aqueous enzymatic extraction (Cheok
334 (Cheok et al., 2017) et al., 2016). Papain, a proteolytic enzyme used in food industries as a meat
335 tenderizer, for stabilizing and chill-proofing beer and in baking processes, is recovered from the
336 latex of papaya fruit (Rocha et al., 2016).

337 Conventional treatment based on grinding and heating is commonly used to revalorized the
338 papaya by-products, but new technologies are been also studied to improve the extractions,
339 diminishing the time and the degradation products, like Pulsed Electric Fields (Parniakov et al.,
340 2016).

341 2.3.6. Banana by-products

342 Banana peel is a waste produced in large volume annually by food-processing industries. For
343 1 ton bone-dry banana peel could be obtained 430 kg of protein or 170 kg of citric acid, 170 kg
344 of pectin, 325 m³ of ethanol, and 220 m³ of methane (Pathak et al., 2017). Banana by-products
345 are also a good source of bioactive compounds with high added value (Singh et al., 2016).
346 Banana by-products have a high phytochemicals concentration that consist mainly of phenolic
347 compounds, flavonoids, proanthocyanidins (Vu et al., 2016), carotenoids (α - and β -carotene
348 and xanthophyll) (Davey et al., 2009), sterols and triterpenes (Hernández-Carranza et al., 2016)
349 and saponin (Siddique et al., 2017). Banana peel contained large amounts of dopamine and L-
350 dopa, catecholamines with a significant antioxidant activity (González-Montelongo et al.,
351 2010). For the extraction of the bioactive compounds, techniques like the use of solid-liquid
352 with hot organic solvents, alkaline, enzyme-assisted, ultrasound, and supercritical fluid
353 extraction, have been studied (Hernández-Carranza et al., 2016).

354 2.3.7. Pineapple by-product

355 Pineapple processing generates large quantities of solid and liquid waste. The peel represents
356 the largest portion (30-42% w/w), followed by the core (9-10%) and stem (25%). These by-
357 products could be a potential source of sucrose, glucose, fructose, saccharose, fiber, bromelain
358 and phenolic compounds (Dorta and Sogi, 2016).

359 Bromelain, a protease with a wide range of industrial application in food, beverage and
360 cosmetic, may be recovered from pineapple by-products including stem, fruit, leaves, and peel.
361 The high demand for bromelain has led to the need for a high purified bromelain production at
362 low cost (Ramli et al., 2017) like membrane systems (Nor et al., 2017) or aqueous two-phase
363 system (Upadhyay et al., 2010).

364 Pineapple by-products may be utilized for the recovery of phenolic compounds. Gallic
365 acid, catechin, epicatechin and ferulic acid were found to be the main polyphenolic in pineapple
366 peels (Li et al., 2014). Also, a high content of anthocyanin, α and β -carotene were also found
367 (Da Silva et al., 2014). The most common system used to extract them are based on solid-liquid
368 extractions, using water or organic solvents (Upadhyay et al., 2010).

369 Pineapple pomace containing high amount of dietary fiber was used for fortificant extruded
370 products (Selani et al., 2014). Peel flour, which presents a good prebiotic potential, could be
371 used to support probiotic bacteria in the gut (Dorta and Sogi, 2016).

372 2.3.8. Olive oil by-products

373 The olive oil industry generates by-products from the three olive oil extraction system, the
374 liquid (alpechin) and the solid (orujo) with a 45-55% of humidity, and from the two olive oil
375 extraction system, the solid with high humidity (70-80%) called alperujo. The presence of
376 bioactives compounds in all these by-products are similar, being concentrated in the solid
377 phases, hence it is necessary to apply a pretreatment to extract them and to facilitate the
378 solid/liquid separation. Despite many studies have been carried out to treat the alperujo and the
379 orujo (chemicals, physicals, biological or mix systems) only the thermal treatments are been
380 industrially used (Fernández-Bolaños et al., 2002). Two kinds of thermal treatments are been
381 used, a malaxation at 55-65 °C and a steam treatment at 150-170 °C. The effects of the thermal
382 treatment are based on the solubilization of sugars and phenols, the easier separation of the
383 phases, the final solid with lower humidity (30-55%) and the oil richer in minor components
384 and concentrated two or three times in the final solid (Lama-Muñoz et al., 2011).

385 The bioactive compounds present in the olive oil industry are mainly the phenols and sugars.
386 Hydroxytyrosol is one of the most actives and remarkable phenol in the olive because of its high
387 anti-oxidant, anti-inflammatory and anti-platelet potency in humans (Fernández-Bolaños et al.,
388 2008). In addition to hydroxytyrosol, there are other interesting phenols in the by-products such
389 as 3,4-dihydroxyphenylglycol, triterpenic acids, lignans and secoiridoids, all of them with
390 potential anti-inflammatory and anti-oxidant activity (Rubio-Senent et al., 2012). The
391 application of solid/liquid extraction with an organic solvent or thermal pretreatment following
392 by chromatographic or liquid/liquid extraction leads to produce a commercial extract rich in
393 phenols being the hydroxytyrosol the majority (Rubio-Senent et al., 2013b). Beside the phenols,
394 the carbohydrate fraction has been shown biological activities, like oligosaccharides with low or
395 high molecular weight, pectins or phenolic glucosides, with antioxidant, antiproliferative or
396 probiotic activities (Rubio-Senent et al., 2015b).

397

398 **3. ANAEROBIC DIGESTION**

399 Anaerobic digestion of agricultural waste has been widely reported in literature.
400 Biomethanization of agricultural waste not only delivers directly usable energy, as
401 biogas but also retains the nutrients contained in the biomass (Braun, 2007). The
402 obtained energy could be used for the energy demands of the extraction of added-value
403 compounds in a biorefinery system, but also to supply to other energy requirements
404 such as water pumping. The three groups defined in the previous section has been
405 followed. The main considerations for the biomethanization of each group are described
406 below.

407 **3.3. Lignocellulosic agricultural by-products**

408 Anaerobic digestion of lignocellulosic by-products is usually characterized by a low
409 biodegradability, especially due to the lignin content of this kind of by-products. For
410 example, winter harvested switchgrass showed a much lower biodegradability than the

411 fresh summer harvested switchgrass, probably due to a much lower lignin fraction than
412 the winter harvested switchgrass (Frigon et al., 2012; Table 3). In order to enhance
413 biodegradability, trials applying strong pretreatments such as steam explosion and/or
414 alkaline dosage are usually done in order to mainly breakdown the lignin fraction (Table
415 3) (Bauer et al., 2009; Estevez et al., 2012; Frigon et al., 2012; Monlau et al., 2012;
416 Rodriguez et al., 2017; Sambusiti et al., 2013; Xie et al., 2011). Other authors have
417 proposed the addition of acid compounds instead (Hassan et al., 2016; Monlau et al.,
418 2012). The selection of a pretreatment must be in accordance with the composition of
419 the agricultural waste to be treated. Acidic pretreatments removed more than 90% of
420 hemicelluloses and uronic acids whereas alkaline and oxidative pretreatments were
421 more effective in dissolving lignin (Monlau et al., 2012). For example, Xie et al. (2011)
422 reported that a thermal pretreatment (100°C) with NaOH addition allowed the
423 solubilization of 65.6% of lignin, 36.1% of hemicellulose and 21.2% of cellulose.

424 As an alternative to the chemical addition in thermal pretreatments, steam explosion
425 could pretreat the lignocellulosic waste without the addition of chemicals and with
426 minimal sample handling (Estevez et al., 2012). Steam explosion could increase the
427 methane yields up to 50%, (Monlau et al., 2012). Hemicelluloses solubilization
428 debilitates the lignocellulosic structure of the substrate and facilitates thus, its
429 microbiological degradation. However, the overproduced methane due to the
430 pretreatment is not necessarily enough to cover the energy requirements of the applied
431 pretreatment (Serrano et al., 2017b).

432 It is not so unusual to observe a decrease of methane production after using a high-
433 temperature pretreatment. The decrease of methane yield when high temperatures are
434 applied is usually attributed to the formation of substances inhibiting the
435 microorganisms responsible for the anaerobic digestion process (e.g. phenolic

436 compounds or furan derivatives) as well as to the loss of sugars due to pseudo-lignin
437 formation (Ghasimi et al., 2016; Monlau et al., 2012; Rodriguez et al., 2017). The
438 inclusion of an extraction step of added values after the pretreatment and before the
439 digester reduce this drawback (Serrano et al., 2017a). During the extraction step, the
440 concentration of some inhibitors can be reduced in the anaerobic influent, whereas the
441 benefit derived from the recovery of high added-value compounds could compensate
442 the extra energy requirements of the pretreatments.

443 **3.4. Vegetable by-products**

444 Anaerobic digestion of vegetable by-products is mainly characterized by high fiber
445 content, mainly hemicellulose and, to a lesser extent, cellulose (Ji et al., 2017).
446 Vegetable by-products present higher methane yields and biodegradability values than
447 the obtained for lignocellulosic by-products, probably due to a lower content of lignin
448 (Table 3 and Table 4). Reported values of methane yield for vegetable by-products can
449 reach really high values, e.g. 390 mL CH₄/g VS (Volatile Solids) for onion waste (Ji et
450 al., 2017; Menardo and Balsari, 2012), 320 mL CH₄/g VS for potato waste (Parawira et
451 al., 2004), or 198 mL CH₄/g VS for carrot waste pomace (Garcia et al., 2011).
452 Moreover, biodegradability values were also higher than the obtained for lignocellulosic
453 by-products (Table 3 and Table 4). Typical biodegradability values for vegetable by-
454 products have been reported in a range from 50 to 70%, in VS (Garcia et al., 2011; Ji et
455 al., 2017; Menardo and Balsari, 2012). The high methane yield and biodegradability
456 values advise against the implementation of pretreatments methods to enhance the
457 biomethanization step, although it would be necessary for some previous extraction
458 procedures.

459 Unfortunately, some drawbacks have been reported for the anaerobic digestion of
460 vegetable by-products. Several vegetable wastes have rather low pH values, e.g. a pH of

461 3.2 for onion, a pH of 5.6 for Lettuce, or a pH of 5.06 for pepper (Ji et al., 2017;
462 Menardo and Balsari, 2012), which can cause the acidification of the anaerobic process.
463 In that sense, Lubberding et al. (1988) described the accumulation of butyric and valeric
464 acid during the biomethanization of onion waste. Moreover, although biodegradability
465 during the biomethanization is relatively high, a post-treatment would be desirable for
466 the total stabilization of the organic matter. The employment of alkaline pretreatments
467 in the extraction process could compensate the low initial pH of the vegetable waste,
468 whereas the subsequent composting process could mineralize the anaerobic effluent.

469 **3.5. Fruit by-products**

470 Anaerobic digestion of fruit by-products is mainly characterized by high sugar and
471 fiber content, mainly hemicellulose (Bouallagui et al., 2005; Ji et al., 2017). As in
472 vegetable by-products, lignin fraction is lower than the reported for lignocellulosic by-
473 products. However, the presence of husks in some fruit by-products, such as olive husks
474 or strawberry achenes, could entail significant lignin concentrations (Serrano et al.,
475 2017a; Serrano et al., 2017b; Siles et al., 2013). Reported values of methane yield for
476 fruit by-products widely vary according to the treated fruit, or if the biomass is
477 composed of the whole fruit or just the peel of the fruit (Table 5). The highest reported
478 methane yields were reached for pineapple waste, i.e. 413 mL CH₄/g VS, kiwi waste,
479 i.e. 371 mL CH₄/g VS (Menardo and Balsari, 2012), and two-phase Olive Mill Solid
480 Waste (OMSW), i.e. 373 mL CH₄/g VS (Rincón et al., 2013). However, significant
481 lower methane yield values were reported for fluted pumpkin peels and banana peels,
482 i.e. 164 mL CH₄/g VS and 188 mL CH₄/g VS, respectively (Bardiya et al., 1996;
483 Dahunsi et al., 2016) (Table 5). The marked difference in the methane yield could be
484 due to the carbohydrates percentage into the waste and/or the different fiber
485 composition (Bardiya et al., 1996; Ji et al., 2017). Biodegradability values of fruit by-

486 products could reach values higher than the reported for vegetable and lignocellulosic
487 by-products. In fact, biodegradability values up to 90%, in VS, for orange peel (Siles et
488 al., 2016), 90%, in VS, for strawberry extrudate (Siles et al., 2013), or 73.8%, in VS, for
489 pineapple pulp (Namsree et al., 2012) (Table 5) have been reported.

490 In order to improve the biomethanization of fruit by-products, several kinds of
491 pretreatments have been proposed. Pretreatments focus on the removal or reduction of
492 potential inhibitors, this results in an improvement of the process stability and/or the
493 methane yield. For example, removal of D-limonene, which is an economically
494 interesting compound, was reported as necessary to avoid the inhibition in the
495 biomethanization of orange peel (Martín et al., 2010; Siles et al., 2016). The recovery of
496 phenolic compounds also reduces the inhibition risk on the biomethanization of waste
497 such as OMSW (Serrano et al., 2017a). Additionally, other authors have reported
498 improvements in the methane from 230 to 312 mL CH₄/g VS, i.e. 36% higher, in the
499 biomethanization of strawberry extrudate due to the reduction of the lignin content
500 through the removal of the achenes from the strawberry extrudate (Siles et al., 2013).
501 However, the application of thermal pretreatments for increase the solubilization of the
502 waste does not usually result in significant methane yield improvements (Dahunsi et al.,
503 2016; Rincón et al., 2013). The low reported improvements could be a consequence of
504 the low cellulose content of fruit waste, which means that the hydrolytic process during
505 anaerobic digestion is not the rate-limiting step (Ji et al., 2017). Therefore, the energy
506 requirements of some pretreatments only could be compensated if the extraction of high
507 added-value compounds is implemented prior the biomethanization (Serrano et al.,
508 2017b).

509 The anaerobic digestion of fruit by-products could present some problems if the
510 process is not correctly monitored. For example, the low pH values of fruit by-products

511 can result in the acidification of the anaerobic process. Moreover, an additional nitrogen
512 source could be required due to the low nitrogen content of some fruit waste (Belhadj et
513 al., 2014; Gil et al., 2015). To avoid these inconveniences, several authors have
514 proposed the addition of co-substrates to provide buffering capacity, to stabilize the pH,
515 and/or to optimize the nutrient balance (Belhadj et al., 2014; Suryawanshi et al., 2013).

516

517 **4. COMPOSTING OF DIGESTATE**

518 After anaerobic digestion process, a wet residue called digestate is produced. The digestate is
519 a mixture of partially stabilized organic matter, microbial biomass and inorganic compounds
520 (Serrano et al., 2014b). Due to its content on carbon, nitrogen and phosphorus, several authors
521 have proposed that the digestate could be employed as nutrient source for agriculture
522 (Albuquerque, J. A. et al., 2012; Kaparaju et al., 2012; Vaneekhaute et al., 2017). Therefore,
523 the nutrient cycle would be closed by returning the nutrients contained in the agricultural wastes
524 to the soil. Direct application of digestate to the soil has been widely proposed (Albuquerque, J.
525 A. et al., 2012; Gómez-Brandón et al., 2016).

526 However, the application of the digestate without a post-treatment could not be always
527 suitable or safe (Monfet et al., 2017). By one hand, direct application of digestate to soil
528 presents operational difficulties derived from its viscosity, odor, and high humidity, which could
529 complicate its handling (Bustamante et al., 2012; Walker et al., 2009). By another hand,
530 digestate composition includes a high content of potentially phytotoxic compounds such as
531 volatile fatty acids and ammonia (Hanajima et al., 2007; Kaparaju et al., 2012). These
532 compounds can cause harmful effects on seed germination, and plant growth and development
533 (Kaparaju et al., 2012; Tiquia et al., 1996). Other minority compounds of the digestate, such as
534 heavy metals, phenolic compounds or salts can also produce phytotoxicity if good agricultural
535 practices are not ensured (Albuquerque, José Antonio et al., 2012; Bustamante et al., 2012;
536 Kaparaju et al., 2012). In addition, although anaerobic digestion usually reduces the pathogenic
537 load respect the untreated substrates, prions and spore-forming bacteria could be present in the

538 digestate (Gómez-Brandón et al., 2016). So, direct application of digestate to the soil could
539 entail the spread of pathogens in the environment, such as *Salmonella*, *Campylobacter*, *Yersinia*
540 *enterocolitica*, and *Cryptosporidium* (Bustamante et al., 2012; Vaneeckhaute et al., 2017).

541 Due to the potential risk for the agricultural soil derived from the direct application of the
542 digestate, full biological stabilization of the organic matter from digestate can be ensured by
543 aerobic maturation or complete composting (Monfet et al., 2017). During composting, the
544 organic matter is turned into stabilized humic substances through mineralization and
545 humification, resulting in a significant decrease in volume (Gutiérrez et al., 2017). In addition,
546 pathogens are mainly removed by the heat generated in the thermophilic phase (Bustamante et
547 al., 2013; Gutiérrez et al., 2017). Moreover, the compost has been proposed for soil remediation
548 processes such as in situ heavy metal removal, immobilization of pesticides, and removal of
549 emerging pollutants (Cerdeira et al., 2017; Kuppusamy et al., 2017; Zhou et al., 2017). Therefore,
550 composting as final biorefinery step is a win-win strategy for the agricultural sector, which
551 allows the recovery of the nutrients presented in the digestate and to minimize the
552 environmental pollution risk.

553 Composting of digestate has been widely proposed as post-treatment of anaerobic digestion
554 of animal manures and/or sewage sludge (Bustamante et al., 2014; Spencer, 2007; Torres-
555 Climent et al., 2015). However, composting of digestate from agricultural waste has not been
556 extensively reported in literature. One of the main challenges for the application of composting
557 processes to digestate is the high humidity content of the digestate. Due to its low cost, one
558 reported strategy to compensate the humidity content of the digestate is the employment of
559 vegetables as a bulking agent during the composting process (Bustamante et al., 2012;
560 Bustamante et al., 2013). As an example, Bustamante et al. (2012) reported the employment of
561 vine shoot prunings as a bulking agent for the composting of digestate obtained from the
562 anaerobic co-digestion of cattle slurry and silage. According to these authors, the use of vine
563 shoot prunings as bulking agent reduced the temperature of the process, the electrical
564 conductivity, and N losses during the composting process. Other agricultural wastes proposed as

565 bulking agents for composting digestate were wheat straw, exhausted grape marc or pepper
566 plant prunings (Bustamante et al., 2013). Moreover, the use of agricultural waste as a bulking
567 agent could dilute a possible excess of pollutants, such as metals, which could be found in large
568 concentration in digestates not only from agricultural waste but also for sewage sludge or
569 animal slurries (Moral et al., 2008).

570 Other reported strategy to allow the composting of digestate is the separation and further
571 composting of the solid fraction of the digestate (Holm-Nielsen et al., 2009). This separation
572 reduces the humidity content of the substrate to be composted (Bustamante et al., 2012).
573 Moreover, composting only the solid fraction of digestate can improve the quality of the
574 obtained compost, contributing to the elimination of pathogens, and to reduce the odor emission
575 by decreasing the concentration of volatile compounds, which could also reduce the potential
576 phytotoxicity (Smet et al., 1999; Tchobanoglous and Kreith, 2002). Tambone et al. (2015)
577 studied the solid fraction of digestates obtained from anaerobic digestion of pig slurry, energy
578 crops and agro-industrial residues. For the composting of the solid fraction of these digestates,
579 the employment of a lignocellulosic bulking agent did not give remarkably different results at
580 the final product (Tambone et al., 2015).

581 A liquid fraction of digestate is also obtained through the separation processes. Usually,
582 liquid fraction of digestate is rich in nitrogen but poor in phosphorous. Therefore, it could be
583 applied as fertilizer due to its high N/P ratio and relatively high ammonia content might be of
584 particular interest in P saturated areas (Sigurnjak et al., 2017; Vaneckhaute et al., 2013). The
585 fertilizer potential of the liquid fraction of an anaerobic co-digestion plant with an input feed
586 consisting of 30% pig manure, 30% energy maize and 40% organic waste originating from the
587 food industry was studied by Sigurnjak et al. (2017). According to these authors, the
588 employment of the liquid fraction of digestate did not show significant differences in crop yield
589 and soil quality at harvest in comparison to the employment of inorganic fertilizers.
590 Nevertheless, the same authors advised that experiments on longer-term are required to fully

591 evaluate the effects of the continuous application of liquid fraction of digestate on crop growth
592 and soil fertility (Sigurnjak et al., 2017).

593 Other authors proposed the obtaining of magnesium ammonium phosphate hexahydrate, i.e.,
594 struvite, from the liquid fraction of digestate (Akhiar et al., 2017; Estevez et al., 2014). Struvite
595 crystallization can help control the N/P ratio while simultaneously producing a slow release
596 phosphate (Sheets et al., 2015). Unfortunately, struvite crystallization of liquid fraction of
597 digestate from agricultural waste present some disadvantages. By one hand, the dissolved
598 phosphate content in digested agricultural feedstocks may be lower than the requirement for a
599 proper struvite crystallization (Sheets et al., 2015). Therefore, external phosphate may be
600 needed, with the consequent increases costs. By the other hand, the liquid fraction of digestate
601 from agricultural waste usually presents high levels of calcium. The calcium competes with
602 magnesium to form calcium phosphate precipitates, reducing the precipitation of struvite (Martí
603 et al., 2010). The suitability of each strategy for recovering the nutrients for the soil will depend
604 on the actual agricultural system, i.e. soil requirements of nutrients, concentration of nutrients in
605 the wastes, generated volume of biomass or the investment capacity of the agricultural sector.

606

607 **5. ECONOMIC ASPECTS**

608 The sustainability of a biorefinery approach should consider not only the environmental
609 aspects but also the economic interest for the agricultural sector. The most significant aspects
610 that should be considered for the integrated evaluation of the processes involved in the
611 biorefinery approach are described below.

612 **5.1. Valuable compounds extraction**

613 The extraction section (excluding pretreatment) is most likely to involve the highest costs of
614 both inversion and operation of the biorefinery approach (Serrano et al., 2017a; Serrano et al.,
615 2017b). These costs depend on the selected extraction process: solid-liquid extraction,
616 ultrasound, supercritical fluid extraction, biochemical, etc. The inversion cost of a steam
617 explosion facility treating 100000 t/year of pinewood was reported as 4.2 M€ (Shafiei et al.,

618 2014) and the price of a thermal hydrolysis system with 170 °C steam for treating 30000 t/year
619 of different organic waste was reported as 1.0 M€ (Cano et al., 2014). These inversion costs can
620 be taken as reference for a preliminary economic assessment (Aden et al., 2002).

621 Due to the high price of the extracted compounds, the highest economical incomings also
622 come from this section, warranting a high profitability for the biorefinery. Nowadays the price
623 of potential extracted products reaches values from 0.9-1.2 €/kg for lignin (Purelignin, 2017) to
624 520 €/kg for hydroxytyrosol (solution 10 % wt). However, these prices can fluctuate according
625 to the market changes, affecting the result of an economic assessment. In the case of pectin, the
626 global demand grown from 30000 t/year in 2009 to 60000 in 2015 with a price increment from
627 9.2-10.9 €/kg to 12.6 €/kg in the same period (Ciriminna et al., 2016). In other cases, the
628 apparition of a new process could increase the offer reducing the price. In this sense, in order to
629 study the profitability of the biorefinery, a sensibility analysis of the influence of the extract
630 price must be done.

631 **5.2. Anaerobic Digestion**

632 In this stage, the final use of the methane generated involves different costs, inversion and
633 operational, and revenues possibilities. The alternatives for the use of the methane are listed
634 below:

635 - Direct use (without purification). The gas is conducted to a close consumer (0.6-15 km)
636 for use in boilers or ovens as fuel. The profitability may be compromised if the distance to the
637 consumer is higher than 5 km.

638 - Direct use (after purification). The gas, after being treated for purification, is conducted
639 to the natural gas grill. Due to the high cost of the technology, this alternative can be only
640 applied to great scale projects.

641 - Electricity generation. Capacity, energy efficiency, and costs are a function of the
642 generation technology, i.e. internal combustion generator (350 kW-3MW), turbine (1-6 MW) or
643 micro-turbines (30-200 kW).

644 - Electricity and heat generation. Simultaneous electricity generation with the recovery of
645 useful heat. Higher capital costs than only power generation due to the heat recovery system,
646 but higher global energy efficiency of the generation system. Efficiencies of cogeneration
647 systems can be assumed 33 % (electricity) and 55 % (heat) (Cano et al., 2014).

648 The election process of one of these alternatives must be carried out taking into consideration
649 different aspects as:

650 - Potential methane or heat consumers around the plant. The profitability of a project with
651 the direct use of methane or cogeneration are conditioned to the existence of potential
652 consumers of the generated methane or heat depending on the case.

653 - Local electricity market regulation. The price of the electricity can vary depending on
654 the electricity market regulation of each country or region. The production of “green energy”
655 can be incentivized by increasing the sale price of the electricity generated from biogas respect
656 to that generated from fossil resources (Ruffino et al., 2015). Also, projects including
657 cogeneration systems can receive economic incentives due to the higher energy efficiency
658 (Real_Decreto_436, 2004).

659 - Waste availability. The capacity of waste supply to the biorefinery or the seasonal
660 variation of this capacity are key parameters on the project design. Some power generation
661 systems as gas turbines have a very low flexibility to changes in the gas charge. On another
662 hand, the high cost of a cogeneration system makes this technology only appropriate for high
663 capacities due to the economy of scale.

664 - Internal energy requirement of the biorefinery. An exhaustive study, taking also into
665 consideration the parameters describes before, must be done to determine the optimal energy
666 integration level of the anaerobic digestion system in the biorefinery. It can be noticed that, due
667 to the existence of economic incentives given for the production of energy from removable
668 sources, an energy self-sufficient design doesn't have to be the best option from an economic
669 point of view. A higher integration level involves high inversion costs, a more complex

670 operation and reduces the revenues for methane or energy sale. On another hand, the energy
671 integrations reduce the operational cost derived from the use of electricity and/or fuels.
672 Regarding to the energy requirement of the anaerobic digestion, in a cogeneration system,
673 between the 30 to 50 % of the heat produced can be consumed, to keep the optimal temperature
674 in the digester (Cano et al., 2014; Serrano et al., 2014a), and around 15 % of the electricity
675 (Serrano et al., 2014a). The thermal energy requirement drastically increases when a thermal
676 pretreatment is necessary (Aden et al., 2002; Shafiei et al., 2014). By energy integration, the
677 methane generated by the anaerobic digestion can supply the energy necessary for the
678 production of the steam employed in the pretreatment. In some cases, the methane production
679 excess the pretreatment requirement but, in others, is necessary an additional fuel to cover the
680 steam demand (Cano et al., 2014). On another hand, according to (Franchetti, 2013) up to 80 %
681 of the thermal energy employed in the pretreatment with steam can be recovered as useful heat.

682 **5.3. Composting**

683 The landfill treatment and disposal of the digestate, with a cost in Spain around 25 € per
684 tonne (excluding transport), have been reported as an important fraction of the anaerobic
685 digestion costs (Cano et al., 2014). But “the digestate problem” can be turned into a revenue
686 source by optimal recovery and recycling of nutrients. In UE, additional composting following
687 digestion adds an additional cost up to 30 € per tonne of digested waste and the composted
688 digestate generate prices between 0-50 €/t. The range of price is justified by the different
689 demand depending on the region. I.e. regions with high manure offer, have a lower demand for
690 compost. In order to evaluate the potential use of the compost for agriculture porpoises and to
691 estimate the potential revenue associated with its sale, some critical aspects must be taken into
692 consideration (Saveyn and Eder, 2014):

693 - Fertilisers market. As the compost is used as a substitute of industrial fertilizers, the
694 price of these products can be taken as a reference to estimate the price of the compost. The
695 reference fertilizer must be chosen in base on the compost composition and its nutrients
696 concentration.

697 - Transport costs. The use is usually limited to areas sited in less than 100 km from the
698 biorefinery. For higher distances, the transports cost would be higher than the value per tonne.
699 - Regulation of compost use. Depending on the country or region, the use of compost can
700 be regulated in terms of composition-quality and/or maximum amount of compost that can be
701 applied per year in the same area.

702 As the price of compost is usually lower than the composting costs, with the aim to obtain
703 higher added values products, several alternatives previously cited in this review have been
704 proposed in the literature for nutrients recovery from the digestate. All of them are based in the
705 first step in which solid and liquid phase are separated (Monfet et al., 2017, Gutiérrez et al.,
706 2017, Vaneckhaute et al., 2017). The compost would be only obtained from the solid fraction,
707 reducing the composting cost. But additionally, the solid-liquid separation process has an
708 energy cost of 3.5 kWh/t (Tampio et al., 2016). The costs of the global nutrient recovery process
709 and the obtained revenue will depend on the kind of final product selected.

710 **5.4. Economic viability assessment**

711 Table 6 shows the results of a preliminary economic viability assessment performed for the
712 valorization of two different waste streams, two-phase olive mill solid waste (OMSW) and
713 orange peel waste (OPW), within the bio-refinery concept: steam explosion as thermal
714 pretreatment (SE) + valuable compound extraction (VCE) + anaerobic digestion with
715 cogeneration (AD/CHP) + composting of digestate (CD). Net present value (NPV), internal rate
716 of return (IRR) and payback period (PB) were used to evaluate the economic viability of the two
717 bio-refinery approaches. IRR is calculated as the discount rate for which the Net Present Value
718 (NPV) reaches a value of zero.

719 Technical and economic assumptions employed for the economic viability assessment are:

720 - Plant capacity of 50,000 t/y of waste. The bio-refinery is assumed to be located in Spain and
721 to operate 8000 h along 12 months per year with a 25-year lifetime (IET/1045/2014).

722 – Steam Explosion and valuable compound extraction. OMSW (Serrano et al., 2017b):
723 4,200,000 € (SE) + 21,000,000 € (VCE) Investment; 2% of the investment (SE) + 25,000,000
724 €/y (VCE) operational and maintenance costs; electricity consumption 15 % of electricity
725 production in AD; high-pressure steam requirement 905 MJ/t OMSW (SE); 12.7 kg of 10 wt%
726 phenol extract per kg OMSW. OPW (Forgács et al., 2011): 2,600,000 € (SE+VCE); 2% of the
727 investment operational and maintenance costs (SE+VCE) (general assumption); electricity
728 consumption 1 kWh/t OPW (SE+VCE); high-pressure steam requirement 790 MJ/t OPW
729 (SE+VCE); 7.1 kg/t OPW.

730 – Anaerobic digester and co-generation biogas engine. 39.95 l CH₄/kg OMSW (Serrano et al.,
731 2017b); 82.16 l CH₄/kg OPW (Forgács et al., 2011; Calabrò and Panzera, 2017); 3398 € of
732 construction cost per installed kWe (TrustEE, 2017); electric energy self-supply of the
733 anaerobic digester 15% of the electricity generated by the co-generation biogas engine
734 (Angelidaki et al., 2006); operational and maintenance costs 2% of the construction cost;
735 complementary costs ought to be considered regarding feasibility studies and administrative and
736 authorization requirements (30,000 €) (González-González and Cuadros, 2013). The methane
737 production was calculated based on lab scale experiment by applying a scale up factor of 85%.
738 In the case of OPW, the experimental methane yield obtained in (Forgács et al., 2011) was
739 adjusted for a campaign of 8 month using fresh waste and 4 months using ensiled waste with a
740 30 % reduction of volatile solids Calabrò and Panzera, 2017)

741 – Energy Integration. The efficiency in the energy obtained through a cogeneration biogas
742 engine is considered 33% for electricity and 55% for thermal energy (30% in hot water and 25%
743 in exhausted gas). Hot water is employed to keep the operation temperature of the AD reactor
744 (Cano et al., 2014). The thermal energy contained in the exhausted gas from the biogas engine
745 and an additional stream of natural are used to obtain the necessary high-pressure steam supply
746 for the steam explosion treatment. Any excess electricity is fed into the grid. Composting
747 penalizes the electricity sale by 3,5 kWh per ton of digestate (Tampio et al., 2016).

748 – Prices. 42.67 €/MWh of electricity (creara.es, 2018), 0.04 €/kWh of consumed Natural gas
749 (endesaonline.es, 2018), 520 €/kg of 10% phenol extract (Ciriminna et al., 2016), 1 €/kg of

750 Limonene (Negro et al., 2017b). No profit has been considered from the sale of compost. The
751 composting cost is assumed to be equal to the incoming from the sale of it. Under the
752 economical point of view, composting would be justified by the aim to avoid the cost of bio-
753 waste dumping.

754 – Economic outlook. The value of the discount rate is assumed to be equal to the Spanish ten-
755 year bond yield. The industrial profit taxes are fixed at 25%. Additionally, a minimum
756 admissible price for the extracted valuable product has been calculated by imposing a NPV
757 equal to zero.

758 Both wastes show promising results with positive NPV and IRR higher than 7.5%, which is
759 an adequate IRR value according to the Spanish regulation for small scale AD plants (Biogas,
760 2014). In the case of OPW, a payback period of more than 5 years would be an inconvenience in
761 order to face the inversion. A reduction of the price of the valuable extracted over 91% (phenol
762 extract) and 70% (D-limonene) would turn this bio-refinery pathway as unprofitable for OMSW
763 and OPW respectively.

764 The profitability of the inversion would be higher if the electricity excess is used to supply the
765 electricity requirement in the olive oil production plant or orange juice plant respectively. In this
766 case, the excess of electricity, of a portion of it, could be assumed as a save cost, with a price of
767 123,4 €/MWh (EUROSTAT, 2018), higher than the sale price of 42.67 €/MWh.

768

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1307 **Table 1.** Bioactive compounds in fruit and vegetables by-products.

By-product	Phenolic	Flavonoids	Anthocyanin	Carotenoids	Phytosterols	References
Citrus	Cumaric, caffeic, ferulic acids	Hesperidin, hesperetin Naringin, naringenin, Luteolin, tangeretin	cyanidin 3-glucoside and cyanidin 3-(6''-malonyl-glucoside)		In oil mixture peel/seed	(Fabroni et al., 2016; M'hiri et al., 2017; Ndayishimiye et al., 2017)
Tomato		Quercetin, rutin, kaempferol, naringenin		β carotene, lutein, lycopene		(Fattore et al., 2016; Gharbi et al., 2017)
Onion		Quercetin and kaempferol glycosides	In red onion	α - and β carotene		(Sharma et al., 2015; Wiltshire et al., 2017)
Carrot	chlorogenic acids, dicaffeoylquinic acids			α - and β - carotene		(Sánchez-Rangel et al., 2016)
Potato	chlorogenic , caffeic, protocatechuic and ferulic acid	Catechin, quercetin and kaempferol rutinoside	anthocyanidins			(Akyol et al., 2016)
Grape	caffeic, gallic, protocatechuic, 4- hydroxybenzoic and syringic acid, hydroxytyrosol,	(+)-catechin, catechin dimer and trimer, (-)- epicatechin, procyanidin B1 and B2, quercetin-3-O- rhamnoside, luteolin, resveratrol (stilbene)	Proanthocyanidins and anthocyanins		tocopherol tocotrienols	(Barba et al., 2016; Teixeira et al., 2014)
Mango	gallic acid, salicylic acid	Epicatechin, quercetin	hydrolysable tannins, anthocyanins			(da Silva and Jorge, 2017; Dorta and Sogi, 2016)
Papaya	Gallic, chlorogenic, caffeic , protocatechuic acid	quercetin, kaempferol		lutein, lycopene cryptoxanthin β -carotene	α -tocopherol	(de Moraes Crizel et al., 2016)
Banana	Syringic, tannic, gallic, cinnamic, p- coumaric, ferulic	Gallocatechin gallate. quercetin	proanthocyanidins	α - and β - carotene, xanthophyll	B-sitosterol, campesterol, stigmasterol	(Singh et al., 2016; Vu et al., 2016)

	acid.					
Pineapple	Gallic acid, ferulic acid	catechin, epicatechin	anthocyanins	α and β -carotene		(Da Silva et al., 2014; Li et al., 2014)
Olive oil	Hydroxytyrosol, 3,4-dihydroxyphenylglycol, secoiridoides.	Luteonin, apigenin.				(Fernández-Bolaños et al., 2002; Rubio-Senent et al., 2013b)

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Table 2. Health benefits from compounds of fruits and vegetables by-products.

By-product	Compound	Uses and health benefits	Reference
Citrus	Pectin	Thickening agent, gelling, stabilizer	(John et al., 2017)
	Dietary fiber	Regulation of intestinal transits, prevention of diabetes, hypertension, cardiovascular disease and colon cancer.	
Tomato	Pectin	Thickening agent, gelling, stabilizer	(Morales-Contreras et al., 2017)
Onion	Fructooligosaccharides	Prebiotic effect	(Smith et al., 2015)
Carrot	Pectin	Thickening agent, gelling, stabilizer	(Jabbar et al., 2015; Jafari et al., 2017)
	Antioxidant dietary fiber	Benefits from fiber together with antioxidant activity.	
Potato	Water-soluble polysaccharides	Additive in food, pharmaceutical and cosmetic preparation.	(Jeddou et al., 2016)
Grape	Antioxidant dietary fiber	Benefits from fiber together with antioxidant activity	(Zhang et al., 2017)
Mango	Antioxidant dietary fiber	Benefits from fiber together with antioxidant activity	(Matharu et al., 2016)
	Pectin (highly-esterified)	Used in marmalade, dairy and meat preparations	
	Soluble dietary fiber	Control diabetes	
Pineapple	Dietary fiber	Fortifying extruded products. Prebiotic effect.	(Selani et al., 2014)
Olive oil	Oligosaccharides, pectins, phenolic glucosides	Prebiotic agents, antioxidant, antiproliferatives	(Rubio-Senent et al., 2015a; Rubio-Senent et al., 2013a)

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Table 3. Methane yield and biodegradability of different lignocellulosic by-products

Waste	Pretreatment	Methane yield (mL CH ₄ /g VS)	Biodegradability (in VS)	Reference
Switchgrass (winter and fresh summer harvested)	Pretreatment: grinding + alkalisation + autoclaving.	140 (winter harvested) 298 (fresh summer harvested)	No available	(Frigon et al., 2012)
Grass (<i>Pennisetum hybrid</i>) Hay	Grass pretreatment: 30 min with water vapour. Hay pretreatment: Steam-explosion at 175 °C for 10 min.	190 (raw grass) 236 (raw hay) 198 (pretreated grass) 281 (pretreated Hay)	No available	(Rodriguez et al., 2017)
Dried grass silage	Pretreatments at 100°C and NaOH loading rates of 1%, 2.5%, 5% and 7.5% by VS mass in grass silage	360 (NaOH 1%) 402 (NaOH 2.5%) 450 (NaOH 5%) 453 (NaOH 7.5%)	76.9% (NaOH 1.0%) 85.3% (NaOH 2.5%) 95.2% (NaOH 5.0%) 96.7% (NaOH 7.5%)	(Xie et al., 2011)
Salix biomass	Steam-explosion pretreatment at 170-230 °C, during 5-15 min.	161 (raw Salix) 234 (pretreated Salix, 230 °C, 10 min)	No available	Estevez (Estevez et al., 2012)
Sunflower stalks	Two thermal (55 and 170 °C) and five thermo-chemical pretreatments (NaOH, H ₂ O ₂ , Ca(OH) ₂ , HCl and FeCl ₃)	192 (raw sunflower stalks) 259 (pretreated sunflower stalks, 55 °C with 4% NaOH for 24 h)	No available	(Monlau et al., 2012)
Wheat Straw	Steam-explosion pretreatment: 160-200 °C, 10-20 min.	276 (raw wheat straw) 331 (pretreated wheat straw, 180 °C, 15 min)	Maximum VS removal efficiency: 46.3% (200 °C, 10 min).	(Bauer et al., 2009)
Wheat Straw	Thermal pretreatment at 80 °C, 10 min + H ₂ O ₂ (7.5%).	188 (raw wheat straw) 274 (pretreated wheat straw)	41.1% (raw wheat straw)	(Hassan et al., 2016)
Barley Straw (BS) Wheat Straw (WS) Rice Straw (RS)	Thermal pretreatment in autoclave at 90 and 120 °C during 30 min.	240 (raw BS) 197 (raw WS) 182 (raw RS) 338 (pretreated BS, 120°C) 299 (pretreated WS, 120°C)	No available	(Menardo et al., 2012)

		261 (pretreated RS, 120°C)		
Peanut hull (<i>Arachis hypogaea</i>)	Mechanical (grinding into < 20 mm) + thermal (80 °C, 70 min) + 3 g NaOH/100 g TS (24 h, at 55 °C).	112 (raw peanut hull) 182 (pretreated peanut hull)	26% (raw peanut hull) 47% (pretreated peanut hull)	(Dahunsi et al., 2017)
Ensiled sorghum forage (ESF) Wheat straw (WS)	Thermal and thermo-alkaline pretreatments at 100 °C, and 160 °C for 30 min, without and with the addition of NaOH solutions at 1% and 10% g NaOH/g TS.	269 (raw ESF) 204 (raw WS). 361 (pretreated ESF, 10% NaOH, 100 °C) 341 (pretreated WS, 10% NaOH, 100 °C)	84-85% (pretreated ESF and WS, 10% NaOH, 100 °C)	(Sambusiti et al., 2013)

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1316 **Table 4.** Methane yield and biodegradability of different vegetable by-products

Waste	Pretreatment	Methane yield (mL CH ₄ /g VS)	Biodegradability (in VS)	Reference
Onion skin Cauliflower stems Potato skin Carrot petioles Beet leaves	No pretreatments	400 (Onion skin) 331 (Cauliflower stems) 267 (Potato skin) 309 (Carrot petioles) 231 (Beet leaves)	No available	(Ji et al., 2017)
Onion waste	No pretreatments	340	50–70%	(Lubberding et al., 1988)
Onion waste	Mechanical pretreatment (pieced to 1-2 cm size)	378	61%	(Menardo and Balsari, 2012)
Carrot waste pomace	No pretreatments.	198	56% COD removal	(Garcia et al., 2011)
Potato waste	No pretreatments.	320 mL CH ₄ /g VS	No available	(Parawira et al., 2004)
Mixture Tomato waste: sewage sludge (5:95)	No pretreatment	159	95%	(Belhadj et al., 2014)
Tomato waste Cucumber waste	No pretreatment	299 (Tomato waste) 177 (Cucumber waste)	No available	(Gil et al., 2015)
Tomato waste Cucumber waste	No pretreatment	320 (Tomato waste) 260 (Cucumber waste)	79% (Tomato waste) 43% (Cucumber waste)	(Jagadabhi et al., 2011)

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1320 **Table 5.** Methane yield and biodegradability of different fruit by-products

Waste	Operational conditions	Methane yield (mL CH ₄ /g VS)	Biodegradability (in VS)	Reference
Banana peel Pineapple waste	No pretreatment	188-240 (Banana peel) 413 (Pineapple waste)	36% (Banana peel) 58% (Pineapple waste)	(Bardiya et al., 1996)
Peels of Fluted pumpkin (<i>Telfaria occidentalis</i>)	Combination of mechanical (grinding into < 20 mm), thermal and chemical (NaOH) (3 g NaOH/100 g TS for 24 h at 55 °C)	164 (mechanical, thermal and chemical pretreatment) 161 (mechanical and chemical pretreatment)	67.2% COD reduction	(Dahunsi et al., 2016)
Pineapple pulp and peel	No pretreatment	111-430	56.0-73.8%	(Namsree et al., 2012)
Mango Peel waste and Pomegranate seed de-oiled cake (PSDC)	No pretreatment	300-410 (Only mango peel waste) 610 (Mango peel waste and PSDC)	23.4% (Only mango peel waste) 25.8% (Mango peel waste and PSDC)	(Suryawanshi et al., 2013)
Banana Skin Tomato waste	No pretreatment	277 (Banana skin) 298 (Tomato waste)	74% (Banana skin) 77% (Tomato waste)	(Ji et al., 2017)
Orange peel	Steam distillation (6 h) for 70% Limonene removal	230-290	84-90% COD removal	(Martín et al., 2010)
Orange peel	Steam distillation (6 h) for 70% Limonene removal	332 (thermophilic conditions) 230 (mesophilic conditions)	84-90% COD removal	(Siles et al., 2016)
Kiwi waste	No pretreatment	371	58% in VS	(Menardo and Balsari, 2012)
Strawberry extrudate	Thermal pretreatment (120 °C, 2 atm, 5-60 min).	285 (raw strawberry extrudate) 339 (pretreated strawberry extrudate)	No available	(Serrano et al., 2015)
Strawberry extrudate	Mechanical sieving (1 mm mesh size).	230 (raw strawberry extrudate) 312 (pretreated strawberry extrudate)	90% in VS	(Siles et al., 2013)
Winery waste	No pretreatment	250	79% in COD	(Siles et al., 2011)
Two-phase OMSW	Thermal pretreatment at 100, 120, 160 and 180 °C during 60, 120 and 180 min for each temperature studied.	373 (raw OMSW) 392 (pretreated OMSW, at 120°C, 180 min)	No available	(Rincón et al., 2013)

		380 (pretreated OMSW, at 180°C, 180 min)		
Two phase OMSW	High temperature thermal pretreatment (170 °C for 60 min) and phenol extraction	261 (raw OMSW) 290 (OMSW, thermal pretreatment) 350 (OMSW, thermal pretreatment and phenol extraction)	57.3% (raw OMSW) 63.4% (OMSW, thermal pretreatment) 75.3 (OMSW, thermal pretreatment and phenol extraction)	(Serrano et al., 2017a)
Two phase OMSW	Steam-explosion pretreatment: 200 °C, 5 min (pressure: 42 kg/cm ²) and phenol extraction	280 (raw OMSW) 294 (OMSW, steam-explosion) 261 (OMSW, steam-explosion and phenol extraction)	66.4% (raw OMSW) 65.0% (OMSW, thermal pretreatment) 60.5 (OMSW, thermal pretreatment and phenol extraction)	(Serrano et al., 2017b)

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1323 **Table 6.** Economic viability assessment of OMSW and OPW AD-biorefineries

Electricity to the grid (GWh/year)	Extracta production (t/year)	Initial investment cost (€)	Incoming (€/year)	Annual costs (€/year)	NPV (M€)	IRR (%)	Payback period (years)	Limit extract^a price (€/kg)^b
4.0	633	28,013,747	329,354,041	25,444,066	4.7·103	815	<1	42.0
11.3	353	8,355,042	834,871	196,947	3.86	5	17	0.3

1324 a Extract: 10 %w phenol extract for OMSW and D-limonene for OPW

1325 b NPV = 0;

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